
Safety Assessment of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate as Used in Cosmetics

Status: Draft Amended Report for Panel Review
Release Date: November 10, 2025
Panel Meeting Date: December 4-5, 2025

The Expert Panel for Cosmetic Ingredient Safety members are: Chair, Wilma F. Bergfeld, M.D., F.A.C.P.; Donald V. Belsito, M.D.; David E. Cohen, M.D.; Samuel M. Cohen, M.D., Ph.D.; Curtis D. Klaassen, Ph.D.; Allan E. Rettie, Ph.D.; David Ross, Ph.D.; Paul W. Snyder, D.V.M., Ph.D.; and Susan C. Tilton, Ph.D. The Cosmetic Ingredient Review (CIR) Executive Director is Bart Heldreth, Ph.D., and the Senior Director is Monice Fiume, M.B.A. This safety assessment was prepared by Christina Burnett, M.S., Senior Scientific Analyst/Writer, CIR, and Jinqiu Zhu, Ph.D., Toxicologist, CIR.

| Public Comment | CIR | Expert Panel | Re-Review | Rpt Status |
|--|---|---|--|--|
| | <p>announce</p> <p>JACT 4(3):267-303, 1985; IJT 24 (Suppl 1):37-40, 2005 (RR); IJT 36 (Suppl 2):34-42, 2017 (RR)</p> | <p>>15 years since last review</p> <p>PRIORITY LIST</p> <p>Are new data cause to reopen?</p> <p>YES</p> <p>DRAFT AMENDED REPORT December 2025</p> <p>Table IDA TAR</p> <p>IDA</p> <p>DRAFT TENTATIVE AMENDED REPORT</p> <p>Table</p> <p>Issue TAR</p> <p>DRAFT FINAL AMENDED REPORT</p> <p>Table Different Conclusion</p> <p>Issue FAR</p> | <p>New Data; or request for cause (FDA)</p> <p>Re-Review Panel</p> <p>Are new ingredients appropriate for inclusion/re-open?</p> <p>Yes</p> <p>RE-REVIEW SUMMARY</p> | <p>Dibutyl Phthalate was placed on the 2 Priorities List following nomination by FDA for cause</p> |
| <p>60-day public comment period</p> <p>PUBLISH</p> | <p>DAR</p> <p>Table</p> <p>IDA Notice</p> <p>Draft TAR</p> <p>Table</p> <p>Tentative Amended Report</p> <p>Draft FAR</p> <p>Table</p> <p>Final Amended Report</p> | | | |



Commitment & Credibility since 1976

Memorandum

To: Expert Panel for Cosmetic Ingredient Safety Members and Liaisons
 From: Christina L. Burnett, M.S., Senior Scientific Analyst/Writer, CIR
 Jinqiu Zhu, Ph.D., Toxicologist, CIR
 Date: November 10, 2025
 Subject: Amended Safety Assessment of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate as Used in Cosmetics

Enclosed is the Draft Amended Report on the Safety of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate as Used in Cosmetics. (It is identified as *report_Phthalates_122025* in the pdf document.) Dibutyl Phthalate was placed on the 2024 Priorities List following nomination by the FDA for cause. The Panel first published the Final Report of the Safety Assessment of Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate in 1985, and concluded that these ingredients are safe for topical application in the present practices of use and concentration in cosmetics (*originalreport1985_Phthalates_122025*). Upon re-review in 2002, the Panel reaffirmed the original conclusion, as published in 2005 (*rereview2005_Phthalates_122025*). In December 2012, the Panel deliberated on studies separately concerning endocrine disruption and diabetes and Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate, and Butyl Benzyl Phthalate; however, the Panel chose not to re-open the safety assessment of these ingredients and published their discussion as a re-review summary in 2017 (*rereview2017_Phthalates_122025*). The data from the original report and from the re-review evaluated by the Panel prior to publishing the 2005 re-review summary (data document included herein as *RRdata_Phthalates_122025*), have been summarized in the Draft Amended Report, in *italicized text*.

According to RLD that CIR received in 2024, Diethyl Phthalate is used in 168 formulations, with most of the uses reported in fragrance preparations. Additionally, the RLD reported Dibutyl Phthalate is used in 2 manicuring preparations and Dimethyl Phthalate has no uses. VCRP survey data received in 2023 reported Diethyl Phthalate was used in 1 skin care formulation; no uses were reported for Dibutyl Phthalate or Dimethyl Phthalate. When comparing the VCRP data received in 2023 to that received in 2001, the frequencies of use for these phthalate ingredients have greatly decreased since the 2005 re-review was published; in 2001, Dibutyl Phthalate was reported to have 150 uses (most in manicuring preparations), Diethyl Phthalate was reported to have 73 uses (most in fragrance preparations), and Dimethyl Phthalate was reported to have 12 uses (most in non-coloring hair preparations).

The results of the concentration of use survey conducted by the Council in 2025 indicate Diethyl Phthalate has a maximum concentration of use range of 0.1 - 0.15%, with 0.15% reported in leave-on face and neck products (*data_Phthalates_122025*). No concentrations of use were reported for Dibutyl Phthalate or Dimethyl Phthalate; however, responses to the survey indicated that Dibutyl Phthalate and Diethyl Phthalate may be present in cosmetics as impurities. In the 2005 re-review, the maximum concentration of use range for Dibutyl Phthalate was 0.0038 - 15% (with 15% reported in manicuring preparations). Diethyl Phthalate was reported to have a maximum concentration of use range of 0.00003 - 11% (11% was reported in perfumes), and Dimethyl Phthalate was reported to have a maximum concentration of use range of 0.00002 - 2% (2% was reported in hair spray).

Additional supporting documents for this report package include a flow chart (*flow_Phthalates_122025*), report history (*history_Phthalates_122025*), a search strategy (*search_Phthalates_122025*), a data profile (*datapofile_Phthalates_122025*), the minutes from all the meetings at which these ingredients were discussed during the original review (*originalminutes_Phthalates_122025*), and the transcripts from recent meeting discussing strategy (*transcripts_Phthalates_122025*).

If no further data are needed, the Panel should formulate an updated Discussion and issue a Tentative Amended Report. However, if additional data are required, the Panel should be prepared to identify those needs and issue an Insufficient Data Announcement.

Phthalates History

1985– The CIR Final Report on the Safety Assessment of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate was published in the *Journal of the American College of Toxicology*. The Panel concluded that these ingredients are “safe for topical application in the present practices of use and concentration in cosmetic.”

2005 – A re-review summary was published reaffirming the 1985 report’s conclusion.

2017 – A re-review summary was published reaffirming the 1985 report’s conclusion. Prior to the publication, in 2012, the Panel had deliberated studies concerning endocrine disruption and diabetes for Dibutyl Phthalate, Diethyl, Dimethyl, Butyl Benzyl Phthalate. However, the Panel chose not to reopen and published their discussion as a re-review summary.

June 2023 – CIR received communication from members of the FDA nominating Dibutyl Phthalate to the 2024 Priority List, for-cause. The Panel agreed to this request.

March 2024 - The Panel discussed the strategy of the preparation of the Draft Amended Report for Dibutyl Phthalate. The Panel determined that Diethyl Phthalate and Dimethyl Phthalate should be included in the safety assessment as much of the published literature involves data on these ingredients together; however, the Panel stressed that any use, exposure, or risk data from one ingredient could not be read-across to the other ingredients in the report, citing in part very different use conditions from one ingredient to the next. The Panel also offered guidance on the presentation of the data, especially data relating to endocrine effect, in the updated safety assessment. Additionally, the Panel discussed the current regulatory status of Dibutyl Phthalate in the European Union and in several US states. The Panel requested clarification on the status of these regulations, and the data that support such.

June 2025 – In a strategy memo, the Panel was asked:

- Does the Panel, or any other stakeholder, have a particular expert in these areas they would like to invite to give a presentation on these DART and endocrine studies?
 - The Panel responded in the affirmative and provided potential experts to invite.
- Does the Panel support the idea of having Dimethyl Phthalate in a separate re-review proposal document or would the Panel prefer that this ingredient stay in the safety assessment with Dibutyl Phthalate and Diethyl Phthalate?
 - The Panel determined to keep these ingredients together in one report.

Phthalates Data Profile* - December 2025 - Christina Burnett

| | | | | Toxicokinetics | | | Acute Tox | | | Repeated Dose Tox | | | DART | | Genotox | | Carci | | Dermal Irritation | | | Dermal Sensitization | | | | Ocular Irritation | | Clinical Studies | |
|--------------------|--------------|---------------|------------|---------------------------|--------------------|--------|-----------|--------|------------|-------------------|--------|------------|--------|--------|----------|---------|--------|--------|-------------------|--------|-------|----------------------|--------|-------|---------------|-------------------|--------|---------------------------|--------------|
| | Reported Use | Method of Mfg | Impurities | log P/log K _{ow} | Dermal Penetration | ADME | Dermal | Oral | Inhalation | Dermal | Oral | Inhalation | Dermal | Oral | In Vitro | In Vivo | Dermal | Oral | In Vitro | Animal | Human | In Vitro | Animal | Human | Phototoxicity | In Vitro | Animal | Retrospective/Multicenter | Case Reports |
| Dibutyl Phthalate | X O | O | O | X | X O | X O | | X O | X | O | X O | O | | X O | X O | | X | X O | | X O | O | | X O | O | O | X | X O | X O | X O |
| Diethyl Phthalate | X O | O | O | X | X O | X O | O | X O | X | O | X O | | X O | X O | X O | | O | | | X O | O | | X O | O | O | | O | | |
| Dimethyl Phthalate | O | O | O | X | X O | O | O | X O | | O | X O | | O | X O | X O | | O | | | O | O | | X O | O | | | O | | |

* "X" indicates that new data were available in a category for the ingredient. "O" indicates data were reported in the original safety assessment.

Phthalates

| Ingredient | CAS # | InfoB | SciFind | PubMed | TOXNET | FDA | EU | ECHA | IUCLID | SIDS | ECETOC | HPVIS | NICNAS | NTIS | NTP | WHO | FAO | NIOSH | FEMA | Web |
|--------------------|----------|-------|---------|--------|--------|-----|----|------|--------|------|--------|-------|--------|------|-----|-----|-----|-------|------|-----|
| Dibutyl Phthalate | 84-74-2 | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ |
| Diethyl Phthalate | 84-66-2 | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ |
| Dimethyl Phthalate | 131-11-3 | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ | √ |

Search Strategy***PubMed***

From 1999 to present:

(Dibutyl Phthalate) OR (84-74-2[EC/RN Number]) = 2849 hits

(Diethyl Phthalate) OR (84-66-2[EC/RN Number]) = 1239 hits

(Dimethyl Phthalate) OR (131-11-3[EC/RN Number]) = 822 hits

ECHA

Entry for CAS # 84-74-2, 84-66-2, and 131-11-3 resulted in finding dossiers for Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate, respectively. Pertinent data not found in the original report is summarized in the amended report.

Typical Search Terms

- INCI names
- CAS numbers
- chemical/technical names
- additional terms will be used as appropriate

LINKS

Search Engines

- Pubmed - <http://www.ncbi.nlm.nih.gov/pubmed>
 - appropriate qualifiers are used as necessary
 - search results are reviewed to identify relevant documents
- Connected Papers - <https://www.connectedpapers.com/>
- DeepDyve - <https://www.deepdyve.com/>

Pertinent Websites

- wINCI - <https://incipedia.personalcarecouncil.org/winci/ingredient-custom-search/>
- FDA Cosmetics page - <https://www.fda.gov/cosmetics>
- eCFR (Code of Federal Regulations) - <https://www.ecfr.gov/>
- FDA search databases: <https://www.fda.gov/industry/fda-basics-industry/search-databases>
- Substances Added to Food (formerly, EAFUS): <https://www.fda.gov/food/food-additives-petitions/substances-added-food-formerly-eafus>
- GRAS listing: <https://www.fda.gov/food/food-ingredients-packaging/generally-recognized-safe-gras>
- SCOGS database: <https://www.fda.gov/food/generally-recognized-safe-gras/gras-substances-scogs-database>
- Inventory of Food Contact Substances Listed in 21 CFR: <https://www.cfsanappsexternal.fda.gov/scripts/fdcc/index.cfm?set=IndirectAdditives>
- Drug Approvals and Database: <https://www.fda.gov/drugs/development-approval-process-drugs/drug-approvals-and-databases>
- FDA Orange Book: <https://www.fda.gov/drugs/drug-approvals-and-databases/approved-drug-products-therapeutic-equivalence-evaluations-orange-book>
- OTC Monographs - <https://dps.fda.gov/omuf>
- Inactive Ingredients Approved For Drugs: <https://www.accessdata.fda.gov/scripts/cder/iig/>
- FEMA (Flavor & Extract Manufacturers Association) GRAS: <https://www.femaflavor.org/fema-gras>
- HPVIS (EPA High-Production Volume Info Systems) - https://iaspub.epa.gov/opthpv/public_search.html_page
- NIOSH (National Institute for Occupational Safety and Health) - <http://www.cdc.gov/niosh/>
- NTIS (National Technical Information Service) - <http://www.ntis.gov/>
 - technical reports search page: <https://ntrl.ntis.gov/NTRL/>
- NTP (National Toxicology Program) - <http://ntp.niehs.nih.gov/>
- EUR-Lex - <https://eur-lex.europa.eu/homepage.html>
- Scientific Committees (SCCS, etc) opinions: https://health.ec.europa.eu/scientific-committees_en https://health.ec.europa.eu/scientific-committees/scientific-committee-consumer-safety-sccs_en
- ECHA (European Chemicals Agency – REACH dossiers) – <https://echa.europa.eu/>
- European Medicines Agency (EMA) - <http://www.ema.europa.eu/ema/>
- OECD SIDS (Organisation for Economic Co-operation and Development Screening Info Data Sets)- <http://webnet.oecd.org/hpv/ui/Search.aspx>
- EFSA (European Food Safety Authority) - <https://www.efsa.europa.eu/en>
- ECETOC (European Centre for Ecotoxicology and Toxicology of Chemicals) - <http://www.ecetoc.org>
- AICIS (Australian Industrial Chemicals Introduction Scheme)- <https://www.industrialchemicals.gov.au/>
- International Programme on Chemical Safety <http://www.inchem.org/>
- Office of Dietary Supplements <https://ods.od.nih.gov/>
- FAO (Food and Agriculture Organization of the United Nations) - <http://www.fao.org/food/food-safety-quality/scientific-advice/jecfa/jecfa-additives/en/>
- WHO (World Health Organization) IRIS library - <https://apps.who.int/iris/>
- a general Google and Google Scholar search should be performed for additional background information, to identify references that are available, and for other general information - www.google.com <https://scholar.google.com/>

JUNE 2023 MEETING – REQUEST TO REOPEN FOR CAUSE

Belsito Team – June 12, 2023

DR. BELSITO: Priorities. So basically, we've just been asked to prioritize -- that's in admin, right?

DR. SNYDER: Yeah.

DR. KLAASSEN: Yes.

DR. BELSITO: So, since our March meeting we received communication from the FDA nominating ingredients for cause, specifically Toluene and Dibutyl Phthalate. So, we're going to be doing accelerated re-reviews on those. And then there was something here that I just want Monice or someone to clarify. So, it basically said that instead of just doing a re-review summary, we're going to fully open this or something?

MS. FIUME: So, are you talking about Toluene?

DR. BELSITO: Yeah.

DR. SNYDER: We never reviewed it before.

MS. FIUME: Well, it is on our list of items to be re-reviewed. It's currently on Christina's docket. Right, you have Toluene?

DR. BELSITO: Right. We reviewed both of them before.

MS. BURNETT: I think so. I don't know.

DR. SNYDER: Oh, that's the TPO. I was talking about TPO. Yeah.

DR. BELSITO: Right.

DR. SNYDER: I'm sorry, TPO is what I was talking about.

MS. FIUME: Right. TPO is the only one. Dibutyl Phthalate was just re-reviewed in 2017.

DR. BELSITO: Right.

MS. FIUME: But Toluene was scheduled for consideration for re-review this year, so you will be seeing that soon.

DR. BELSITO: Right. But it says, "The CIR will present the panel with a draft amended report on this ingredient instead of an abbreviated re-review document."

MS. FIUME: Okay. So instead of getting the table that you have been --

DR. BELSITO: Right. We are actually going to get a written document?

MS. FIUME: Assuming that you were going to accept FDA's request to reopen it.

DR. BELSITO: I think if FDA comes to us with a request for cause, we have to -- I don't know -- yeah.

MS. FIUME: Which is why you'll get an actual report person versus do you want to reopen? Here's the table of data that we found and then -- just taking that step out.

DR. BELSITO: Right, okay. So, we're going to -- yes, we're reopening Dibutyl Phthalate and Toluene for cause. And I think the third ingredient -- I mean, this is the type of stuff that I want to see happening. Something's going on in Europe, there's a concern about this material for reproductive toxicity, we need to be looking at it, number one. Number two, we've never even reviewed it. So, yes, I personally would like it added to the 2024 priority list.

DR. SNYDER: Agreed.

DR. KLAASSEN: It would be interesting to know why they wanted these first two chemicals. We don't -- why they want us to do Dibutyl Phthalate?

DR. BELSITO: Because it's a huge issue in endocrine disruption --

DR. KLAASSEN: Right, right.

DR. BELSITO: -- and --

DR. KLAASSEN: But I don't think there's any new data since the last time we did it, but maybe there is. And how about Toluene? I mean, I'm not against doing it, I'm just wondering. It'd be nice if they said why.

MS. FIUME: So, I'm looking at the memo and the email that was originally sent on March 20th, it's PDF Page 26. It just says that they're proposing it.

DR. BELSITO: Yeah. This is from Prashiela.

DR. KLAASSEN: Yeah. It says nothing really.

DR. BELSITO: Right.

MS. FIUME: Sorry, Priya has Toluene. So, Priya will be bringing that back probably in September.

MS. BURNETT: And Phthalates.

MS. FIUME: Yeah.

DR. BELSITO: I mean, both of them have gotten a lot of press, you know, bad press.

DR. KLAASSEN: Yeah, I know about the phthalates always do.

DR. BELSITO: Well, Toluene for carcinogenicity.

DR. RETTIE: So, the phthalates are the less (inaudible) issues, right?

DR. BELSITO: Right. I'm surprised that they are supposedly only one reported use because they used to be used in a lot of nail enamels. But I guess now everyone's using acrylic, so I don't know.

DR. KLAASSEN: Well, let's do them.

DR. SNYDER: Been there, done that.

DR. BELSITO: They're also used in a lot of fragranced products to hold the fragrance on the skin as a fixative, I think.

MS. KOWCZ: No.

DR. BELSITO: No?

MS. EISENMANN: Diethyl.

MS. KOWCZ: The Diethyl.

DR. BELSITO: Yeah, diethyl. Okay.

Cohen Team – June 12, 2023

DR. COHEN: All right. Now we're going to Priorities. Okay, for the 2024, draft priorities, we asked for propolis to be accelerated. Two other ingredients were initially proposed and were removed from the list and it was determined that cannabidiol should be reviewed singly.

The others are listed here, some with pretty high frequencies of use reported. Any comments on this? I mean, I don't know if we're going to have a really in depth conversation about this, are we?

DR. HELDRETH: I think that the main point was that FDA had actually asked for three additions to our prioritization. Two of these are request for accelerated rereviews, so Toluene and the Dibutyl Phthalate.

Now Toluene was actually already in our in-house pipeline. We were already working on it, so that one's definitely coming back your way. Dibutyl Phthalate, we haven't started working on yet. But now that FDA has requested it, we've went ahead and added it, unless the Panel has an objection to accelerating that be reviewed.

So, the only real question, I think, for the panel is do they want to add this Trimethylbenzoyl Diphenylphosphine Oxide to the prioritization list for next year?

DR. SLAGA: I think we should accelerate it.

DR. COHEN: Yeah. That's a question to the Panel. We should add them.

DR. TILTON: Yeah, I agree.

DR. ROSS: New data. I agree.

DR. HELDRETH: Okay. That's easy.

DR. ROSS: Bart, could I ask you, what was the reason for -- or maybe you don't know -- why FDA nominated Toluene and the Dibutyl Phthalate? Was there a specific reason?

DR. HELDRETH: Prashiela stepped out?

DR. ANSELL: Our FDA person just --

DR. COHEN: We can ask her when she comes back.

DR. ROSS: Ah, okay.

DR. COHEN: These are plastics, the phthalates, right?

DR. HELDRETH: Plasticizer, yeah.

DR. ROSS: Yeah, they're phthalates. Toluene is a little different.

DR. COHEN: Yeah, Toluene is going to be a bit different.

DR. HELDRETH: Well, we've looked at the phthalates before.

DR. ROSS: Yeah.

DR. BERGFELD: And there's a lot of endocrine disruption with that group.

DR. COHEN: So, it's interesting. In 2017, the panel reaffirmed it, so this would be a real short cycle.

DR. HELDRETH: Right.

DR. COHEN: Prashiela, a question. No, no, no, it's okay. For the priority list, the FDA nominated some items, one was Toluene. Do you know why Toluene was nominated?

DR. MANGA: I'm going to have to get back to you on that one. Let me take a quick look at what we --

DR. COHEN: And the phthalates, the dibutyl phthalate?

DR. MANGA: I think there's just a lot of interest in phthalates right now. It's come up quite a bit. The Toluene is being used in a lot of nail products.

DR. ANSELL: Historically.

DR. MANGA: Historically.

DR. COHEN: Are you talking about the Toluene sulfonamide resins or just Toluene?

DR. ANSELL: No Toluene is a diluent.

DR. ROSS: I think Toluene is being reviewed quite a bit at IARC on its own, but also in connection with Benzene.

DR. ANSELL: Right. Also not used anymore, so.

DR. ROSS: Yes.

DR. ANSELL: But we fully support accelerating anything FDA ask us to.

DR. COHEN: We're good. Yeah. So are we.

DR. HELDRETH: Which is a question, I just wondered why they --

DR. BERGFELD: Actually, we really like it when they ask.

DR. ANSELL: Yes. More than support it, encourage it.

DR. MANGA: We appreciate that.

DR. COHEN: No, it's nice we're being paid attention to. And the other one was -- Annex 3 was a little more self-explanatory.

DR. BERGFELD: What was that?

DR. COHEN: The Trimethylbenzoyl Diphenylphosphine Oxide.

DR. ROSS: Yeah it's more data. Yeah.

DR. HELDRETH: Yeah, it looks like there may be some repro concerns with that one.

DR. COHEN: Some? I didn't hear what you said.

DR. HELDRETH: Repro -- DART issues with that ingredient.

DR. COHEN: Repro. Okay. All right, so I think we're aligned on the priorities.

DR. BERGFELD: I think when we present this, it would be nice if you, the FDA, presented the reasons for bringing them forth.

DR. COHEN: Just like a sentence.

DR. BERGFELD: It would be very nice.

Full Panel – June 13, 2023

DR. BELSITO: So the FDA has asked us to move Toluene and Phthalates up for cause. And I would agree with doing that. And also, it was brought to our attention that a material that we haven't reviewed, trimethylbenzoyl dimethyl phosphine oxide, is being looked at by the European Chemical Agency, ECHA. And they're very concerned about the safety of this. It's a substance of very high concern (SVHC), and I think we should move that up on our Priority List as well.

And I think this is the type of thing that needs to be done, where we're monitoring what other safety organizations are looking at, perhaps, flagging ingredients that we weren't aware of. And we should continue to do this type of thing.

DR. BERGFELD: Any comments, Dr. Cohen?

DR. COHEN: No, I thought we might have heard from the FDA a little more why they were nominated.

DR. BERGFELD: Jan, do you want to talk about the nominations?

DR. HELDRETH: We also have Dr. Manga online.

DR. BERGFELD: Manga too?

DR. HELDRETH: She had to return to the office.

DR. MANGA: Hi, this is Prashiela. So these three ingredients came up because we've had a couple of inquiries about these being used in nails -- I'm sorry, I'm getting a bit of feedback from the room.

DR. BERGFELD: We can hear you.

DR. MANGA: So these ingredients have been noted particularly for the use in nail products. And that was why we were interested. And then, as Don mentioned, at least for the TPO, that is coming up as a new ingredient. We were concerned that it be reviewed given the other reviews that are going on.

Toluene is now one of the California Department of Toxic Substances Control products that effective January 1, 2023, nail products containing Toluene will become priority products. And, so, we felt that this was also one that needed to be looked at once again.

In terms of Dibutyl Phthalate, this is one which was included when FDA amended the food-additive regulations, to no longer provide for 25 plasticizers in various foods contact applications. They did this because the uses were abandoned, but given that this one was included in these amendments, we felt that it would be timely for CIR to review it as well.

DR. BERGFELD: Thank you very much. We're really appreciative of the FDA coming in and suggesting these particular ingredients.

DR. MANGA: Sure. Can I just make one other announcement quickly, please? Dr. Jannavi Srinivasan is there in person today, she was there yesterday. She will be serving as the FDA liaison when either Linda or I are not available. I just wanted to let you know that she'll be representing FDA when we're not available. Thank you.

DR. HELDRETH: Thank you, Prashiela.

DR. BERGFELD: Thank you.

MARCH 2024 MEETING – STRATEGY MEMO #1

Belsito Team – March 28, 2024

DR. BELSITO: Here we've got comments from Women's Voice of the Earth. That's under an Admin doc, right?

MS. BURNETT: Mm-hmm.

MS. FIUME: Yes.

DR. BELSITO: Yeah, Admin Dibutyl Phthalate. Women's Voices of the Earth, regarding this document, sent a lot of information about various state legislatures banning dibutyl phthalate and some of them also banning dimethyl and diethyl phthalate. The first question I had for you all is should we include these legislations that have passed and/or pending in our report? I didn't think we should. Comments? Allan? Paul? Curt?

DR. SNYDER: Yeah, my comment is similar to the one you just made, Don. We just need to report on the science. All of the other stuff is just kind of extraneous. I mean, we're aware of it and we try to see what their justification is if there's any science to support it, but that's -- I think we largely -- don't include that in our reports.

DR. BELSITO: Right. Curt and Allan, you're okay with that?

DR. KLAASSEN: I was just going to mention, we've often noted what's happening in Europe, for example, in our reports. I don't think we've ever said what was happening in various states. That's kind of a small difference maybe. Why do we say what's happening in Europe, like, this is banned in Europe? We will often say that, or, it's in this category in Europe. I don't feel strongly. I would just like to kind of point out a little bit of a --

MS. BURNETT: The European legislation usually has the scientific opinion attached. The Panel uses that to review what caused Europe to arrive at their conclusions. Whereas, the state legislations might not necessarily have science reports to go with it.

DR. RETTIE: All right. When we can append a citation to European material, then I think that's useful. State by state here, that would be very unwieldy unless we have a general acknowledgment that this is a contemporary issue of concern here.

DR. BELSITO: Although for toluene, I'll point out, we do talk about the California legislation.

MS. BURNETT: Right.

DR. BELSITO: We're going to need to be consistent here.

MS. BURNETT: With using -- and I think we did this in preparing the strategy memo, we did look at the California product database as they have concentrations listed that are not necessarily reported when we do the concentration of use survey. In that respect, we did use the California website to glean some numbers. I don't remember. I don't think in the memo that we put anything about the current status of being banned in California. I can't remember exactly. I do know we did use some of the numbers from the database.

DR. BELSITO: Yeah. When we get to that document, then, we'll have to be sure that we're not referencing California limits if our decision is not to reference state limits since they're not necessarily science-based. Is everyone happy with that?

DR. RETTIE: Yeah.

DR. BELSITO: Curt, Paul, Allan?

DR. SNYDER: Yes, I'm fine with that approach.

DR. RETTIE: Yes. Seems good.

DR. KLAASSEN: Fine.

DR. BELSITO: Okay. Then moving on with the other -- so without a use concentration for dibutyl phthalate in particular, and given the state regulations, the question is how would we handle this? My proposal is, first of all, that we keep dibutyl and dimethyl in this report, not just do diethyl phthalate because I don't think you can just do one as a stand-alone, and that we assume, as we typically have done, that the highest concentration that we're reporting is the limit for all of them. The concentration that we'll report for diethyl phthalate will be the maximum concentration in that product category for DMP and DBP. Is everyone hap- --

DR. SNYDER: Don?

DR. BELSITO: What?

DR. SNYDER: I'm sorry. I'm sorry. Sorry. My notes say that we had zero uses reported. We had one use at 0.1 percent in a foot powder. Is that correct?

MS. BURNETT: Diethyl phthalate has one reported use.

DR. SNYDER: Is that the foot powder?

MS. BURNETT: Yes.

DR. SNYDER: Yeah.

MS. BURNETT: No. That one is a hand and body skin product. The concentration of use, yes, is in the foot powder. There's one use reported in the VCRP and one concentration of use -- oh, two. No, there's several concentrations of use reported. The maximum is 0.12 in a foot powder. I'm sorry.

DR. SNYDER: Yeah. Okay, that's what I thought.

MS. BURNETT: Dr. Belsito, could you back up? You said you were leaning towards including dimethyl phthalate, but not including diethyl? So it'd just be two of the three?

DR. BELSITO: No, no. I'm in favor of including all three.

MS. BURNETT: All three? Okay, thank you.

DR. BELSITO: Paul? Curt? Allan?

DR. SNYDER: Yes, I'm fine with that.

DR. RETTIE: Yeah, that's fine. The ethyl and methyl are not in Annex 2 in Europe. The butyl is. Then there's a dozen of others, of the longer chain ones that are in Annex 2. I was kind of curious about how they split that out, but I'm fine with looking at all three.

DR. BELSITO: Curt, are you fine with keeping all three?

DR. KLAASSEN: Yeah, I guess. It's more complete that way and the way we usually look at these, although I was thinking, as far as economy of work, if we would just look at the endocrine and reproductive effects of dibutyl phthalate, just look at that aspect of it, which is really the essence of the concern. I think that's kind of not the way we usually do things, so I think if it's best if we probably stick the three of them together and do them again and again and again.

DR. BELSITO: Okay. Just to remind everyone, basically, this was an FDA request only for dibutyl phthalate. I think that it's hard for us to assess dibutyl phthalate without including data from diethyl phthalate in particular.

MS. BURNETT: Right.

DR. BELSITO: I would just go with all three.

MS. BURNETT: Yep. Much of the published literature is either a combination of two of them or all three or even more than that.

DR. BELSITO: Okay. Then there were a couple of questions. We've resolved the first about including all three. Then there was a question about do we have any preference how to organize all of the studies on endocrine reproductive and developmental effects? Should the endocrine studies be subheaded under Developmental and Reproductive Toxicity? Or should the studies be classified in the stand-alone section? I thought they should be subheadings of DART. Anyone else have thoughts on that?

DR. SNYDER: I think it can go either way. The writer needs to see how it's easiest for them to organize it.

DR. BELSITO: Christina, so then it's going back to you. What are your recommendations?

MS. BURNETT: I think subheadings under DART make more sense because the way some of the studies that I've seen so far have been set up, you'll start off as a traditional DART. Then they'll focus in on testicular development or the anal-genital distance, so it will be like a two-part study. Previous iterations of this report, the data was kind of patch-worked throughout the report. For me, it was hard to follow in order to summarize it. I feel that everything under DART makes sense and it is interrelated. Again, I'll defer to you.

DR. SNYDER: Well, I think it's how you think you can organize it best for us to review.

MS. BURNETT: Okay.

DR. SNYDER: I think you summarized it there. I'm fine with that.

MS. BURNETT: Right. It's going to be -- I assume, since there's so much data, that tables would be very helpful. It would be subheaded tables under DART. I mean, obviously, whatever I put together, we'll review at the next iteration with all the data, and you can tell me whether it worked or not.

DR. BELSITO: Okay. And then the quest- --

DR. SNYDER: I have faith in you, Christina. I have faith in you.

MS. BURNETT: Thank you.

MS. FIUME: We all do.

DR. SNYDER: Yes.

MS. BURNETT: I will say that I am enlisting help with Jinqiu on this one because some of it's a little over my head this time.

DR. SNYDER: I just want to know how you draw the short straw all the time on getting these difficult ones.

MS. BURNETT: I don't know. Then with all this new data, there's a lot of zebra fish studies and different methodologies that we have not included before. Since we are interested in alternative methodology, I am not -- I mean, I get the gist of these zebra fish studies, but I don't know what the important information of methodology and outcomes are.

DR. BELSITO: I've never heard of that use of zebra fish, so I actually did a -- not a deep dive, but a medium dive into zebra fish and DART. To summarize, what I got was the high structural similarities of the reproductive axis of zebra fish as

compared to humans has resulted in zebra fish often being utilized as the model in the assessment of reproductive toxicity of environmental chemicals, including herbicides.

MS. BURNETT: Right.

DR. BELSITO: A lot of the zebra fish studies were done by the EPA. Without seeing the data, I think we can use the zebra fish data for DART endpoint, but open it up to other members of my panel.

DR. RETTIE: Yeah, I'd be curious to hear -- I was just curious to hear what Paul and Curt had to say about zebra fish studies for DART. I don't know too much about that.

DR. SNYDER: Well, I think it's been a pretty much high level (audio skip) way to look for a signal. Now, they're -- because the model is showing some utility, it's being applied in other instances. I think we're obligated to have a look-see --

MS. BURNETT: Okay.

DR. SNYDER: -- and see if it matches up with other data. We have a lot of data. It's not like we're only going to rely on these non-mammalian studies.

MS. BURNETT: Right.

DR. SNYDER: If it collaborates what we already have, I think we need to look at some of it, yeah.

MS. BURNETT: Okay. Should I just focus on the zebra fish studies that give DART data? Or are there any other endpoints that they're possibly are useful or just ignore it, just focus on the DART?

DR. BELSITO: I would just do DART data.

DR. SNYDER: Yeah, just endocrine effects, yes.

MS. BURNETT: Okay.

MS. FIUME: Paul, just to take that --

DR. BELSITO: Curt, you're okay with the zebra?

MS. FIUME: Oh, this is Monice. I was --

DR. KLAASSEN: Yes, that's fine.

MS. FIUME: I was just going to say, to take it a step further than for generally all reports, if we find studies on zebra fish, we should bring them into the documents for the Panel to take a look at?

DR. SNYDER: Yeah, well, I think so, especially if it's the ingredient --

DR. BELSITO: Yeah.

DR. SNYDER: -- yeah, for repro, yeah.

MS. FIUME: Okay. Thank you.

DR. SNYDER: Do you agree with that, Don?

DR. BELSITO: Yes.

MS. BURNETT: Yeah. If we -- not to skip back to the previous report, but in the pyrogallol table, I do know that there were catfish studies, which were new. We've not included catfish studies. In this case they were acute and short-term toxicity studies, not DART studies.

DR. SNYDER: Yeah, that's a different issue for me. Those are different, yeah.

MS. BURNETT: Okay. I'll include it for now. Then you can tell me, once you see that first draft, whether you like it or not.

DR. BELSITO: Yeah.

DR. SNYDER: Sure.

DR. BELSITO: Also, my recommendation for particularly the DART data since there's so much of it out there is to keep as much in tabulative format and limit the text.

MS. BURNETT: Okay.

DR. BELSITO: Basically, refer us to a table.

MS. BURNETT: Okay. Yep.

DR. BELSITO: Not say too much in the text.

MS. BURNETT: Okay. That kind of makes some things easier.

DR. BELSITO: On PDF page 3, where you say, "For further consideration, state CIR staff have calculated the dermal exposure level of diethyl phthalate from various product categories based on a Council survey," --

MS. BURNETT: Yes, that was --

DR. BELSITO: -- my question is why --

MS. BURNETT: I'm sorry. Go ahead.

DR. BELSITO: Why are we using the survey when we have California data suggesting a higher use? Shouldn't we be taking a more conservative approach and using the highest level reported in the California data.

MS. BURNETT: I believe Jinqiu Zhu did calculate a second one using that. If you go past all the calculations on PDF 4, he did add on a sentence saying, "In light of maximum use concentration reported to the California product database, the calculated MOS 86,206 for the bath products at one percent --

DR. BELSITO: Okay.

MS. BURNETT: -- and 10,489 for skincare products at 0.2 percent."

DR. BELSITO: Yeah, I know, but that's sort of -- I mean, to me, it's sort of a little -- I mean, it's important, but it -- I almost would rather not have to make that statement --

MS. BURNETT: Right.

DR. BELSITO: -- and basically say at the beginning, up front, that we have a level from the Council survey. However, we had information from California of higher use in these product categories. As a conservative estimate, we're using this higher level of use.

MS. BURNETT: Okay.

DR. BELSITO: Then go through all the data --

MS. BURNETT: All right.

DR. BELSITO: -- rather than just adding a sentence at the end.

MS. BURNETT: Okay. We'll flip it. We'll recalculate it and show the calculations using the California data.

DR. BELSITO: Right.

MS. BURNETT: Okay. Yeah.

DR. BELSITO: Then the next one was, does the Panel prefer to request survey data on additional phthalates and perform further calculations? This survey was just on diethyl phthalate, right?

MS. BURNETT: Part of this calculation that Jinqiu Zhu recreated was from the re-review from 20- -- that was published in 2005. That's the same calculation, I think, was in there or was similar to that that the Panel put into the re-review summary.

DR. BELSITO: Okay.

MS. BURNETT: Some of the assumptions and stuff were based on what was done then. It might not be what is currently done in our margin of safety calculation. Jinqiu Zhu can update that if you like.

DR. BELSITO: Yeah.

DR. KLAASSEN: I would like to emphasize again, I did not like to see five significant values for 28,846.

MS. BURNETT: Okay.

DR. KLAASSEN: Let's change that to at least 28,800.

MS. BURNETT: Okay.

DR. RETTIE: Three significant figures.

MS. FIUME: Right, Christina, because we have the survey data for those other two phthalates.

MS. BURNETT: Yes, we already --

MS. FIUME: Yeah.

MS. BURNETT: Yes. We already -- or Council provided that a little while ago. We could ask them, I suppose, if there's any updates to it.

MS. FIUME: It was part of the 2023 survey, correct?

MS. BURNETT: Yeah.

MS. FIUME: Yeah.

MS. BURNETT: It's probably as up-to-date as we're going to get.

MS. FIUME: Because was that sentence to ask were there any other additional phthalates? Or is it still referring to the two we're adding in?

MS. BURNETT: I think, internally, we had that discussion about the -- because the very last re-review included benzyl butyl phthalate -- or butyl benzyl phthalate. We're not including that in this query because it's not quite structurally the same.

MS. FIUME: Okay.

MS. BURNETT: That might have been a -- we might not have updated it from that decision.

DR. BELSITO: Okay, yeah. I mean, I think we should just do the phthalates we're reviewing, not other phthalates.

MS. BURNETT: Right.

DR. BELSITO: Okay. We're clear on where we're going on this, Christina?

MS. BURNETT: I believe so.

DR. KLAASSEN: There's a lot of data out there.

MS. BURNETT: Oh, I know. I also have two large archive boxes sitting in my house from the previous reviews with all the data in that too, so, yeah, there's a lot.

DR. BELSITO: Okay. Should be fun.

MS. BURNETT: Mm.

DR. BELSITO: Any other comments on phthalates for Christina before we move on?

DR. SNYDER: None for me.

DR. RETTIE: All good.

DR. KLAASSEN: I'm good.

MS. BURNETT: Thank you.

DR. BELSITO: Okay.

MS. BURNETT: Thank you.

DR. BELSITO: Okie doke.

Cohen Team – March 28, 2024

DR. COHEN: Okay. The last one is dibutyl phthalate. So, the FDA requested dibutyl phthalate -- for us to look at this -- and we got a number of questions back from the staff that we can go through on dibutyl phthalate, but before I open it, I just wanted some clarification because it seems like we could spend a lot of time on this. Is it true, did PCPC agree on banning dibutyl phthalate for the Maryland House Bill 643 that banned dibutyl phthalate amongst a series of other chemicals on Bill 643?

And the reason I'm asking is, like, it's like a patient that's DNR. You don't start doing cat scans and biopsies on it when there's a DNR order on it. So, if PCPC has signed a DNR order on dibutyl phthalate do we go through the exercise of reviewing this a great deal?

DR. EISENMANN: We are no longer supporting this ingredient, yes.

DR. COHEN: Right. And I saw the questions to us that we need to be impartial, and we need to look at the scientific data and to what end? Why are we doing this? Because I did pull the PCPC note to the Maryland legislature, and I looked at the House Bill 643 and it's like they don't support the product anymore. Why can't we just indicate that we're no longer -- use not supported at that point. I know there's more to it than that but that's my visceral response.

DR. EISENMANN: No, I thought that it would be nice to have another category for re-reviews that said something like publish some kind of a statement that says this ingredient is no longer in use. This is a historical, we are no longer going to -- so in other words, your old conclusions don't stand anymore because it's hard when you've published reports, what to do the

original reports that you're not going to review it anymore. You're not going to make a determination whether it's safe or not and you're just not going to review it anymore because it's not used. But there isn't a process yet for you guys to do that.

DR. BERGFELD: We could send a note to the steering committee and request that that be considered, and we can put this on hold.

DR. COHEN: Table it. There's a lot of very important questions from the CIR staff here and a lot of detail and there's a margin of safety. But, again, I don't want us to spin our wheels for no reason. It would be nice to have another category where our recommendations are withdrawn, and we will not re-review. No longer supporting use.

DR. HELDRETH: So, would we want to make it similar to an insufficient data conclusion that's based on zero use where we eventually are saying use not supported? Because, I mean, we're not saying unsafe, we're just saying we don't have enough information to support the safety of this.

DR. EISENMANN: Well, I'm not sure you're reviewing the safety anymore. I mean, we're not asking you to look at the new data, you're just -- it's not used anymore.

DR. HELDRETH: And that's the same situation with insufficient data, zero use.

DR. COHEN: But there's a lot of data on this.

DR. EISENMANN: Right. There's a lot of data on it and one thing -- we really should not use data on diethyl phthalate to estimate exposure for dibutyl phthalate. The uses for dibutyl phthalate were totally different than how diethyl phthalate was ever used so you really don't really have any use information to base anything. You should not use the dye -- I think that was one of the questions.

DR. COHEN: Yes. It is one of the questions.

DR. HELDRETH: Right. So, we have a lot of data, but we are insufficient for use and concentration data.

DR. EISENMANN: Correct. Which is why I hate to see you spend a lot of time on it because you're never going to get that information because they're not using it anymore.

DR. HELDRETH: Right.

DR. EISENMANN: So, there's nothing to hang your hat. There's no exposure to hang your hat on for this ingredient anymore.

DR. ROSS: It's not unusual that we would look at ingredients that we don't have concentrations on but what is very unusual with this one is that it's not being supported anymore, which is now is common sense.

DR. COHEN: That's the point.

DR. ROSS: Yeah.

DR. COHEN: Look, we just finished one that has one use and we're going to go through the whole exercise, right? And so, in this case we can't say there's no data. There's an abundance of data and we would review that. So, if we went out with an IDA, what's our IDA? We want concentration of use and use, we don't get it, but then we still have this massive report to review at the next iteration. So, I just didn't want to -- that's why I pulled the PCPC letter and the House Bill because I thought this was very different.

And I don't think we need to sort of fall on our sword that we can review the data scientifically for absolutely no practical purpose.

DR. HELDRETH: Right. Right. So, then would the request be to the Steering Committee that when we get things like this in front of us the Panel can simply say the use of this is not supported. I mean, it goes into that same bucket. And then, if somebody decides later, ten years from now, I want to use dibutyl phthalate then they're left with the impression well, I need to provide data and information to the Panel to reopen that. I mean, it's the same kind of situation.

DR. BERGFELD: I think we would have to ask the Steering Committee for specific declaration here.

DR. COHEN: Yeah.

DR. HELDRETH: Right.

DR. BERGFELD: And have it call it something specific and then have a mechanism developed of what would be said and what would be retracted. So, I think we have to just table this until we get an answer in the Steering Committee if that would be allowable.

DR. HELDRETH: Yeah.

DR. ROSS: It's certainly interesting points from WVE with all the pending state legislature actions that's compounded. Maybe you want to bring in those at some point, too. But anyway, I totally agree with your proposed way forward, Wilma, I think it's a good way to go.

DR. COHEN: And the WVE was instrumental in the way we adjudicated this one in my opinion because I went and looked at all those links, the state laws, and so good. So, we're going to table this.

DR. HELDRETH: Can I ask since we have FDA representatives here, would it be useful for our Steering Committee to agree to and our Panel to eventually issue that the use is not supported for this because FDA was the one that asked for us to look at dibutyl phthalate again. So, I want to know, is this responsive in the way FDA was hoping for if we moved forward with that?

DR. MANGA: Bart, I'd probably have to take this back and maybe I can give you an answer tomorrow.

DR. HELDRETH: Yeah. That would be great. We want to have good customer service. You asked for this and that's why we're looking at this.

DR. MANGA: No, I appreciate that, and I appreciate that you guys took this up. So, yeah, happy to provide you with an answer tomorrow.

DR. HELDRETH: Wonderful. Thank you, Prasheila.

DR. MANGA: Sure.

DR. COHEN: Okay.

DR. BERGFELD: Could I ask Bart a question?

DR. HELDRETH: Sure.

DR. BERGFELD: Bart, I don't have at home the Steering Committee's sort of document that we have. I wonder if you could maybe copy the part where we might fit this in and just include it in some kind of mailing tonight?

DR. HELDRETH: Oh, you mean get this out to the Steering Committee tonight?

DR. BERGFELD: No, not tonight. To the Panel where this might fit. It's quite a document that gives us great guidance and we'd be looking at some kind of report that allows it to go as use not supported where that would fit. What that'd be called and where that would fit.

DR. HELDRETH: So, you mean the excerpt from the CIR procedures? Yeah, yeah. I can put something like that together and I can even present it in my report to the Panel tomorrow if you would prefer that.

DR. BERGFELD: That'd be great. That would really be great because it gives just another mechanism to do with something like this.

DR. HELDRETH: Okay. I'll do that.

DR. COHEN: There used to be something in the resource about the types of conclusions we had, I remember.

DR. BERGFELD: Yeah, it's documented in pages. I just don't have it here to look at.

DR. COHEN: Yeah, no. I just couldn't find it immediately. Okay.

DR. HELDRETH: Yeah, our conclusion types are safe as used, safe with qualifications, unsafe, and then there's insufficient data that breaks it into two pieces currently. If there are uses and we get no data response from two years of a final safety assessment, that's use not supported. If there are no reported uses after two years on an insufficient data conclusion then that's converted to zero use. And when we get a zero use one, we never look at that again unless somebody brings us data to say reopen it.

DR. BERGFELD: So, it falls somewhere under that category.

DR. HELDRETH: Probably. Probably. I mean, I think that's what we would bring to the Steering Committee. Do you want to use this exactly or do you want another separate procedure that fits right next to this in a similar fashion?

DR. COHEN: That would go under -- that would mean that we're bringing this to a draft report and issuing an IDA, right?

DR. BERGFELD: Not necessarily.

DR. HELDRETH: No. I mean, ideally this could be brought to you as one of these abbreviated re-review documents like we just did for pyrogallol. You could say, hey, nobody's using this. There's no support for this. Why should we move forward with it? And if the Steering Committee provides us with a process to go forward that'd be the end of it. You'd just kick it into that bucket, and we wouldn't think about it again until somebody came forward and said, hey, hey. I'm using this and I have data.

DR. COHEN: So, under the current mechanism what could we do? What could we say, there's no use --

DR. HELDRETH: Yeah. At the current mechanism with just those conclusions that we have available for us, we would have to trudge through to a final report with an insufficient data conclusion, wait two years --

DR. COHEN: No, no. Forget it.

DR. HELDRETH: Right. Right. So maybe we can draw a new pathway on that little flow chart we brought up earlier. It goes right from that bucket right from the abbreviated re-review.

DR. COHEN: Okay. I heard the read across discussion, at least, the back half of it, and there's no point in reiterating it now and it'll already be discussed tomorrow. Is there anything --

DR. BERGFELD: Did you want me to summarize that I have here before me?

DR. COHEN: Sure, sure.

DR. BERGFELD: Okay. The read across working group met. It's composed now -- (audio skip).

DR. COHEN: We lost you. We lost your audio for a second. For a second.

DR. BERGFELD: Okay. I said the read across working group met and the new chair will be Allan Rettie. Members of the working group is Curt Klassen, David Ross, and Susan Tilton. Ad Hoc is Don Bjerke. Consultant to be invited is Dan Lieber. The movement or the summary of what they will -- examination of the need for a read across to collect the CIR past read across documents and see how we've handled those. And three, look at the read across documents that Europe and the reach group uses.

And four, if the CIR has a read across document presently, which is a living document that will ultimately have some changes to it. So that's what we did.

DR. COHEN: Nice. It seemed like that thing could've gone on for hours, but it was more succinct. All right, so are we calling it a day?

DR. BERGFELD: We are.

Full Panel – March 29, 2024

DR. BELSITO: Yeah. So, this was brought up for cause by the FDA for an earlier review and fast tracked. And several questions have been posed to us. First, does the panel want to include, along with dibutyl phthalate, diethyl and dimethyl phthalate in our expedited review. And my team felt we should include these because it would be difficult to evaluate dibutyl phthalate in the absence of information from the other phthalates.

Did we have any preference on how to organize the myriad of endocrine and DART studies, and we felt that it should be as much as possible in a tabular form rather than in written form. Whether we would accept studies from zebrafish. And in my review, actually the zebrafish have been said to have high structural similarities with the reproductive axis of humans and they're utilized as a model so we would accept that.

Apparently, there were some studies on catfish that we decided we would not accept. And that's where we were to include diethyl and dimethyl phthalate along with a dibutyl phthalate re-review and go from there.

DR. COHEN: So, Don, we took a completely different view on this. Ultimately, we thought we would table this issue, primarily because dibutyl phthalate has no reported uses. In March of 2021, the PCPC wrote a letter in favor of Maryland House Bill 643 banning dibutyl phthalate. And so, we have no uses, we have no support by industry for the use of the product, and we thought that taking this memorandum forward to do a really comprehensive report on this would be a complete waste of time.

And so, we thought if we table dibutyl phthalate until Cosmetic Direct comes around, perhaps, to see if there's any uses, we can consider doing that but in the face of it not being used, we wanted to just spare the workload in just not doing this.

DR. BELSITO: I guess my only concern, and I'll let Bart and/or Monice, and FDA comment. I mean, this came to us as a request from the FDA.

DR. COHEN: We asked them about it, and we were hoping to get a response today about that.

DR. BELSITO: I just feel like we've been asked to do it. And then to turn around and say we're going to table it, is probably not the right thing. And I guess this is now the time to raise the discussion that was pointed out at the beginning of our meeting, as to if there are no reported uses for a material it goes into a new category. So perhaps we should have that discussion now.

And if this were to go forward, you know, that MoCRA effort for mandatory is supposedly July. But we probably wouldn't get this back until September or December anyway. In which case, at that point, hopefully, if it were to come back in December, we would have definitive information via the MoCRA regulations of, you know, is there existing use out there. If we table it, then we're no closer to getting it on the agenda in September or December.

DR. BERGFELD: Comment, David?

DR. COHEN: Well I was hoping to get a comment from FDA on what they felt about moving this along. Someone has raised their hand, I just can't tell who it is.

DR. MANGA: Hey, David, it's Prashiela. Hey, good morning, everyone. So, FDA would be okay with tabling it. Just to give you some perspective of where we are with Cosmetics Direct, as was mentioned, July is the deadline. We're still working on the system. As you know, it was an unfunded mandate so we're working in a resource-limited situation. We are developing our backend systems, and we hope to have that up and running as soon as possible. So, I can't really give you a timeline on when we'd be able to get you the data. That's sort of where we are.

In terms of tabling it, we'd be okay with that. If in terms of putting it into a new category, we'd be interested to hear what that new category or moving it into sort of a hold status, what that would look like. And I'd be happy to comment once we've had that discussion.

DR. BERGFELD: Bart, can you add to this?

DR. HELDRETH: Sure, absolutely. So, as I discussed in my Director's report, there was some discussion on the Cohen team about petitioning the CIR Steering Committee to create a new category for re-reviews in general, not just the dibutyl phthalate. But re-reviews in general where the evidence in front of the panel demonstrates that it's not in use, and essentially discontinued use in the U.S. since the last time they reviewed it. And the idea being that this is, essentially, insufficient data points. And if that's not available, and it seems like listening to industry it looks like it will not become available, then should the panel really carry on with the whole process of opening up a draft report and going through all the way to a final report.

Or could the Steering Committee approve a new category for such incidences, where a re-review summary would come out with a new conclusion of use not supported. The verbiage, of course, can be different than that, but the idea is that we don't have the data to support the safety because we don't know how it's used and what concentration it's used at.

DR. COHEN: Don, I mean, so I haven't seen before where the PCPC has been supporting a ban on a product, and of course no uses, and we would go ahead and deliberate on this. For what endpoint? You know, the industry themselves have supported the Maryland ban.

DR. BERGFELD: Don, can you respond?

DR. BELSITO: Yeah. I mean, I guess, my only response to that is it would've been nice, I think, for PCPC to at least notify the Panel that they were going to do that.

DR. BERGFELD: Well, I think they did it before we even started on this.

DR. COHEN: This is from 2021.

DR. BELSITO: No, I understand. As a member of the Expert Panel for Cosmetic Ingredient Safety, I just think it would be nice if we were all informed when PCPC is going to issue statements in support of banning cosmetic products that we, as a panel, had previously said were safe as used.

DR. BERGFELD: We can write another letter requesting that in the future. Mm-hmm. I think we need to do that officially. It's an expectation that we have.

DR. COHEN: That's a good point, Don.

DR. BELSITO: Beyond that, if the FDA who brought this petition to us is okay with us tabling it, I think it's fine. On the other hand I think that it's something that, if anything, we need to give our imprimatur to the actions that PCPC took and bring this up. Regardless, I think that diethyl phthalate is still used and there's a lot of controversy about that. And so I think that the phthalates in general should undergo an expedited review.

And I would just like to move it forward to look to getting it on the December agenda. Hopefully by that time we'll have -- Cosmetic Direct will know exactly how many of these phthalates are being currently used and what the current concentrations are, and we can come out with an informed decision.

DR. COHEN: Yeah.

DR. BERGFELD: So, what I'm hearing you say is you don't like the word table? You would rather have the word--

DR. BELSITO: Yeah, because that puts it in limbo and we're no further along.

DR. BERGFELD: Yeah.

DR. COHEN: Well, but we're tabling it pending perhaps a new category. And I certainly support what you're saying about looking at the other phthalates, as long as it didn't contain dibutyl phthalates in it at this point. Because it's just going to create a tremendous amount of work for no useful end product.

DR. BELSITO: Okay. I mean, Paul, Curt?

DR. SNYDER: I think we have to keep in mind largely what we do is for the manufacturers, and we have a report out there that says it's safe as used. And that's not withstanding an issue. So even if we table it, they're going to defer to the last report, and it says safe as used. So, I think, I'm not quite certain, it has to be very clear as to why we're tabling.

As usual, I'm supportive of Don and I don't like to leave these things hanging. Let's just keep moving along and get them off our plate instead of piling our plate up with all these loose ends. And if we can move it along, I think we should do that.

DR. BERGFELD: Well, we can postpone until we have the Cosmetic Direct information. And supposedly, we might have some of that in December.

DR. COHEN: So, Paul, I certainly understand what you're saying, but if we move this along the staff now starts writing a draft report on this, right? So that I don't want to do. Number two, what we're looking to have is use not supported which resends the old one, right, we just can't do that right now.

DR. SNYDER: I guess I'd be more supportive of that if we knew it was going to be in a relatively short timeline. I just don't want this to linger out there and then kind of get lost in the weeds and then -- you know what I mean? We've got to just bring these things to fruition.

DR. BERGFELD: Bart?

DR. HELDRETH: Since, instead of having a re-review/abbreviated summary with a question to open or not or a draft report that we're proceeding in the process here. Since we have only a strategy memo here, I think maybe I'm hearing that the Panel has some consensus that if it came back in December, with the all likelihood of Cosmetics Direct being able to provide us with some more sound numbers on frequency of use at the very least, that that would be good timing.

So, we could just agree to -- if it is truly the Panel's consensus to adding these two other phthalates, adding the endocrine and DART end tables, looking at the zebrafish and not looking to the catfish studies, and including that Cosmetic Direct data that you would see a draft amended report on these three ingredients in December. So, we don't have to worry about tabling or not, it's just a matter of that will be the timing of it.

DR. ROSS: Bart, could I ask a question? How quickly do you think we'd get an answer back on the new category not supported? I mean, would that be next week, next month, three months' time? I mean, what's the timeline?

DR. HELDRETH: You're talking about the petition to the Steering Committee?

DR. ROSS: Yeah.

DR. HELDRETH: So, first we'd have to schedule a meeting of the Steering Committee which doesn't always happen very quickly. And I was hoping to have a meeting of that full Steering Committee instead of just posing a question on them. So, yeah, that would likely be somewhere around the September meeting, I would guess, we would be able to get something scheduled.

DR. ROSS: Okay.

DR. BERGFELD: I think it's two-prong here. We're going to go forward with this other category with the Steering Committee, that's a given. The other part is do we go and update this draft, as proposed by Bart, and nothing extra, and put it on the December schedule for review? Thinking that we will get the Cosmetic Direct information about use and frequency, and perhaps at that time we'll also have this other category that we could possibly use.

DR. BELSITO: I mean, first of all, we don't even need that other category by December, because it's going to be our first go around, right? And then we're going to have another 60 days comment period. So, if we start this in December, my best guestimate is we're not going to be signing off on this until June of 2025.

DR. BERGFELD: Probably.

DR. COHEN: So, I'm just going through this mind exercise. We get Cosmetic Direct back and there's some uses there. We have PCPC supporting a ban of the product that some states have enacted already. And we're going to go ahead and issue a safe as used if we --

DR. BELSITO: We don't know what we're going to issue yet, David. We've got to look at all the new endocrine data. Right now, we have a safe as used conclusion for dibutyl phthalate.

DR. COHEN: I understand that. I understand. I'm just looking at the practical aspects of this.

DR. BELSITO: We'll have to struggle that when we look at the data, right? We haven't looked at all the data yet.

DR. COHEN: The question is, is the assembly of the data for dibutyl phthalate a reasonable thing to have the staff and us do? That's what I'm --

DR. BELSITO: Yes. Because just for the reason you said, PCPC has supported the ban and yet our current document says it's safe as used.

DR. COHEN: That's why if we have another option, we can finish this without all that.

DR. BELSITO: We can't exercise that as an option without reviewing the data.

DR. ROSS: You don't have that category yet anyway.

DR. COHEN: We don't.

DR. ROSS: Yeah. So that's the issue. I mean, you could exercise (audio skip) .

DR. BELSITO: We don't have the absolute data that it's not used.

DR. ROSS: It's not supported. The use isn't supported, I guess, that's a different conclusion as you pointed out.

DR. COHEN: Right. So right now, we have zero uses for dibutyl phthalate and dimethyl phthalate.

DR. BERGFELD: Other than the workload that you're concerned about, what is the other issue here?

DR. COHEN: Well, number one, the workload is not an insignificant issue with an item like this. And two, we go through the exercise and regardless of what we find on this -- if we find that we can support its use, it seems silly to me, right, that we would support its use in a situation where industry has recommended a ban. Okay. And so that's incongruous to me.

DR. BELSITO: How do you know we're going to support the use, David? We've got to look at the new data. We have to look to see what the concentration of the use is. Right now, it's a voluntary system. And we can't assume it's zero just because someone hasn't voluntarily reported that they're using it.

DR. COHEN: Okay.

DR. ROSS: This gets back to the mandatory reporting. We're sort of winding ourselves around the axel again here with respect to coming back to that same point. But if we're not going to have that -- it sounded like we're not going to have that, necessarily, even in December.

DR. COHEN: Can I go back to --

DR. BERGFELD: We will have why the PCPC banned it, they should be giving us that information. What did they put together to make that decision? What was the information?

DR. COHEN: That would be helpful, but I've been sort of countered with precedent issues at this meeting today. And so, can you give me a precedent issue where industry has supported a ban on a product, and we've gone ahead and decided to adjudicate it and come up with a decision on it.

DR. SNYDER: My only comment to that, is we're the expert Panel for the safety of cosmetic ingredients, we're the driver. And so we need to do what we think is best in our expert opinion in moving these ingredients along and what stays and how we do it. I think we sometimes lose track of that, and we get influenced by all these outsiders and things. But we, ultimately -- the Panel -- needs to decide what we think is in the best interest for the safety of these ingredients for the consumer.

DR. ROSS: Yeah. I agree with that.

DR. COHEN: I agree with that.

DR. BERGFELD: We still need to know the data or the information regarding the PCPC statement, why they did it and what drove it. We just need to know. And we need a letter to them saying when they do such things they should at least announce it to us and give us this information.

I see we're sort of at an impasse right now and I think this is only a strategy type of discussion. So, Bart, what do you think the strategy is?

DR. HELDRETH: Yeah. Since it's a strategy memo it really just comes down to a consensus of the Panel for us to move forward. I think what I'm hearing is that we should proceed with a number of things. We should proceed with waiting, to some extent, until we get the Cosmetic Direct data to come in and see if there are uses or not. And once that information is available, proceed with a draft amended report for dibutyl phthalate/dimethyl phthalate/diethyl phthalate, including some of the data and categorization of the data as was questioned in the strategy memo.

But then at the same time, also, petitioning the Steering Committee about the future use of a new conclusion category for re-reviews of ingredients that are not in use. And a letter to industry requesting some explanation of why they were requesting this ban in the state of Maryland, and to communicate with this Panel about future motions in such directions.

DR. COHEN: Yeah.

DR. BELSITO: I agree.

DR. BERGFELD: I think that's right on. Anyone disagree with that?

DR. ROSS: Just some editorial addition there, Bart. You know, scientific rationale for these bans, which it would allow us to come to some balanced opinion when we actually see all of the data. In our team meetings yesterday, just getting back to one issue, there was some discussion and I think it came from PCPC, I may be wrong there. But to not include the additional phthalates. But that's where we have concentration of use data. So I had, actually, in my notes to include them, but there was some discussions there not to include them.

DR. EISENMANN: The discussion is not to use the use --

MS. KOWCZ: in read across.

DR. EISENMANN: Not to use the use information for diethyl for dibutyl because historically dibutyl was used very differently.

DR. ROSS: So, it's not, not include them -- using the double negative. It's just --

MS. KOWCZ: Yes, we just wanted to make sure that it's just the singular ingredient, Dr. Ross.

DR. ROSS: Okay.

MS. KOWCZ: And I just wanted to reiterate real quick that a lot of the times the decisions, as you've said, are not scientifically based. The bans are -- a lot of times we fight for not having ingredients banned because the science speaks otherwise. But we lose the battle, and so we'll definitely take that back to our internal staff and we'll definitely try to do better on the legislative piece. Because, Bart, we've also not had a Steering Committee in, I think, two years, so I think we're definitely due for that. And so maybe this is a good time to discuss that as well.

We are very open with what we can and cannot share. And a lot of times, as you said, Dr. Ross, the bans are not based on science. We're not happy about them either. Just wanted to share that point. Thank you for letting me say that.

DR. ROSS: Excellent comments. And I think we'd like to see the science behind those bans. And since I have to wade through a lot of those data with respect to the endocrine, I think for me it would be best as a separate section just after the DART. Many, but not all of the endocrine effects are repro. So, you do have separate ones.

I think it's best to make it a separate section. And I think it was Paul commented, or Don commented, that they wanted to see that in tabular form, and I'm totally in agreement with that. It's an easier read. So.

DR. BERGFELD: All right. I think we've done our duty here. I think we're going to move on now.

JUNE 2025 MEETING – STRATEGY MEMO #2

Belsito Team – June 9, 2025

DR. SNYDER: All right. We'll move on to the Phthalates. The Phthalates were brought up in the 2024 Priorities List in the 1985 report. They were safe for topical application. In 2002, we reaffirmed that conclusion. In 2012, we looked at it, we did not reopen.

In 2024, due to a large number of new studies with repro and developmental effects, endocrine effects, we've now opened this up. And it's on Admin Page 9 of 19. The question is, do we want to have Dimethyl Phthalate as a separate re-review proposal document, or would you prefer to have this ingredient in a safety assessment with Dibutyl and Diethyl Phthalate? I think that's the question, right?

DR. BELSITO: Well, no, there was no reported uses, right? The question was whether to drop it?

DR. SNYDER: Drop "it" as in which one, the Dimethyl?

DR. BELSITO: Yeah. Wasn't that the question?

DR. HELDRETH: Yeah, the idea was, since we're here looking at these again, and there's no reported uses for the Dimethyl phthalate, and it seems like this is something that's being abandoned altogether, why not just kick it out into a use not supported re-review summary?

DR. BELSITO: My comment on that, Bart, was, if we can read across, why are we getting rid of it? Because number one, with MoCRA, we know that not all companies have to report, right? So it could be being used, just not reported. And number two, that with the VCRP, it was all voluntary, so it could be being used.

So, we really don't know that it's not used. We just don't have reports in the RLD that it's used.

DR. SNYDER: True. So, how do you want to handle then?

DR. BELSITO: I wanted to keep it in unless you thought we can't read across.

DR. RETTIE: I think we can read across from Dimethyl to Diethyl, at least. So, it should be fine for read across.

DR. SNYDER: Okay.

DR. BELSITO: And, Christina, you were asking for an expert. I have the perfect person; he's in the Netherlands. That would be expensive to bring him over, but he might do it virtually.

His name is Aldert, not Albert, A-L-D-E-R-T, Piersma, P-I-E-R-S-M-A. He's in Utrecht, Netherlands. And his, email, if you want it, is Aldert H. Piersma, so, alderthpiersma@gmail.com. He's the repro endocrine expert on the RIFM Panel as well. And if you do a lit search on him, he's done a huge number of publications on phthalates.

DR. SNYDER: Okay.

MS. BURNETT: Thank you.

DR. BELSITO: And I think what we're asking him is to comment on the validity of whatever studies we're being presented with. Because all of these endocrine disruptions, so many of them, the systems aren't really -- what's the word I want -- representative for humans. The doses are extraordinarily high. He would be the person who would be able to comment on these studies that we're going to get. Because I think it's going to overwhelm us.

DR. SNYDER: I think that's a great idea.

DR. KOWCZ: We also have an alternate, if you're interested. We have Dr. George Daston, who has also presented to the Panel before, from Procter & Gamble.

DR. SNYDER: What's it -- spell his name?

DR. KOWCZ: George Daston. D as in Daryl, A-S, as in Sam, T as in Tom, O, N as in Nancy. You know, if Aldert can't make it, or if for some reason the timing doesn't work, we also have George Daston. He has presented, Paul, numerous times.

DR. BELSITO: He would be a lot cheaper to have in person.

DR. KLAASSEN: I have two people as well from the United States. Paul Foster, who has done a lot of this work. And in fact he's done it with Earl Gray. I think they have been -- they're both very recently retired. They've done a lot of work with these phthalates. Anyhow, that's a couple of other names, so you got plenty of names.

DR. SNYDER: Okay.

DR. BELSITO: Well, maybe it would be good to have -- I mean, since we have three people in the US, to have them actually see if they can physically come to the Panel because we've seen the issues with virtual presentations. But, to have two of them so we get more than one person's ideas. And, of course, if they both agree, that makes it very much easier for us, right?

DR. SNYDER: Well, I like the idea of having George come if he's presented to us before. Because he would have a history with us and -- yeah.

DR. KOWCZ: He does have a history with you.

DR. HELDRETH: Is George not in the EU right now, too?

DR. KOWCZ: I'm sorry?

DR. HELDRETH: I thought George was in the EU as well.

DR. KOWCZ: No.

DR. HELDRETH: No? Okay. That must have been --

DR. EISENMANN: I gave him the dates, and your next two meetings are virtual. Of those two meetings, the December one would work for him.

DR. SNYDER: Okay.

DR. BELSITO: Oh, okay.

DR. SNYDER: All right.

DR. BELSITO: Do we want to see if one of Curt's gentlemen or Aldert?

DR. SNYDER: I think so. I like the idea of getting two experts.

DR. BELSITO: Or if Albert could make the virtual meeting in December?

DR. SNYDER: I think so. I think that's a good -- we'll just -- and if we only get one, then we get one. But if we can get two, that's fine, right? Virtual, it's not that -- we're not flying anybody out or anything, so let's do it.

DR. KLAASSEN: I mean, this topic keeps bouncing up, not just with the phthalates, but in general. And it would be good to have two or three people talk about it and see what they say. That's my feeling. But I don't know how many of these would even be willing to do it, and what their schedules are, so whatever.

DR. RETTIE: A point of education for me is, has this Panel used the EPA's screening program for endocrine disruptors before? Is that something you've considered? They have a whole pile of assays that I found out: receptor bindings, steroidogenesis, thyroid function, all that stuff.

DR. SNYDER: No.

DR. RETTIE: We haven't typically looked at that before?

DR. SNYDER: No.

DR. RETTIE: Okay.

DR. SNYDER: Again, this is a little bit like the prostaglandins, it's a very complex ingredient. But I think that's why it's important that we get the experts, like we had today, to provide us some basis. So, I think that's a good plan.

DR. KLAASSEN: You know, in regard to the structure activity relationship of these chemicals, it is known that some of these phthalates have quite different toxicity than others. And, is it smart coming in to putting them all in one category? Or, I guess I'm thinking maybe after we hear these experts, we make that decision at that time, if we should do it as one category or three.

I mean, if you know already -- let's just say, as an example, in a group of compounds -- that one is really known to do this one thing, and the other six compounds in the group is absolutely known not to do that one thing, would you put them in a group or not?

DR. BELSITO: No. No, but I think you're spot on, Curt. That's what we ask these experts. Can these phthalates be grouped in terms of potential endocrine disruption?

DR. KLAASSEN: Okay.

DR. RETTIE: And isn't it also known that the monoesters are at least as toxic as the diesters, maybe more so? So, it could be a metabolic component there, a bit of SAR. I don't know that field.

DR. SNYDER: We can get clarification of all those questions.

DR. RETTIE: Yeah.

DR. HELDRETH: Yeah, and if it turns out that it doesn't seem like they fit together, like Dr. Klaassen is proposing, then we can split it out at that point.

DR. SNYDER: Perfect. All right. Good discussion.

Cohen Team – June 9, 2025

DR. DAVID COHEN: All right. Dibutyl Phthalate was placed on the 2024 Priorities list following nomination by the FDA for cause due to restrictions imposed on uses of plasticizers in food-contact applications.

We first published a Final Report on the safety assessment of Dibutyl Phthalate, Dimethyl Phthalate and Diethyl Phthalate in 1985, and concluded that these ingredients are safe for topical application in the present practices of use and concentrations in cosmetics. On re-review in 2002, the Panel reaffirmed the original conclusion as published in 2005.

In December of 2012, the Panel deliberated on studies separately concerning endocrine disruption and diabetes for Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate and Butyl Benzyl Phthalate. However, the Panel chose not to reopen the safety assessment of these ingredients and published their decision as a re-review summary in 2017.

In 2024, the CIR staff initiated a literature review. There's a lot of data out there. And the questions to us were, "Does the Panel or any other stakeholder have a particular expert in these areas they would like to invite to give a presentation on these DART and endocrine studies?" Anybody know anyone?

DR. EISENMANN: I have a comment. George Daston of Procter & Gamble would be happy to come and talk to you.

DR. DAVID COHEN: Who is it, again?

DR. EISENMANN: George Daston. He's talked to you before. He would be available for the December meeting, if that works out.

DR. DAVID COHEN: Okay. And the CIR noted that in 2024, RLD and the 2025 concentration of use by the Council indicate no current uses for Dimethyl Phthalate. The 2023 VCRP also reported no uses. Twelve uses were reported to the VCRP, and the maximum concentration of use reported was 2 percent in hairspray in 2001.

So, does the Panel support the idea of having Dimethyl Phthalate in a separate re-reviewed proposed document? Or would the Panel prefer to have this ingredient in the safety assessment with Dibutyl and Diethyl Phthalate?

So, if we keep it together, we could come to split conclusions. There's no problem with that. I just thought, if we have them as a separate report, we're kind of duplicating our efforts by going into this. Was there another thought behind this? Or you're just using a use not supported?

MS. FIUME: Use not supported, yes. It wouldn't be duplicating efforts; it would be use not supported.

DR. DAVID COHEN: Okay.

MS. FIUME: So it would be a re-review summary.

DR. DAVID COHEN: So you just spit it out and it goes to use not supported?

MS. FIUME: And it would be a re-review summary, right, Christina? Is that what we discussed when we discussed that? So it would probably be not reopened because it would be use not supported, like some of our re-review summaries today.

DR. BERGFELD: How long does it have to be not used? Is there a time period on it, or just when you do the survey they say, "not used," and you're done?

MS. FIUME: I don't know. I don't think there has been a time period on it so far. Because then if they --

DR. DAVID COHEN: You want to take a break? You want to take a two-minute break?

DR. ROSS: Sure, go ahead.

DR. DAVID COHEN: We could --

MS. FIUME: Yeah. I was going to say because if it goes use not supported, and then someone wants to use it, they can petition the Panel saying that they are using it and want it looked at.

DR. DAVID COHEN: So, look, from that perspective, I spoke too soon, right? So, let's split it so the actual review of the other two is limited. We agree? You guys alright with that?

DR. BERGFELD: I agree.

DR. TILTON: I agree.

DR. DAVID COHEN: So, the Phthalate discussion tomorrow is basically just answering these questions, correct?

MS. FIUME: Yes. It's to provide some guidance for Christina because there's so much information out there.

DR. DAVID COHEN: Yeah, it's going to be quite a report.

DR. DAVID COHEN: When do you think that hits the docket?

MS. BURNETT: Well, we're hoping to have the speaker in the same time that the Panel looks at it the first time. And if we go with PCPC's recommendation, that earliest would be December.

Full Panel – June 10, 2025

DR. DAVID COHEN: Right. Dibutyl Phthalate was placed on the 2024 priorities list following nomination by the FDA for cause, due to restrictions imposed on the use of plasticizers and food contact applications. The Panel first published on Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate in 1985 with the conclusion of safe.

In 2002, it was reviewed and the Panel reaffirmed the original conclusion. In 2012, the Panel deliberated studies concerning endocrine disruption and diabetes for Dibutyl Phthalate, Diethyl, Dimethyl, Butyl Benzyl Phthalate. However, the Panel chose not to reopen and published their discussion as a re-review summary in 2017.

This is going to come up, many of the studies have expanded. Mechanisms of action since 2024, a new literature search was sent. So we have mechanisms of action, gene expression and other specialized areas. And it may be beneficial for us to have a talk on DART and endocrine studies.

Does the Panel or any stakeholders have any opinions about who to invite regarding DART and endocrine studies? One suggestion from the PCPC was George Daston from P&G, who would be available. There were two other people that I found

that might be able to. One was Linda Birnbaum who is the Director of the NIEHS. Another was Dr. Swan from Mount Sinai. But I'm not sure that she may provide the information we need in this particular setting.

DR. BELSITO: Yeah, our team had discussed this. Yeah, I think George is fine but he is industry. It would be nice to get a non-industry person, get two opinions.

DR. SNYDER: We also liked George because he's presented to the Panel before, so he's familiar with our issues. So I think George and one additional person, that's what we kind of concluded.

DR. BELSITO: I had recommended a guy from Netherlands by the name of Aldert Piersma, who is actually the DART endocrine person on the RIFM panel. And if you look him up, he's done a lot of work on phthalates.

DR. ROSS: He sounds great, Don.

DR. SAM COHEN: Yeah, I think that one would be good. I would not have Dr. Birnbaum, because she has not done any work on this. And was the director of NIEHS, but was dismissed from that. She really has no experience with phthalates other than hand waving.

DR. DAVID COHEN: Okay.

DR. SNYDER: So I think we would suggest you go forward with George and Aldert.

DR. BERGFELD: Are you okay with that, Bart?

DR. KLAASSEN: I would like to also suggest two other excellent people who have worked on this area. One is Paul Foster and the other is Earl Grey.

DR. BELSITO: Are they academics, Kurt?

DR. KLAASSEN: They are government. They were at EPA.

DR. BELSITO: Okay.

DR. KLAASSEN: Or NIEHS, I can't remember for sure which one.

DR. ROSS: Yeah, Paul has done a lot with NTP as well, yeah.

DR. BERGFELD: Do you need anything else?

DR. HELDRETH: I can reach out. I've heard already that George is available for December. I'll reach out to the others and see what their availability is and then we can go from there.

DR. BERGFELD: Would you have them virtually or live or both? Whatever?

DR. BELSITO: We're virtual in December.

DR. HELDRETH: We're virtual in December. And that'll make things easy if we end up bringing in Dr. Piersma since he's in the Netherlands.

DR. BELSITO: Bart, your mic.

DR. HELDRETH: Yeah, being virtual will also make things easy for Dr. Piersma if we're going to have him do a presentation from the Netherlands.

DR. BELSITO: Yeah.

DR. BERGFELD: Okay. Anything else to discuss on the Phthalates, other than --

DR. BELSITO: Yeah. There was a second question about splitting Dimethyl Phthalate out into a separate re-review, and we didn't really think it was necessary, unless someone has a reason.

DR. DAVID COHEN: Our original motion was to split it and then do use not supported. However, in some late, breaking information from FDA, the RLD has three reported uses. So, we're not going to split them.

DR. SNYDER: We agree with that.

DR. BERGFELD: Okay. Any other questions regarding this strategy? Seeing none, then we're going to move on.

JULY 25-26, 1983 PANEL MEETING – FIRST PUBLIC REVIEW

Full Panel

The following conclusion of the report was unanimously approved:

“On the basis of the available data, the Panel concludes that Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate are safe for topical application in the present practices of use and concentration in cosmetics.”

Dr. Hoffman suggested mentioning in the Introduction that this review dose not include the carcinogenic ingredient Di(2-ethylhexyl) Phthalate (DEHP).

Dr. Bergfeld requested the clinical data be expanded in the text of the report, and Dr. Hoffman suggested reference be made to the Russian article L.E. Milkov in the clinical section of the report.

Subject to minor revisions, the document will be announced as a Tentative Report for a 90-day comment period.

JUNE 18-19, 2002 MEETING – FIRST RE-REVIEW

Full Panel

Dr. Belsito said that his Team determined that the available data (summarized in the Draft Report) that have been identified in the published literature since the Panel’s Final Report on Dibutyl Phthalate was published in 1985 warrant a decision by the Panel to reopen its original safety assessment.

Dr. Belsito said that the concern at this point relates to the anti-androgen effect of Dibutyl Phthalate, and that the Panel needs to do a thorough risk analysis to determine the extent to which its use in cosmetic products (primarily in nail products) contributes to the potential body load of Dibutyl Phthalate. He added that the Panel needs to examine how cosmetics containing this ingredient are used, what the likely absorption would be, and then do a risk assessment on that.

The Panel voted unanimously in favor of reopening the Final Report on the safety of Dibutyl Phthalate in cosmetics.

Regarding the Panel’s deliberations on the Phthalates in Teams, Dr. Andersen recalled discussions of additional data that are under development. He said that it is his expectation that, as the studies are completed, the data will be provided to CIR and incorporated into the next draft (Scientific Literature Review) of the CIR report on Dibutyl, Diethyl, and Dimethyl Phthalates.

NOVEMBER 18-19, 2002 MEETING – DRAFT AMENDED REPORT

Full Panel

At the June 18-19, 2002 Panel meeting, the Panel voted unanimously in favor of reopening the Final Report on the safety of Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate in cosmetics. This decision was made after reviewing data that have entered the published literature since the Final Report was published in 1985 with the following conclusion: On the basis of the available data, the Panel concludes that Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate are safe for topical application in the present practices of use and concentration in cosmetics.

Dr. Marks said that after discussing the anti-androgen effect of Dibutyl Phthalate in great detail and doing a risk analysis (particularly, a margin of safety evaluation), his Team determined that the Panel’s original conclusion on Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate should not be changed and, thus, that the Panel’s original safety assessment should not be reopened. However, it was agreed that the conclusion should be edited as follows to reflect the current wording of CIR’s conclusions: Based on the available information included in this report, the CIR Expert Panel concludes that Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate are safe for use in cosmetic products in the present practices of use and concentration.

In light of his Team’s decision, Dr. Marks added that there should be a lengthy discussion of all of the new data that have been presented and a statement indicating the basis for reaffirmation of the Panel’s original conclusion, taking into consideration the no-effect of Dibutyl Phthalate, in particular, the exposure from cosmetics, and the margin of safety that has been devised. Dr. Marks said that his Team determined that there is a significant margin of safety in relation to reaffirming the Panel’s original conclusion.

Dr. Bergfeld gave each Panel member an opportunity to comment.

Dr. Slaga said that the data used for the calculations adds a tremendous margin of safety, without taking into consideration that Dimethyl, Diethyl, and Dibutyl Phthalates have different potencies (i.e., ranging from essentially no activity to active [Dibutyl Phthalate]). Therefore, he added that the total cosmetic exposure is diluted by the potency of these Phthalates, Dibutyl Phthalate being the one for which, chemically (because it is used in nail products), its reaction does not make it available as

much as the calculations would indicate. Dr. Slaga also said that greater exposure is associated with Diethyl Phthalate, but, that its activity is weaker, thereby causing another dilution and resulting in a greater degree of safety.

Dr. Katz said that it is common knowledge at FDA that cosmetics represent one area in which Phthalates are used, and that the issue of Phthalates and exposure, as expressed during an earlier discussion, will be addressed on a broader perspective. Relative to FDA's review of some of the data, she added that FDA has also been unable to find the burden of proof or show that there is now a safety issue regarding the use of Phthalates in cosmetics.

Dr. Katz stated that FDA will review the CIR report on Phthalates in more detail, in an effort to identify any additional information that will be useful in further discussions within the agency.

Rachel Weintraub commented on data provided by Dr. McEwen at yesterday's Team meetings, specifically, the potential levels of Phthalates in humans. She said that there appear to be inherent limitations to this analysis, considering that all of the products that contain Phthalates are not known. Ms. Weintraub added that this is supported by a report provided by Health Care Without Harm, which discloses many other products containing Phthalates that the Panel had not been aware of.

Ms. Weintraub noted that Dr. McEwen's analysis includes a number of products (i.e., hair spray, deodorant, perfume, and nail polish), but does not include others which may contain Phthalates in the fragrances or flavors that are not indicated on the label (e.g., shampoos and body lotions).

Regarding yesterday's Team discussion (Dr. Mark's Team), Ms. Weintraub recalled comments on the additive effects of DBP and DEHP (source: poster by Foster et al.). She stressed the importance of including all of the data discussing the additive effects, rather than use of the poster by Foster et al. only.

Dr. McEwen said that the memorandum to CIR that he completed this morning addresses the aggregate exposure, as presented in a document from Health Care Without Harm, taking the most conservative NOEL and the 95th percentile maximally exposed subset of the population. This includes all DBP exposure, because it is a biomeasurement of that population. Dr. McEwen stated that the memorandum to CIR will be available to the public.

Dr. McEwen expressed industry's appreciation of information on Phthalates from the public. He said that the documents provided (particularly, the study on exposures from various products) helped him arrive at a decision as to the probable human exposure to Phthalates.

Dr. Bergfeld asked if anyone in the audience wished to make a brief statement.

Charlotte Brody, with Health Care Without Harm, expressed disappointment over the fact that her small, non-profit organization, in collaboration with two other small groups, had to present the cosmetics industry with data on Phthalate levels in its own products. She said that there should be more transparency in terms of which products contain Phthalates, and that this information needs to be made available.

Ms. Brody added that it is difficult for her to respond to Dr. McEwen's use of Health Care Without Harm's study, considering that his information was presented this morning. She said that this is not her idea of a public process. Similarly, she expressed disappointment over the fact that, in her opinion, calculations (based on industry's data on exposure) done in Teams yesterday will not be made available to everyone.

Ms. Brody asked the Panel to consider the question of whether or not the population would be better off with less or more Phthalate exposure. She said that the message to industry should be that less Phthalates in cosmetics would be better than more.

The Panel unanimously concluded that the Final Report on Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate should not be reopened and that the Panel's original conclusion should not be changed.

In summary, the Panel compared the current uses and consumer exposures with the available safety test data, and concluded that not only are exposures low compared to levels shown to produce adverse effects in animals, but that there was a high margin of safety between exposures and levels demonstrated to produce no-observable-effects in animal tests. Therefore, the Panel concluded that these ingredients are safe for use in cosmetic formulations in the current practices and concentrations of use, and that there was no need to reopen the safety assessment. This conclusion, an extensive presentation of the new scientific studies and other data considered by the Panel, and the rationale for the decision will be included in CIR's Annual Review, which presents the rationale for decisions not to reopen prior safety assessments.

Dr. Bergfeld asked that the Panel have an opportunity to review the Annual Review prior to its announcement.

Dr. Andersen said that the Annual Review will be made available to the Panel prior to its announcement.

FEBRUARY 6-7, 2003 MEETING – FIRST RE-REVIEW SUMMARY

Full Panel

A CIR Final Report with the following conclusion was published in 1985: On the basis of the available data, the Panel concludes that Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate are safe for topical application in the present practices of use and concentration in cosmetics. Since this conclusion was issued, many additional studies have appeared in the scientific literature. These studies, along with current frequency of use and use concentration data, were considered by the CIR Expert Panel.

On November 19, 2002, the CIR Expert Panel announced its decision to not reopen the Final Safety Assessment on the Phthalates and asked that a summary of the newly available data and a discussion of the issues be prepared for the Panel's review.

Dr. Bergfeld stated that the Panel has been provided with the re-review summary and discussion on the Phthalates, to be published in the Annual Review, for review at this Panel meeting.

Dr. Andersen noted that this request was made because of the understanding that documentation of the decision not to reopen (i.e., the summary for inclusion in the Annual Review) the Final Report on Phthalates would be different in comparison with other decisions not to reopen that have been made. He said that this is based on the need to present and discuss a large amount of new information on the Phthalates.

Dr. Andersen also stated that the Panel now has an opportunity to comment on the summary and discussion that have been prepared.

Dr. Snyder said that the Panel's reason for not reopening the Final Report should be stated in the first paragraph of the summary.

In response to Dr. Snyder's comment, Dr. Andersen proposed the following statement: Based on its consideration of the data discussed below, the Panel decided not to reopen this safety assessment.

DECEMBER 10, 2012 MEETING – SECOND RE-REVIEW/STRATEGY MEMO

Dr. Belsito's Team

DR. BELSITO: Right. Exactly. Thank you. Okay. We have time for a few more here. So let's at least start the phthalate discussion. And I guess since Alan is here and he had a chance to review this guidance for industry document and it was just sitting here when I got here and I haven't even seen it yet, basically this came from the FDA Center for Drug Evaluation and Research.

DR. ANDERSEN: That's correct.

DR. BELSITO: So, not from cosmetics.

DR. ANDERSEN: Right.

DR. BELSITO: And their recommendation is in drugs to do what?

DR. ANDERSEN: Phthalates in particular dibutyl phthalate and diethylhexyl phthalate have uses as excipients in drugs, and because there are alternatives to perform those excipient functions at the Center for Drug Evaluation and Research is recommending that the two phthalates not be used. And the science behind it is nothing different than what you have reviewed. There's really no new data. What's different is that the Center for Drugs Evaluation and Research has applied the precautionary principle and not a risk assessment. What you guys have done is a risk assessment with large margins of safety vis a vis cosmetics. So, I would argue that this doesn't have any real impact on the phthalate question, but you needed to know it exists.

DR. BELSITO: Okay.

DR. BRESLAWEK: Dr. Belsito? I might point out that the FDA guidance specifically states that the recommendations in this guidance do not address the use of DBP or DEHB in other types of FDA regulated products.

DR. BELSITO: Well, diethylhexyl phthalate is not a cosmetic ingredient anyway, right? I mean, because we deal with what, dibutyl, diethyl, and --

DR. ANDERSEN: Butyl benzyl.

DR. BELSITO: Butyl benzyl.

DR. ANDERSEN: Yeah, that's correct but I think that sentence in the guidance document was written more for the Center for Devices and Radiological Health, which does have diethylhexyl phthalate as a cross to bear because of its use in tubing to keep it flexible.

DR. BELSITO: Okay. So anyway, what we're presented here were three different studies. One done on a South Bronx population of children looking at phthalates in airway inflammation as measured by nitric oxide, and then two studies looking at diabetic populations. One from Uppsala, Sweden and the other based upon the NHANES data in the United States.

I thought the phthalates and airway inflammation -- I mean, they were able to draw some lines but when I looked at the scattergrams it looked like it was all over the place and I had a real hard time making any sense of it.

The diabetic studies -- I'm not a statistician and I was just sort of overwhelmed with the statistical analysis of these studies, but also impressed that there did seem to be somewhat of a correlation with urinary phthalates and diabetes. And then a possible explanation for this, you know, based upon the nuclear praxisome proliferating activity and the fact that there are drugs that target that to treat diabetes.

So, I don't know if that's cause for re-opening because I'm not sure that phthalates from cosmetic preparations are absorbed to a level where that will reasonably occur. Also, they were looking at phthalates that aren't used in cosmetics, so I really thought that when I was looking at those levels of non-cosmetic phthalates that there were other sources that were likely more important if this was even real, which I'm not sure it is. But there were sources of phthalate exposure such as freezing plastic water bottles or microwaving in plastic whatever that probably were more important than what we were seeing in terms of exposure from cosmetics.

But that was my own personal view, so I open it up to people who know more about phthalates and diabetes and airway inflammation than I do.

DR. SNYDER: Who would that be? (Laughter) Yeah, I read the data on all these. I thought they both did a good bit of addressing the study limitations of both studies. I still think that the big missing link is what's the underlying mechanism. You know, we have these associations but how does that link that they're related, I think is one issue.

I did want to know that we did not have -- I didn't understand why we did not have the Gaithera, 2004 reference was not in our document, in our report previously. But I thought that we had addressed the issues related to -- in the old report we did address the metabolism issues and the enzyme systems as it pertained to the diabetes report. So, I thought we had addressed some of that even though it wasn't directly linked maybe at the time to diabetes, but we were already aware of those issues about the enzyme systems regarding the metabolism of the phthalates.

So, I wasn't all that concerned about the new data set in regards to what we already know about the phthalates.

DR. LIEBLER: I think the papers establish that there is a relationship between the parameters measured and nothing more. I -- you know, for example between exhaled nitric oxide and urinary phthalate metabolites. There is a statistically significant relationship that may or may not be biologically significant, and so I felt that way about all three of these papers. Essentially there are three variations on the same type of study.

And so I said, okay, well if we did use this to re-open what the heck would we do once we did that? Because we can't interpret these studies at any level mechanistically that would inform our evaluation of our prior conclusion. So for that reason I felt that we really can't use these studies to really re-evaluate our conclusion because they provide really no mechanistic insight as to whether or not there's any causal relationship between these ubiquitous environmental contaminants and the disease states -- either airway disease or diabetes.

So I felt if we did re-open it we'd have no place to go and we wouldn't really end up being able to change our conclusion. So for that reason, I suggest we do not re-open.

DR. KLASSEN: I'll basically second what Dan said. You know, these are -- these three studies kind of show a weak association. You know, the data is not that impressive. You know, they are statistical associations. I mean, we know that the -- what the phthalates do biologically, and that's been covered before.

I question that it is a biological significance, or even if it might be reproduce-able in another study. So, there are many explanations for why these associations might occur and to suspect that they really are important is premature at this stage. And I think we should not re-open, because of these papers or anything else.

We also have another paper, I guess, on our desk this morning in regard to the sulfation with the phthalates that we might want to address.

DR. BELSITO: I didn't see that.

DR. KLASSEN: It had to do with sulfation.

DR. BELSITO: Oh, that was part of the child/infant report.

DR. KLASSEN: Yeah, might use it two different ways.

DR. BELSITO: Oh, okay. So we acknowledge the papers, we've read them and don't want to do anything with them. So, Alan, how do we communicate to the public that we did this?

I mean, were we being asked to potentially re-open this on the basis of this data? I thought this was like a panel FYI.

DR. ANDERSEN: Well, it's a panel FYI but you do need to make the decision if the information crosses the threshold to re-open.

DR. BELSITO: Okay.

DR. ANDERSEN: By saying that it doesn't, we'll capture that in the minutes of the meeting and it will become a matter of public record.

DR. BELSITO: Okay.

DR. BERGFELD: But also, it will be in our record and the annual report, will it not? That which we have not re-opened and the reasons given.

SPEAKER: Yes.

DR. BELSITO: I guess the only thing that I have a slight discrepancy with is, you know, when we say, you know, we have no mechanistic clue. I would agree with the airway inflammation but, you know, I thought the argument for the peroxisome proliferator receptors in diabetes was pretty cogent. So, I think they offered a potential mechanism there, so I think we have to come up more with than just a statement we don't think there's any mechanism that helps us.

You know, my point was that they were looking at levels of phthalates that aren't used in cosmetics and in fact the -- as I interpreted the data, the burden from cosmetic use was insignificant compared to burdens from other exposures, particularly given the levels of non-cosmetic phthalates that were found in the urine. So, that was my point. Not that, you know, these were interesting articles. You know, there was an association, there was a potentially plausible explanation for a link. The issue was, we don't think the exposure is largely from cosmetics, we think the cosmetic exposure is negligible and that -- number one, and number two there needs to be further investigations. You know, there were limitations in the study, as the authors readily acknowledge. These studies had limitations, yadda, yadda, yadda.

But to simply say there was no mechanistic explanation for airways, I would agree. But for diabetes, I would have a little bit of a pause.

DR. SNYDER: I agree. I think they did address, and very well -- they said this could potentially be a plausible mechanism, but they even conclude themselves that their data set is insufficient at this time to conclude that that is a mechanism action that further studies are required.

I think it's the same conclusion that we would come up with if they hadn't written that, or that's what we're saying, is that we are aware of these associations but at this time there's no data to support they're nothing more than associations.

DR. BELSITO: I'm sure you can wordsmith it.

DR. ANDERSEN: Message received.

Dr. Marks' Team

DR. MARKS: Okay. Talking about phthalates. So, this morning we found this. The guidance for industry limiting the use of certain phthalates as expediting CDER regulated products. In 2005, the panel decided not to reopen the phthalates particularly at that time we were focused on the possible endocrine disrupter development issues.

We have some new studies. One, an airway study and a couple on diabetes. Alan's summary states the issues well. Is there any reason to reopen?

DR. SLAGA: Do not reopen.

DR. SHANK: Well, those epidemiological studies are not a cause to reopen. I'm not too sure how to handle this. I haven't read it yet. The CDER regulation but if this applies to drugs is it also going to apply to cosmetics eventually.

DR. ANDERSEN: Very specifically not.

DR. SHANK: Okay.

DR. ANDERSEN: Not the medical devices, not the cosmetics, it's dibutylphthalate, diethylhexophthalate as used as excipients in drugs only.

DR. SHANK: Okay. Thank you.

DR. ANDERSEN: I just didn't want to hide it since it came out last Friday. It seemed timely.

DR. SHANK: Yes. Okay. Then I would say don't reopen.

DR. MARKS: Ron Hill, not reopen? Yeah, okay. Now, in terms of the discussion and the re-review.

DR. SLAGA: Well, we definitely have to discuss the epi studies but as Ron pointed out the early association, there's no way to come up with a concentration to relate to cosmetic. There is concern in my eyes that some of these phthalates may have effect on people who are gamma receptors. Which the gamma ones are the ones that the diabetes drugs, the glitazone class of compounds are effective agonists. And then also the alpha, there's a number of different types of fibrates that interrelate to this too that I just think we just discuss that and that's all we have to do.

DR. ANSELL: Well, this is not a re-review. This was three specific papers which questioned whether this should jump out of cycle.

DR. SLAGA: Yeah.

DR. ANSELL: So, I don't know that, I mean that's of course an issue and we can reopen based on --

DR. SLAGA: No, no. I didn't say reopen. This is in a re-review summary.

DR. ANSELL: But this isn't a re-review, right?

DR. MARKS: It's a thick document for not being a re-review.

DR. ANSELL: I thought this was -- I'm sorry. Maybe I'm out -- I thought this was brought forward specifically to assess three papers that we became aware of.

DR. ANDERSEN: That's correct. It's -- you have the option of reopening based on these new data but the minutes of the meeting could be an adequate summary of the basis --

DR. SLAGA: Right.

DR. ANDERSEN: -- for a decision to not reopen it. Like you did at the last meeting with respect to parabens. You said, sorry Charlie, there's not enough new information here to support reopening it. And our previous conclusion is still okay. And that's what I would -- if you choose not to reopen, that's what I'd do here.

DR. MARKS: Leave the minutes and let it stand and not actually publish a re-review summary.

DR. ANDERSEN: Yeah, I think if we did otherwise we'd be publishing a re-review every time somebody published something.

DR. MARKS: Right.

DR. ANDERSEN: No, thank you.

DR. ANSELL: It would make the threshold to look at a paper that came through so onerous that we might not want to do that.

DR. ANDERSEN: Right. Giving you the option of saying, fiddlesticks, this is important. We'd better reopen this. That is what --

DR. MARKS: And so, in the minutes we'll capture that you've already said that, Alan, and Ron Shank is --

DR. ANDERSEN: Yeah and we'd make the note that it specifically excludes any relevance to cosmetics.

DR. MARKS: Okay. So, we'll not reopen phthalates once again and we'll just capture that in the minutes and the biggest three epidemiologic studies don't warrant reopening. Okay.

Full Panel

DR. BELSITO: This is part of our ongoing surveillance of chemicals that have caught the public's attention in various ways. And this was triggered by two reports, one suggesting a linkage between urinary phthalate monofunctional metabolites and airway inflammation as measured by nitric oxide in children from the South Bronx. It was a rather interesting paper with statistical correlation, but when you looked at the individual points, they were really all over the board. And some of the higher phthalate levels were phthalates that are not used in cosmetics, suggesting that exposures were from sources other than cosmetic exposures. And the mechanism of action was speculated, and the authors actually went through great pains to point out the various limitations to their study and what would need to be done.

Then there were two studies on diabetics, one from Sweden and one part that was taken from the U.S. in the Haney study. And again they linked phthalate levels to diabetes, and they specifically subclassified the types of diabetes depending upon phthalate exposure, hypothesizing in effect on nuclear peroxisome proliferating activity receptors because apparently there are diabetic drugs that act via that mechanism. But again --

DR. SLAGA: Antidiabetic.

DR. BELSITO: Antidiabetic, yes, thank you for the correction. But again, despite the statistical linkages, again many of the phthalates were those that aren't used in cosmetic products. And while we took these under advice for lack of a better word, we didn't really feel a need to reopen the phthalate document because of them.

DR. MARKS: Second.

DR. BERGFELD: So a motion has been made and seconded not to reopen. Any other discussion?

DR. MARKS: Our team wanted to know how we would capture this discussion since phthalates are a hot topic and we felt that the minutes would be adequate.

DR. BERGFELD: Alan, can you respond to that? How this will be recorded? It seems to me we have a mechanism in the annual report that appears in The Journal.

DR. ANDERSEN: I think we have traditionally summarized all decisions to not reopen safety assessments and while we haven't published one recently, they are prepared for inclusion in The Journal. I don't know what our track record is at this point in terms of self-initiated re-reviews. That's what I went through was certainly true for the -- as Dr. Belsito referred to a minute ago -- the 15-year mandated re-reviews. Those for sure are done. It's really the Panel's call as to whether this decision is just captured in the meeting minutes or whether a full re-review summary is prepared for publication. In the most recent example, when the Panel reviewed parabens earlier this year, you did not ask for a re-review summary. You just determined to not reopen it based on the handful of new studies that were available.

Procedurally my concern is a new CIR re-review publication every time a paper comes out; that's procedurally a question. But if that's the Panel's desire, we can go in that direction.

DR. BERGFELD: Jim?

DR. MARKS: As I said, our team felt it could be captured in the minutes. Just precisely for what you said, Alan, is that every time a new study came out if we did a re-review summary, it would just become quite burdensome and probably not add much to the safety of these ingredients.

DR. BERGFELD: I'd like to ask a question. Are the minutes available to the public because this, as Don mentioned, is a hot item now and our response would be important?

DR. ANDERSEN: The answer is yes, they are. The post-meeting announcement that will include all of the details of this discussion will be -- if we follow our current practices and procedures -- be online and available by this Friday. So everybody will have -- any interested party will have ample opportunity to see what we did.

DR. BERGFELD: Don?

DR. BELSITO: I guess the only other comment I would make, and you can take it or leave it, would be that, for instance, if you go in to check CIR status and you were to type phthalate, it would come up with our original report and our re-review. It may also be nice when we've taken a special look, even though we've determined not to reopen, not to issue a report, if you typed in phthalates, when it came up it said "discussed at Panel meeting" -- or something to that effect -- "see minutes." And you can hit on it and it links to the minutes. So someone could look and see not only did we re-review phthalates in whatever year, but in December of 2012 we took a look at these three specific papers and decided that for the reasons reflected in the minutes not to go back and reopen the report.

DR. MARKS: That was a question I was going to ask Kevin. Is this searchable? Will it appear?

DR. ANDERSEN: Speaking on Kevin's behalf, I think the answer is it's searchable if we make the extra effort to make it searchable and if there's nothing automatic about it. But I think given Don's comments, message received.

DR. BERGFELD: I'd like to make a comment. Why is it that we cannot do a document, but just not present it for publication but to present it to the linkage in addition to the minutes? I mean it's not -- it's a one-pager usually that updates with the appropriate references, and you can prepare it for the Website, but not particularly for publication..

DR. MARKS: I guess the minutes really capture it the way Don reviewed it, and if we have three more studies that appear in the next six months and we review those studies again, do we do another page document? Pretty soon it'll become a boilerplate as to these three new studies, these two new studies, these five new studies didn't cause us to reopen. So I think it's captured in the minutes as long as the minutes are searchable as Don suggested. I think that's really important so it'll direct you to the meetings of this discussion right today.

DR. HILL: As long as we know that that key word would result in hitting the post-meeting summary. I mean Kevin will know what's needed there to make that happen. To me that would be sufficient except under a circumstance where we really had the need to put together some sort of a paper. There might be conditions, and I'm not sure this would be one of them, but that's just my opinion.

DR. SNYDER: I have a more general comment. I'm less concerned about what we do with the data, just that we're aware of the data. So I'm curious as to what triggered us to be aware of those three publications. And so do we have or is it so to speak on the radar of a particular writer, parabens, the phthalates, the hot button topics, because I think we need to be kept abreast of the current publications. Now whether they rise to the level that we want to consider a reopening or not, I think we should be not waiting 15 years before we ever look at phthalates or parabens again with it being such a public awareness of those ingredients. And I think we need to -- so I'm more concerned about what are we doing to trigger looking for new data on those ingredients.

DR. BERGFELD: Alan?

DR. ANDERSEN: I wish I could give you a presentation of what triggers that happening. It's in the eye of the beholder with respect to the paraben studies that we looked at earlier this year. It just seemed very obvious that we needed to look at that. So it's really between me and the rest of CIR staff flagging that something is important to bring forward. I don't have anything more than we know it when we see it.

DR. BELSITO: And I think that we need to be vigilant in our own various worlds. I mean I get -- a lot of the phthalate issues get directed to Dan and me because we now sit on the RIFM Panel and phthalates are obviously very important to the fragrance industry. And I think -- and then as a dermatologist I have various links that come up to issues related to concerns of chemicals that are in derm products, and paraben is the largest or is the most frequently used preservative in cosmetics. So I think we all need to take our own level of expertise. I'm sure we're all on various Web links that send us information important to what we do. And when we see something coming across about a cosmetic ingredient, we head it back to Alan.

DR. BERGFELD: Halyna, can you respond to how the Council might identify these particular hot button items?

DR. BRESLAWEK: Well, as you can all imagine, we monitor the safety information and concerns about all cosmetic ingredients pretty vigorously. A lot of times it's something we pick up in the press and we decide needs to come to CIR. We've asked CIR for review and opinion. A lot of times our CIR Science and Support Committee will bring something to the table and also the SRTC, the top toxicologists from all the companies who meet on a regular basis, will bring something up. So we've made a practice of bringing articles and issues that raise concerns about ingredients that CIR has reviewed. We make a practice of bringing that to CIR for their review and assessment.

DR. BERGFELD: So basically you're the alert person for us officially?

DR. BRESLAWEK: From our perspective, absolutely, but I just historically know that FDA's done this and that individual members of the Panel have also raised concerns.

DR. BERGFELD: Don?

DR. BELSITO: And the other Website that I follow to find out the latest breaking news, whether it be Internet roar or a scientific publication, is the Environmental Working Group because you can ensure that it pops up very quickly on their site if there are any concerns about cosmetic products..

DR. BERGFELD: Dan?

DR. LIEBLER: One practical point about having, for example, phthalates, this instance of our discussion of phthalates to show up on the Website. But I don't think we necessarily need is for every time phthalates gets mentioned in a transcript incidentally, for example, to pop up in a search. But more along the lines of perhaps the threshold would be if it's on the agenda, there should be an entry and that will probably cover all of these instances then.

DR. ANDERSEN: Well I'm glad you said that because that increases the likelihood that Kevin won't shoot me.

DR. LIEBLER: Especially because I'm right between you and Kevin.

DR. ANDERSEN: But I should -- we've had a lot of discussion about who is watching what's going on in the scientific literature, and I think any interested party should feel empowered to let CIR know if something has appeared that deserves CIR's attention. There's just no reason for there to be any barrier to that happening. So yes, we're vigilant, and we have people who have a self-interest in being vigilant. We have the scientific expertise on the Panel that picks stuff up. All that's great, but anybody ought to be able to raise a red flag and have us pay attention to it.

DR. BERGFELD: We think that we need to vote on this. I'm just surveying if we voted. The conversation went on so long here. But I call for the motion. The motion I believe is not to reopen and to add -- and the discussion led to how we were going to record and link why we did not reopen for the public and our own interest. So if there's no more discussion, let me call for the vote not to reopen. Unanimous. Thank you.

MARCH 18-19, 2013 PANEL MEETING – SECOND RE-REVIEW SUMMARY

Full Panel

Dr. Bergfeld was very pleased with the summary and discussion, and noted that this report sets the precedent for creating this type of document for controversial ingredients that are considered for re-review in the future. She added that the Panel should routinely have an opportunity to review all re-review summaries and discussions before they are published in the Annual Review.

Dr. Andersen said that the re-review summaries could be routinely included on the meeting agenda, for the Panel's comments, prior to publication in the Annual Review. He noted that the public already has an opportunity to comment on these summaries, because the Panel's re-review decisions are announced to the public and a 90-day comment period is observed.

The re-review summary and discussion on Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were approved by the Panel, and Dr. Andersen noted that the announcement of this decision will be followed by a 90-day public comment period.

JUNE 12-13, 2023 – 2024 PRIORITY LIST

Belsito's Team

DR. BELSITO: Priorities. So basically, we've just been asked to prioritize -- that's in admin, right?

DR. SNYDER: Yeah.

DR. KLAASSEN: Yes.

DR. BELSITO: So, since our March meeting we received communication from the FDA nominating ingredients for cause, specifically Toluene and Dibutyl Phthalate. So, we're going to be doing accelerated re-reviews on those. And then there was something here that I just want Monice or someone to clarify. So, it basically said that instead of just doing a re-review summary, we're going to fully open this or something?

MS. FIUME: So, are you talking about Toluene?

DR. BELSITO: Yeah.

DR. SNYDER: We never reviewed it before.

MS. FIUME: Well, it is on our list of items to be re-reviewed. It's currently on Christina's docket. Right, you have Toluene?

DR. BELSITO: Right. We reviewed both of them before.

MS. BURNETT: I think so. I don't know.

DR. SNYDER: Oh, that's the TPO. I was talking about TPO. Yeah.

DR. BELSITO: Right.

DR. SNYDER: I'm sorry, TPO is what I was talking about.

MS. FIUME: Right. TPO is the only one. Dibutyl Phthalate was just re-reviewed in 2017.

DR. BELSITO: Right.

MS. FIUME: But Toluene was scheduled for consideration for re-review this year, so you will be seeing that soon.

DR. BELSITO: Right. But it says, "The CIR will present the panel with a draft amended report on this ingredient instead of an abbreviated re-review document."

MS. FIUME: Okay. So instead of getting the table that you have been --

DR. BELSITO: Right. We are actually going to get a written document?

MS. FIUME: Assuming that you were going to accept FDA's request to reopen it.

DR. BELSITO: I think if FDA comes to us with a request for cause, we have to -- I don't know -- yeah.

MS. FIUME: Which is why you'll get an actual report person versus do you want to reopen? Here's the table of data that we found and then -- just taking that step out.

DR. BELSITO: Right, okay. So, we're going to -- yes, we're reopening Dibutyl Phthalate and Toluene for cause. And I think the third ingredient -- I mean, this is the type of stuff that I want to see happening. Something's going on in Europe, there's a

concern about this material for reproductive toxicity, we need to be looking at it, number one. Number two, we've never even reviewed it. So, yes, I personally would like it added to the 2024 priority list.

DR. SNYDER: Agreed.

DR. KLAASSEN: It would be interesting to know why they wanted these first two chemicals. We don't -- why they want us to do Dibutyl Phthalate?

DR. BELSITO: Because it's a huge issue in endocrine disruption --

DR. KLAASSEN: Right, right.

DR. BELSITO: -- and --

DR. KLAASSEN: But I don't think there's any new data since the last time we did it, but maybe there is. And how about Toluene? I mean, I'm not against doing it, I'm just wondering. It'd be nice if they said why.

MS. FIUME: So, I'm looking at the memo and the email that was originally sent on March 20th, it's PDF Page 26. It just says that they're proposing it.

DR. BELSITO: Yeah. This is from Prashiela.

DR. KLAASSEN: Yeah. It says nothing really.

DR. BELSITO: Right.

MS. FIUME: Sorry, Priya has Toluene. So, Priya will be bringing that back probably in September.

MS. BURNETT: And Phthalates.

MS. FIUME: Yeah.

DR. BELSITO: I mean, both of them have gotten a lot of press, you know, bad press.

DR. KLAASSEN: Yeah, I know about the phthalates always do.

DR. BELSITO: Well, Toluene for carcinogenicity.

DR. RETTIE: So, the phthalates are the less (inaudible) issues, right?

DR. BELSITO: Right. I'm surprised that they are supposedly only one reported use because they used to be used in a lot of nail enamels. But I guess now everyone's using acrylic, so I don't know.

DR. KLAASSEN: Well, let's do them.

DR. SNYDER: Been there, done that.

DR. BELSITO: They're also used in a lot of fragranced products to hold the fragrance on the skin as a fixative, I think.

MS. KOWCZ: No.

DR. BELSITO: No?

MS. EISENMANN: Diethyl.

MS. KOWCZ: The Diethyl.

DR. BELSITO: Yeah, diethyl. Okay.

Cohen's Team

DR. COHEN: All right. Now we're going to Priorities. Okay, for the 2024, draft priorities, we asked for propolis to be accelerated. Two other ingredients were initially proposed and were removed from the list and it was determined that cannabidiol should be reviewed singly.

The others are listed here, some with pretty high frequencies of use reported. Any comments on this? I mean, I don't know if we're going to have a really in depth conversation about this, are we?

DR. HELDRETH: I think that the main point was that FDA had actually asked for three additions to our prioritization. Two of these are request for accelerated rereviews, so Toluene and the Dibutyl Phthalate.

Now Toluene was actually already in our in-house pipeline. We were already working on it, so that one's definitely coming back your way. Dibutyl Phthalate, we haven't started working on yet. But now that FDA has requested it, we've went ahead and added it, unless the Panel has an objection to accelerating that be reviewed.

So, the only real question, I think, for the panel is do they want to add this Trimethylbenzoyl Diphenylphosphine Oxide to the prioritization list for next year?

DR. SLAGA: I think we should accelerate it.

DR. COHEN: Yeah. That's a question to the Panel. We should add them.

DR. TILTON: Yeah, I agree.

DR. ROSS: New data. I agree.

DR. HELDRETH: Okay. That's easy.

DR. ROSS: Bart, could I ask you, what was the reason for -- or maybe you don't know -- why FDA nominated Toluene and the Dibutyl Phthalate? Was there a specific reason?

DR. HELDRETH: Prashiela stepped out?

DR. ANSELL: Our FDA person just --

DR. COHEN: We can ask her when she comes back.

DR. ROSS: Ah, okay.

DR. COHEN: These are plastics, the phthalates, right?

DR. HELDRETH: Plasticizer, yeah.

DR. ROSS: Yeah, they're phthalates. Toluene is a little different.

DR. COHEN: Yeah, Toluene is going to be a bit different.

DR. HELDRETH: Well, we've looked at the phthalates before.

DR. ROSS: Yeah.

DR. BERGFELD: And there's a lot of endocrine disruption with that group.

DR. COHEN: So, it's interesting. In 2017, the panel reaffirmed it, so this would be a real short cycle.

DR. HELDRETH: Right.

DR. COHEN: Prashiela, a question. No, no, no, it's okay. For the priority list, the FDA nominated some items, one was Toluene. Do you know why Toluene was nominated?

DR. MANGA: I'm going to have to get back to you on that one. Let me take a quick look at what we --

DR. COHEN: And the phthalates, the dibutyl phthalate?

DR. MANGA: I think there's just a lot of interest in phthalates right now. It's come up quite a bit. The Toluene is being used in a lot of nail products.

DR. ANSELL: Historically.

DR. MANGA: Historically.

DR. COHEN: Are you talking about the Toluene sulfonamide resins or just Toluene?

DR. ANSELL: No Toluene is a diluent.

DR. ROSS: I think Toluene is being reviewed quite a bit at IARC on its own, but also in connection with Benzene.

DR. ANSELL: Right. Also not used anymore, so.

DR. ROSS: Yes.

DR. ANSELL: But we fully support accelerating anything FDA ask us to.

DR. COHEN: We're good. Yeah. So are we.

DR. HELDRETH: Which is a question, I just wondered why they --

DR. BERGFELD: Actually, we really like it when they ask.

DR. ANSELL: Yes. More than support it, encourage it.

DR. MANGA: We appreciate that.

DR. COHEN: No, it's nice we're being paid attention to. And the other one was -- Annex 3 was a little more self-explanatory.

DR. BERGFELD: What was that?

DR. COHEN: The Trimethylbenzoyl Diphenylphosphine Oxide.

DR. ROSS: Yeah it's more data. Yeah.

DR. HELDRETH: Yeah, it looks like there may be some repro concerns with that one.

DR. COHEN: Some? I didn't hear what you said.

DR. HELDRETH: Repro -- DART issues with that ingredient.

DR. COHEN: Repro. Okay. All right, so I think we're aligned on the priorities.

DR. BERGFELD: I think when we present this, it would be nice if you, the FDA, presented the reasons for bringing them forth.

DR. COHEN: Just like a sentence.

DR. BERGFELD: It would be very nice.

Full Panel

DR. BELSITO: So the FDA has asked us to move Toluene and Phthalates up for cause. And I would agree with doing that. And also, it was brought to our attention that a material that we haven't reviewed, trimethylbenzoyl dimethyl phosphine oxide, is being looked at by the European Chemical Agency, ECHA. And they're very concerned about the safety of this. It's a substance of very high concern (SVHC), and I think we should move that up on our Priority List as well.

And I think this is the type of thing that needs to be done, where we're monitoring what other safety organizations are looking at, perhaps, flagging ingredients that we weren't aware of. And we should continue to do this type of thing.

DR. BERGFELD: Any comments, Dr. Cohen?

DR. COHEN: No, I thought we might have heard from the FDA a little more why they were nominated.

DR. BERGFELD: Jan, do you want to talk about the nominations?

DR. HELDRETH: We also have Dr. Manga online.

DR. BERGFELD: Manga too?

DR. HELDRETH: She had to return to the office.

DR. MANGA: Hi, this is Prashiela. So these three ingredients came up because we've had a couple of inquiries about these being used in nails -- I'm sorry, I'm getting a bit of feedback from the room.

DR. BERGFELD: We can hear you.

DR. MANGA: So these ingredients have been noted particularly for the use in nail products. And that was why we were interested. And then, as Don mentioned, at least for the TPO, that is coming up as a new ingredient. We were concerned that it be reviewed given the other reviews that are going on.

Toluene is now one of the California Department of Toxic Substances Control products that effective January 1, 2023, nail products containing Toluene will become priority products. And, so, we felt that this was also one that needed to be looked at once again.

In terms of Dibutyl Phthalate, this is one which was included when FDA amended the food-additive regulations, to no longer provide for 25 plasticizers in various foods contact applications. They did this because the uses were abandoned, but given that this one was included in these amendments, we felt that it would be timely for CIR to review it as well.

DR. BERGFELD: Thank you very much. We're really appreciative of the FDA coming in and suggesting these particular ingredients.

Safety Assessment of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate as Used in Cosmetics

Status: Draft Amended Report for Panel Review
Release Date: November 10, 2025
Panel Meeting Date: December 4-5, 2025

The Expert Panel for Cosmetic Ingredient Safety members are: Chair, Wilma F. Bergfeld, M.D., F.A.C.P.; Donald V. Belsito, M.D.; David E. Cohen, M.D.; Samuel M. Cohen, M.D., Ph.D.; Curtis D. Klaassen, Ph.D.; Allan E. Rettie, Ph.D.; David Ross, Ph.D.; Paul W. Snyder, D.V.M., Ph.D.; and Susan C. Tilton, Ph.D. The Cosmetic Ingredient Review (CIR) Executive Director is Bart Heldreth, Ph.D., and the Senior Director is Monice Fiume, M.B.A. This safety assessment was prepared by Christina Burnett, M.S., Senior Scientific Analyst/Writer, CIR, and Jinqiu Zhu, Ph.D., Toxicologist, CIR.

ABBREVIATIONS

| | |
|-------------------|--|
| AchE | acetylcholinesterase |
| AKT | protein kinase B |
| ALB | albumin |
| ALT | alanine aminotransferase |
| AOP | adverse outcome pathway |
| AST | aspartate aminotransferase |
| 8-OHdG | hydroxydeoxyguanosine |
| BAL | bronchoalveolar lavage |
| BaP | benzo[a]pyrene |
| CAT | catalase |
| Cdyn | dynamic compliance |
| CGRP | calcitonin gene-related peptide |
| CHO | Chinese hamster ovary |
| CI | confidence interval |
| CIR | Cosmetic Ingredient Review |
| Council | Personal Care Products Council |
| CRP | C-reactive protein |
| Cx43 | connexin 43 |
| DEF | S,S,S-tributylphosphorotrithioate |
| <i>Dictionary</i> | <i>International Cosmetic Ingredient Dictionary and Handbook</i> |
| DMSO | dimethyl sulfoxide |
| DNFB | 2,4-nitrofluorobenzene |
| ECHA | European Chemicals Agency |
| EITC | eosin 5-isothiocyanate |
| ELISA | enzyme-linked immunosorbent assay |
| EPA | Environmental Protection Agency |
| ER | estrogen receptor |
| FDA | Food and Drug Administration |
| Fe ²⁺ | ferrous iron |
| FITC | fluorescein isothiocyanate |
| FOR | fecundity odds ratio |
| FT3 | free thyroxine |
| FT4 | free thyroxine free thyroxine |
| GC-MS | gas chromatography – mass spectrometry |
| GDM | gestational diabetes mellitus |
| GLUT | glucose transporter |
| GPGR | G-protein-coupled estrogen receptor |
| GPMT | guinea pig maximization test |
| GPX4 | glutathione peroxidase-4 |
| GSH | glutathione |
| GSSG | oxidized glutathione |
| hER | human estrogen receptor |
| hEST | human embryonic stem cell test |
| HPLC | high-performance liquid chromatography |
| HRIPT | human repeated insult patch test |
| IARC | International Agency for Research on Cancer |
| IC ₅₀ | 50% inhibition of growth and viability |
| ID ₅₀ | 50% inhibition of differentiation |
| IgE | immunoglobulin E |
| IGF | insulin-like growth factor |
| IL | interleukin |
| JAK 1 | Janus kinase 1 |
| LABC | levator ani plus bulbocavernosus muscles |
| LC-MS/MS | liquid chromatography-tandem mass spectrometry |
| l.o. | leave-on |
| LOAEL | lowest-observed-adverse-effect level |
| LOEC | lowest-observed-effect concentration |
| MDA | malondialdehyde |
| MDI | methylenediphenyl diisocyanate |
| MNG | multinucleated gonocytes |

| | |
|----------------|---|
| MoCRA | Modernization of Cosmetics Regulation Act |
| MOE | margin of exposure |
| MOS | margin of safety |
| MRL | minimal risk level |
| MS | mass spectrometry |
| MTT | 3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide |
| NF- κ B | nuclear factor kappa B |
| NO | nitric oxide |
| NOAEL | no-observed-adverse-effect level |
| NOEC | no-observed-effect concentration |
| NR | not reported |
| Nrf2 | nuclear factor erythroid 2-related factor |
| NTP | National Toxicology Program |
| OECD | Organisation for Economic Co-operation and Development |
| OVA | ovalbumin |
| OXA | oxazolone |
| Panel | Expert Panel for Cosmetic Ingredient Safety |
| PEITC | phenethyl isothiocyanate |
| PET | polyethylene terephthalate |
| PI3K | phosphatidyl inositol 3 kinase |
| POD | point of departure |
| PRDX6 | peroxiredoxin-6 |
| q-PCR | quantitative polymerase chain reaction |
| Re | expiratory resistance |
| REACH | Registration, Evaluation, Authorization and Restriction of Chemicals |
| Res | resveratrol |
| RfD | reference dose |
| Ri | inspiratory resistance |
| RIFM | Research Institute for Fragrance Materials |
| RITC | rhodamine B isothiocyanate |
| RLD | Registration and Listing Data |
| r.o. | rinse-off |
| ROS | reactive oxygen species |
| RT-PCR | reverse transcription-polymerase chain reaction |
| SCCNFP | Scientific Committee on Cosmetic and Non-Food Products |
| SCCP | Scientific Committee on Consumer Products |
| SCE | sister chromatid exchange |
| SCO | Sertoli cell-only |
| SED | systemic exposure dose |
| SF-1 | steroidogenic factor-1 |
| SHBG | sex hormone binding globulin |
| SOD | superoxide dismutase |
| STAT6 | signal transducer and activator of transcription 6 |
| STZ | streptozotocin |
| TDI | toluene 2,4-diisocyanate |
| TG | test guideline |
| TMA | trimellitic anhydride |
| TNF- α | necrosis factor- α |
| TPN | total parenteral nutrition |
| TRPA1 | transient receptor potential ankyrin 1 |
| TSLP | thymic stromal lymphopoietin |
| TT3 | total triiodothyronine |
| TT4 | total thyroxine |
| TUNEL | terminal deoxynucleotide transferase-mediated deoxy-UTP nick labeling |
| VCRP | Voluntary Cosmetic Registration Program |

INTRODUCTION

This assessment reviews the safety of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate as used in cosmetic formulations. According to the web-based *International Cosmetic Ingredient Dictionary and Handbook (Dictionary)*, these ingredients are reported to function as fragrance ingredients, plasticizers, and solvents in cosmetic formulations (Table 1).¹ Diethyl Phthalate is also reported to function as a denaturant.

The Expert Panel for Cosmetic Ingredient Safety (Panel) first reviewed the safety of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate in a report that was published in 1985, with the conclusion that these ingredients are “safe for topical application in the present practices of use and concentration in cosmetics.”² This conclusion was reaffirmed in re-reviews that were published in 2005 and 2017.^{3,4} Excerpts from the summaries of the 1985 report, the draft amended report that preceded the 2005 published re-review, and the 2005 and 2017 re-review summaries are disseminated throughout the test of the document, as appropriate, and are identified by *italicized* text. Additionally, where appropriate, data regarding the monoester metabolites have been included in this safety assessment. This report was placed on the 2024 Cosmetic Ingredient Review (CIR) Priorities List following nomination by the US Food and Drug Administration (FDA) for cause.

This safety assessment includes relevant published and unpublished data that are available for each endpoint that is evaluated. Published data are identified by conducting an extensive search of the world’s literature; a search was last performed in October 2025. A listing of the search engines and websites that are used and the sources that are typically explored, as well as the endpoints that the Panel typically evaluates, is provided on the CIR website (<https://www.cir-safety.org/supplementaldoc/preliminary-search-engines-and-websites>; <https://www.cir-safety.org/supplementaldoc/cir-report-format-outline>). Unpublished data are provided by the cosmetics industry, as well as by other interested parties.

Some chemical and toxicological data on the phthalate ingredients included in this safety assessment were obtained from robust summaries of data submitted to the European Chemicals Agency (ECHA) by companies as part of the Registration, Evaluation, Authorization and Restriction of Chemicals (REACH) chemical registration process.⁵⁻⁷ These data summaries are available on the database for ECHA, and when deemed appropriate, information from the summaries has been included in this report.

CHEMISTRY

Definition and Structure

The definitions and structures of the ingredients included in this review are provided in Table 1. These 3 ingredients are each a short-chain alkyl diester of *o*-phthalic acid, resulting from esterification with butyl, ethyl, or methyl alcohol.

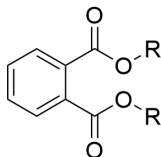


Figure 1. Phthalates, wherein both R substituents are either butyl, ethyl, or methyl.

Chemical Properties

Chemical and properties of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate are described in Table 2. Dibutyl Phthalate is a colorless, oily liquid with a molecular weight of 278.34 g/mol.² The log P_{ow} at 30°C is 4.46.⁵ Diethyl Phthalate is a colorless or pale yellow, oily liquid with a molecular weight of 222.23 g/mol.^{2,6} The log P_{ow} at 41°C is 2.2.⁶ Dimethyl Phthalate is a colorless, oily liquid with a molecular weight of 194.19 g/mol.² The log P_{ow} at 25°C is 1.54.⁷

Method of Manufacture

Phthalate esters can be prepared by the reaction of phthalic acid with alcohol.² Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate are produced industrially by the reaction of phthalic anhydride with butyl alcohol, ethyl alcohol, and methyl alcohol, respectively. Dibutyl Phthalate is manufactured by the esterification of phthalic anhydride with an excess of n-butyl alcohol. Vacuum stripping removes the unreacted n-butyl alcohol. Steam sparging ensures low odor. The phthalate is alkali refined to give a low acid number and is filtered to produce a clear product. The exact manufacturing processes for Diethyl Phthalate and Dimethyl Phthalate are proprietary information.

Impurities

Diethyl Phthalate may contain Dimethyl Phthalate or ethyl methyl phthalate as impurities.² Additional impurities may be isophthalic acid, terephthalic acid, and maleic anhydride.⁸ The purity of phthalate esters has been reported to be > 99%.

Natural Occurrence

Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate have been identified in the organic solvent extracts, root exudates, and essential oils of many different plant species.⁹ These phthalates have also been isolated and purified from various algae, bacteria, and fungi.

USE

Cosmetic

The safety of the cosmetic ingredients addressed in this assessment is evaluated based on data received from the US FDA and the cosmetics industry on the expected use of phthalates in cosmetics. Registration and Listing Data (RLD) obtained from the FDA report frequency of use, and responses to a survey conducted by the Personal Care Products Council (Council) indicate maximum reported concentrations of use; it is these values that define the present practices of use and concentration that are assessed by the Panel. Since 2024, as a result of the Modernization of Cosmetics Regulation Act (MoCRA) of 2022, manufacturers and processors are required to register facilities and list their products (and ingredients therein) with the FDA (i.e. RLD). An exception is made for small businesses (average gross annual sales in the US of cosmetic products for the previous 3-yr period is less than \$1,000,000, adjusted for inflation), which are exempt from MoCRA reporting for most cosmetic product categories. Eye are products, injected products, internal use products, or products that alter appearance for more than 24 h, and the facilities that manufacture these products, are not included in this exemption.¹⁰

According to RLD that CIR received in 2024, Diethyl Phthalate is used in 168 formulations, with most of the uses reported in fragrance preparations (Table 3).¹¹ Additionally, the RLD reported Dibutyl Phthalate is used in 2 manicuring preparations and Dimethyl Phthalate had no uses. The results of the concentration of use survey conducted by the Council in 2025 indicate Diethyl Phthalate has a maximum concentration of use range of 0.1 - 0.15%, with 0.15% reported in leave-on face and neck products.¹² No concentrations of use were reported for Dibutyl Phthalate or Dimethyl Phthalate; however, responses to the survey indicated that Dibutyl Phthalate and Diethyl Phthalate may be present in cosmetics as impurities.

When determining whether to re-open this safety assessment, the Panel considered FDA Voluntary Cosmetic Registration Product (VCRP) survey data submitted to CIR in 2023 as compared to that stated in the previous reports. In 2023, Diethyl Phthalate was reported to be used in 1 skin care formulation; no uses were reported for Dibutyl Phthalate or Dimethyl Phthalate.¹³ When comparing the VCRP data received in 2023 to that received in 2001, the frequencies of use for these phthalate ingredients have greatly decreased since the 2005 re-review was published; Dibutyl Phthalate was reported to have 150 uses (most in manicuring preparations), Diethyl Phthalate was reported to have 73 uses (most in fragrance preparations), and Dimethyl Phthalate was reported to have 12 uses (most in non-coloring hair preparations).³ In the 2005 re-review, the maximum concentration of use range for Dibutyl Phthalate was 0.0038 - 15% (15% reported in manicuring preparations). Diethyl Phthalate was reported to have a maximum concentration of use range of 0.00003 - 11% (11% was reported in perfumes), and Dimethyl Phthalate was reported to have a maximum concentration of use range of 0.00002 - 2% (2% was reported in hair spray).

While not reported in RLD submitted to CIR in 2024, Dibutyl Phthalate a study published in 2011 stated that it was detected in 14 different commercially available black tattoo inks at 0.12 - 691.2 µg/g.¹⁴ In another study from 2017, Dibutyl Phthalate, Diethyl Phthalate and Dimethyl Phthalate had mean concentrations of 0.032, 1670, and 32.25 ppm, respectively, in 42 branded perfumes from 11 countries, including the US.¹⁵ This research group described the mean concentrations of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate in the perfume manufactured in the US to be 0.00095, 3.43, and 43.40 ppm, respectively.¹⁶

Additionally, some of the phthalates may be used in cosmetic powders, and could possibly be inhaled; for example, Diethyl Phthalate is reported to be used in colognes, perfumes, and other fragrance preparations (concentrations not reported). In practice, as stated in the Panel's respiratory exposure resource document (<https://www.cir-safety.org/cir-findings>), most droplets/particles incidentally inhaled from cosmetic sprays would be deposited in the nasopharyngeal and tracheobronchial regions and would not be respirable (i.e., they would not enter the lungs) to any appreciable amount.

Products containing Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate may be marketed for use with airbrush delivery systems. With the advent of MoCRA and the current product categories outlined by the FDA, it is now mandatory that cosmetic products used in airbrush delivery systems be reported as such for some, but not all, product categories in the RLD. In other words, a reliable source of frequency of use data regarding the use of cosmetic ingredients in conjunction with airbrush delivery systems is now available, in some instances. Additionally, the concentration of use surveys are conducted based on product categories as stated in the RLD. None of the reported product categories for these ingredients as listed in the RLD include a designation using airbrush application, so it is possible that this ingredient is used with airbrush delivery systems, but not reported as such. Additionally, no concentration of use data were provided indicating airbrush application. Nevertheless, no consumer habits and practices data or particle size data are publicly available to evaluate the exposure associated with this use type, thereby preempting the ability to evaluate risk or safety. Without information regarding the consumer habits and practices data or product particle size data (or other relevant particle data, e.g., diameter) related to this use technology, the data profile is incomplete, and the Panel is not able to determine safety for use in airbrush formulations. Accordingly, the data are insufficient to evaluate the exposure resulting from cosmetics applied via airbrush delivery systems.

In a review published in 2001, the Research Institute for Fragrance Materials (RIFM) concluded that Diethyl Phthalate is not considered to present any significant toxic liability for its uses as a solvent and vehicle in cosmetic products.¹⁷ No

toxicological endpoints of concern were identified. The Expert Panel for Fragrance Safety has not published a monograph on the phthalate ingredients.

Under European regulations for cosmetic ingredients, Dibutyl Phthalate is listed in Annex II, List of Substances Prohibited in Cosmetic Products.¹⁸ Diethyl Phthalate and Dimethyl Phthalate are not restricted from use in any way under the rules governing cosmetic products in the European Union. In 2004, the Scientific Committee on Cosmetic Products and Non-Food Products (SCCNFP) determined that Dibutyl Phthalate should not be intentionally added to cosmetic products.¹⁹ Dibutyl Phthalate has been classified as a Category 2 toxic substance to reproduction that “may cause harm to the unborn child”. In 2002, the SCCNFP opined that the safety profile for Diethyl Phthalate supported its use in cosmetic products;²⁰ this opinion was reaffirmed by the SCCNFP in 2003²¹ and the Scientific Committee on Consumer Products (SCCP) in 2007.²²

Non-Cosmetic

Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate are used as solvents and plasticizers for nitrocellulose, cellulose acetate, and cellulose acetate-butyrate compositions.² They are used in the manufacture of varnishes and plastics and in insecticides and insect repellents. Dibutyl Phthalate is used as a plasticizer in explosives and elastomers, such as polyvinyl, as a textile lubricating agent, as a resin solvent, and in safety glass, printing inks, paper coatings, and adhesives. Dimethyl Phthalate is used as a camphor substitute in the manufacture of celluloid, as a wetting agent, and as an alcohol denaturant.

Phthalate esters are ubiquitous in the modern environment, found in clothing textiles, packaged foods, and even in household dust.⁸ Dibutyl Phthalate and other phthalates are primarily used as plasticizers in flexible polyvinyl chloride and in cellulose plastics. Additional applications include uses as solvents, lubricants, fixatives, and detergents. Dibutyl Phthalate is used in the manufacture of medical products including denture base materials, and ocular prosthetic. Other products containing Dibutyl Phthalate include plastic wraps and containers for food, paints, plastic toys and inks. Diethyl Phthalate is found in many plastic consumer products and plastic-coated materials, packaging for food and drug products. Dimethyl Phthalate is used as an insect repellent and, with permethrin, as a fabric treatment on tents to protect campers from mosquitoes.

Globally, Dibutyl Phthalate has been detected in commercial polyethylene terephthalate (PET) bottled water at concentrations as high as 82.8 µg/l.²³ Analysis of feminine hygiene products (pads, panty liners, and tampons) in the US found Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate to have mean concentrations of 0.90, 1.35, and 0.55 ng/g, respectively.²⁴ Another study (published in 2020) reported that feminine bactericidal creams and possibly cosmetic deodorant sprays and feminine powders had measurable concentrations of Dibutyl Phthalate (mean 0.52 ng/g), Diethyl Phthalate (mean 0.51 ng/g), and Dimethyl Phthalate (mean 0.60 ng/g).²⁴

In 2012, the FDA issued guidance for industry on limiting the use of Dibutyl Phthalate as an excipient in regulated drug and biologic products, recommending that use of this ingredient be avoided if an alternative is available.²⁵ Dibutyl Phthalate (> 1000 ppm or 0.1%) has been prohibited in the manufacturing, sale, distribution, or importation of children’s toys and childcare articles (16CFR§1307). Diethyl Phthalate is a plasticizer in food packaging material (21CFR§181.27) and an authorized denaturant (27CFR§21.71, 21.72, 21.106, and 21.151). The European Food Safety Authority established a tolerable daily intake for Dibutyl Phthalate at 0.01 mg/kg bw/d based on reproductive effects.²⁶

TOXICOKINETIC STUDIES

Dermal Penetration

In Vitro

The dermal absorption rates of undiluted phthalate esters were determined in an in vitro study using rat and human skin.⁸ The diffusion of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate across the epidermal samples was measured for various exposure periods. In rat skin (9-11 samples), the mean steady-state absorption rates for Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were 93.35 ± 0.94 , 413.67 ± 92.79 , and 410.55 ± 41.77 µg/cm²/h, respectively. In human skin (11-15 samples), the mean steady-state absorption rates for Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were 2.40 ± 0.63 , 12.75 ± 1.12 , and 30.51 ± 6.43 µg/cm²/h, respectively.

In a study of the percutaneous absorption of Dibutyl Phthalate using static diffusion cells with full-thickness rat skin (haired and hairless; 1.76 cm²), an infinite dose of ¹⁴C-Dibutyl Phthalate (50 mg/cm²) was applied to the stratum corneum side of the skin samples for 24 h.⁸ Aliquots (200 µl) of the receptor fluid on the dermis side of the skin sample were collected periodically over the 24-h period. The percutaneous absorption fluxes of Dibutyl Phthalate across the skin sections of haired and hairless male rats were 26 ± 1 and 39 ± 1 µg/cm²/h, respectively. All of the radioactivity contained in the receptor fluid had the same HPLC retention time as monobutyl phthalate. Addition of an esterase inhibitor caused a reduction of 99% of the absorption flux of ¹⁴C, and monobutyl phthalate was barely detectable in the receptor fluid.

In another absorption study, shaved rat skin and human breast skin were exposed to Diethyl Phthalate.⁸ Absorption through the rat skin was rapid and extensive. There was an initial lag for the first 8 h during which the absorption was 1.7%. From 8 - 72 h, absorption was almost linear, reaching 35.9% on occluded skin and 38.4% on unoccluded skin. The mean

rate of absorption in rat skin between 12 - 72 h was 0.094 mg/cm²/h (occluded) and 0.103 mg/cm²/h (unoccluded). The percutaneous absorption of Diethyl Phthalate through human skin was slow and incomplete. The initial lag time was 10 h. At 8 h, 0.1% had absorbed in occluded skin and 0.2% in unoccluded skin. At 72 h, 3.9 and 4.8% of applied Diethyl Phthalate had absorbed in occluded and unoccluded skin, respectively. The mean rate of absorption of Diethyl Phthalate through human breast skin between 12 - 72 h was 0.011 mg/cm²/h (occluded) and 0.014 mg/cm²/h (unoccluded). In this in vitro model, the rate of absorption of Diethyl Phthalate through rat skin was significantly ($p < 0.05$) faster than through human skin.

The in vitro permeation of Diethyl Phthalate was studied using abdominal guinea pig skin mounted into a Franz-type diffusion cell.⁸ The donor side of the skin was pre-treated for 2 h with 10 mM sodium dodecyl sulfate, 10 mM benzalkonium chloride, or 0.5 % polyoxyethylene (10) oleyl ether. After the surfactant solution was rinsed, a solution of 0.05 % Diethyl Phthalate was added to the donor side for 16 - 24 h. A control condition consisted of Diethyl Phthalate treatment without surfactant pre-treatment. Additional skin samples were treated with 0.05% Diethyl Phthalate in a mixture with each of the respective surfactants described above, without surfactant pre-treatment. Skins pre-treated with sodium dodecyl sulfate, benzalkonium chloride, or polyoxyethylene (10) oleyl ether had respective permeation rates of 190, 174, and 110% of the permeation rate of the control condition. Co-treatment of Diethyl Phthalate with sodium dodecyl sulfate, benzalkonium chloride, or polyoxyethylene (10) oleyl ether resulted in Diethyl Phthalate permeation rates of 82.0, 105.7, or 99.5% of the control. The cumulative amount of Diethyl Phthalate permeation through guinea pig skin without surfactant at 24 h was 21.8 µg/cm².

In an in vitro absorption study, ¹⁴C-Dibutyl Phthalate (radiochemical purity > 97%; specific activity 21.1 mCi/mmol) in an oil-in-water emulsion was applied to flow-through diffusion cells (0.64 cm²) using hairless guinea pig dermatomed skin (200 - 320 µm thick) at approximately 0.5 µCi test material/cm².²⁷ Absorption was measured at 24 and 72 h from the receptor fluid consisting of HEPES-buffered Hanks' balanced salt solution and 4% bovine serum albumin. At the end of the 24 or 72 h experiments, the skin was washed and rinsed to measure unabsorbed test material. Tape stripping was utilized to determine how much test material remained in the skin after 24 h. The absorption of Dibutyl Phthalate in the receptor fluid over 24 h was 27.1 ± 1.9%, with 29.2 ± 2.4% remaining in the skin (total skin penetration over 24 h = 56.3 ± 2.7%). Over 72 h, the amount of Dibutyl Phthalate in the receptor fluid was 59.9 ± 3.2%, with 13.7 ± 3.1% remaining in the skin (total skin penetration over 72 h = 75.6 ± 4.7%).

A percutaneous absorption study of Dimethyl Phthalate and 2 other phthalates (dipropyl phthalate and dipentyl phthalate) tested these materials using different carriers, such as ointments, creams, and lotions.²⁸ The study used rat skin in Franz diffusion cells using PBS as the receptor fluid. There was 0.2 g of phthalate per 10 g of carrier. The lag time of the ointment, base cream, and lotion containing Dimethyl Phthalate were 0.891, 0.888, and 0.885 h, respectively. The permeation rate of the ointment, base cream, and lotion containing Dimethyl Phthalate were 1.391, 1.216, and 1.120 µg/cm²/h, respectively. Of the 3 phthalates tested, Dimethyl Phthalate had the fastest permeation rate, regardless of carrier type.

Animal

The dermal absorption of Dibutyl Phthalate was studied in 3 female hairless guinea pigs.²⁷ ¹⁴C-Dibutyl Phthalate (radiochemical purity > 97%; specific activity 21.1 mCi/mmol) in an oil-in-water emulsion was applied in a single dose for 24 h at approximately 0.5 µCi test material/cm² (dosage of 1 mg/cm²; application site 3.0 cm²). The test site was covered to prevent test material from being rubbed off. Urine and feces were collected during the application period. At the end of 24 h, the guinea pigs were killed and the test site was washed and rinsed. The washes and rinses were analyzed to determine how much test material was not absorbed. Skin from the treatment site, ovaries, livers, and kidneys were analyzed for systemic absorption. Approximately 62.0 ± 2.0% of the applied dose was systemically absorbed. Most of the Dibutyl Phthalate was found in the urine (60.4 ± 1.8%) and the amount systemically absorbed in the ovaries, kidneys, and liver were less than 2%. Approximately 2% of Dibutyl Phthalate remained on the skin.

Absorption, Distribution, Metabolism, and Excretion

In Vitro

Dibutyl Phthalate and Dimethyl Phthalate (0.4 mg/ml each) almost completely disappeared after 2 h of incubation with rat liver homogenates, while action with rat kidney homogenates was slower.² Approximately 90% of Dibutyl Phthalate and 95% of Dimethyl Phthalate disappeared during a 5-h incubation. In another study, Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate (1 mg/ml each) was incubated for 16 h with the contents of rat stomach, small intestines, or cecum. The test material was metabolized rapidly to monobutyl phthalate when incubated with the contents of rat small intestine, while metabolism was slower in the presence of rat cecal contents. Monobutyl phthalate was the only product of metabolism; complete hydrolysis to phthalic acid did not occur.

In an in vitro study of the intestinal absorption of Dibutyl Phthalate and Dimethyl Phthalate using an everted gut-sac preparation from the rat small intestine, most of Dibutyl Phthalate and Dimethyl Phthalate hydrolyzed to the corresponding monoester before crossing the intestinal mucosa.² Only 4.5% of Dibutyl Phthalate and 18.8% of Diethyl Phthalate crossed the intestine intact. Inhibition of mucosal esterases by an esterase inhibitor reduced the amount of Dibutyl Phthalate hydrolyzed to monobutyl phthalate. Approximately the same amount of intact Dibutyl Phthalate was absorbed by the intestine

with and without the esterase inhibitor. Intestinal absorption of these compounds may be controlled by the hydrolysis of Dibutyl Phthalate to monobutyl phthalate.

Animal

Dermal

The dermal absorption and distribution/excretion fate of radiolabeled phthalates was examined in using F344 rats.⁸ ¹⁴C-Dibutyl Phthalate, ¹⁴C-Diethyl Phthalate, or ¹⁴C-Dimethyl Phthalate (157 μ mol/kg each) was applied to a 1.3 cm diameter patch of shaved skin and covered with a plastic cap. Rats were then placed in a metabolic cage that allowed separate collection of urine and feces which were collected every 24 h for 7 d. ¹⁴C was excreted in the urine and feces at a nearly constant rate of 10 - 12% of the ¹⁴C-Dibutyl Phthalate dose per 24 h throughout the 7 d of collection. Approximately 24 times as much ¹⁴C was excreted in the urine as in feces. For ¹⁴C-Diethyl Phthalate, 24% was excreted in the first 24 h, then slowed to 11% in the second 24 h, and slowed to 5% every 24 h for the remaining period of measurement. The rate of ¹⁴C-Dimethyl Phthalate excretion was a relatively constant 6 - 7.5% per 24 h over the entire 7 d. Rats were killed on the seventh day, and various tissues and organs were analyzed for the presence of ¹⁴C. Most of the absorbed phthalate had been excreted in the urine and feces. Very little was retained in the tissues.

In another experiment in this report, 10 μ l/cm² neat ¹⁴C-Dibutyl Phthalate was applied to the shaved backs of male Sprague-Dawley rats and covered with a perforated plastic cap.⁸ Groups of 3 to 8 rats were killed at several time points between 0.5 - 72 h after application of the test material. Blood, urine and feces were collected for analysis. Dibutyl Phthalate penetrated the skin rapidly. Approximately 30 min after application, 20% of the applied dose had entered the skin. Within the first 8 h, the absorption flux was 43 μ g/cm²/h. Between 8 - 48 h, the rate of absorption flux increased to 156 μ g/cm²/h. At the 24- and 72-h time points, the radiolabeled test material that absorbed into the skin had radially diffused into the surrounding stratum corneum and/or epidermis. Dibutyl Phthalate was hydrolyzed by skin esterase before reaching systemic circulation. At all time points, unchanged Dibutyl Phthalate accounted for 2% of the plasma radioactivity, whereas monobutyl phthalate and monobutyl phthalate-glucuronide combined accounted for 61 - 85% of the plasma radioactivity. The mean ratio of plasma monobutyl phthalate to monobutyl phthalate-glucuronide was 1.3. In the urine, monobutyl phthalate and monobutyl phthalate-glucuronide together accounted for 50 to 66% of the urinary radioactivity, with no significant correlation to time. The mean ratio of monobutyl phthalate to monobutyl phthalate-glucuronide in urine was 0.73. Over the maximum 72-h exposure, the total urinary excretions of monobutyl phthalate and monobutyl phthalate-glucuronide were 14.1 ± 1 and $20 \pm 1\%$ of the applied dose, respectively. Similar experiments were conducted using Sprague-Dawley female rats and hairless rats for comparison. Female haired rats showed similar Dibutyl Phthalate absorption rates as the males. Male hairless rats had a higher percutaneous absorption of Dibutyl Phthalate (75% of the applied dose). The urinary excretion rate of ¹⁴C-Dibutyl Phthalate in hairless rats was 2.5-fold higher than that of the haired rats.

¹⁴C-Diethyl Phthalate was applied topically to rabbits.² Analysis of urine indicated that approximately 9% of the radioactivity was excreted after 24 h, 14% after 48 h, and 16 to 20% within 72 h. After 3 d of topical exposure, tissue distribution was determined by autoradiography. Radioactivity was detected in the lung, heart, liver, kidney, gonads, spleen, and brain. It was not detected in the skin and subdermal fatty tissue at the site of application.

Radiolabeled Diethyl Phthalate was applied on the shaved backs (25 mm diameter; occluded) of 3 female albino rabbits at approximately 299 - 569 mmol per rabbit.⁸ Blood samples were collected at 1, 24, 48, 72, and 96 h after application, and urine and feces were collected at 24, 48, 72, and 96 h. At 96 h after application, the rabbits were killed and tissues from the vital organs were collected. The results of the study suggested that Diethyl Phthalate is rapidly absorbed through the skin, widely distributed to tissues, and actively excreted. Approximately 50% of the applied Diethyl Phthalate was excreted in 4 d. It was estimated that it would take approximately 1 mo to eliminate the entire Diethyl Phthalate dose applied. There was evidence of higher deposition of Diethyl Phthalate in the adrenal tissue and kidneys than in other tissues.

Oral

Dibutyl Phthalate was administered by gavage to male rats in 2 doses of 0.2 ml 24 h apart.² Urine was collected for 48 h after the first dose, and Dibutyl Phthalate and its metabolites were quantitated. A total of 24.6% of the phthalate moiety was recovered in the urine. The recovered phthalate moiety consisted of 89.8% monobutyl phthalate, 2.7% phthalic acid, 0.4% intact Dibutyl Phthalate, and four other metabolites in very small amounts. The researchers suggested that Dibutyl Phthalate was metabolized by hydrolysis of one ester bond and both terminal and subterminal oxidation of the remaining alkyl chain. The resulting primary and secondary alcohols were, presumably, further oxidized to acids and ketones, respectively.

In a study in which male mice were administered ¹⁴C-Dibutyl Phthalate orally, the radioactivity accumulated in the liver and kidney within 6 h.² The radioactivity was rapidly excreted in the urine and feces. In another oral study in mice, no radioactivity was recovered from hepatic DNA after administration of ¹⁴C Dibutyl Phthalate. Dibutyl Phthalate and its metabolites appeared not to be transported into the nuclei.

¹⁴C-Dibutyl Phthalate was administered orally in dimethyl sulfoxide (DMSO) in a dose of 60 mg/kg or intravenously in saline in a dose of 10 mg/kg to male rats.² Urine and feces were collected, and the amount of radioactivity excreted was

determined. The amount of administered radioactivity excreted varied from 81.4 to 97.7% in the urine and from 1.0 to 8.2% in the feces in the first 24 h after oral or intravenous administration of Dibutyl Phthalate. Several rats were killed, and tissue distribution of radioactivity was determined. Brain, heart, liver, lung, spleen, muscle, adipose, stomach, prostate, and thymus tissues, blood, and the intestinal contents were examined 24 h after oral or intravenous administration of Dibutyl Phthalate. Very little radioactivity was recovered. The elimination of Dibutyl Phthalate from tissues and organs was rapid, and no organ had any significant affinity for accumulation. Rats were administered labeled ^{14}C -Dibutyl Phthalate orally, and bile was collected. Approximately 27.6 to 52.8% of radioactivity was excreted in the bile within 24 h after oral administration of Dibutyl Phthalate. Since more radioactivity was excreted in the bile than in the feces, there was apparently good absorption of Dibutyl Phthalate and its metabolites from rat intestine.

Urinary metabolites were identified in male rats, male hamsters, and male guinea pigs given a single oral dose of 60 mg/kg Dibutyl Phthalate.² All 24-h urine samples contained monobutyl phthalate as the major product, intact Dibutyl Phthalate, phthalic acid, monobutyl phthalate glucuronide, and two other monobutyl phthalate oxidation products. The hamster urine contained an additional oxidation product. The livers from rats were examined 1 h after intravenous dosing of Dibutyl Phthalate and the data obtained indicated that Dibutyl Phthalate was rapidly hydrolyzed to monobutyl phthalate by the microsomal fraction. No phthalic acid was detected. The bile contained monobutyl phthalate and intact Dibutyl Phthalate but not phthalic acid. Since phthalic acid was detected in the urine, it was suggested that its formation must occur at other sites than the liver. It was concluded that the hydrolysis of Dibutyl Phthalate to monobutyl phthalate occurred in the liver, that there was enterohepatic circulation of Dibutyl Phthalate and its metabolites and good absorption from the intestine, and that monobutyl phthalate was the main metabolite of Dibutyl Phthalate and was primarily excreted in urine.

A 500 mg/kg dose of ^{14}C -Dibutyl Phthalate in ethanol was administered by gastric intubation to male rats and the bile was collected every hour for 6 h.² At 6 h after oral administration of Dibutyl Phthalate, 4.5% of the radioactivity was recovered in the bile. At 5 h after intravenous injection of Dibutyl Phthalate, 10% of the radioactivity was detected in the bile. Dibutyl Phthalate bile metabolites included monobutyl phthalate, intact Dibutyl Phthalate, phthalic acid, a monobutyl phthalate glucuronide, and traces of other glucuronides. A small amount of Dibutyl Phthalate appears to be absorbed unaltered from the intestine, and the excretion of Dibutyl Phthalate through the biliary route has a role in its metabolic fate.

In another study, at least 90 % of orally ingested Dibutyl Phthalate was absorbed from the gastrointestinal tract.⁸ Once absorbed, Dibutyl Phthalate distributed to the liver, adipose, kidney, muscle, and some other tissues, but it did not remain in these tissues, thus accumulation in the body is unlikely. Dialkyl phthalates are metabolized by hydrolytic enzymes found in many tissues to produce the corresponding monoester and alcohol. A second hydrolysis can further metabolize the monoester into an alcohol and phthalic acid, but in primates and man the monoester is more likely to be conjugated to glucuronide for excretion. The major route of excretion of phthalates in rats, primates, and humans is in the urine. Very small amounts of Dibutyl Phthalate metabolites are found in the feces, but they have been detected in the bile at levels up to 44% of the Dibutyl Phthalate dose given to rats (60 mg/kg orally).

Single doses of 0.27 g/kg or 2.31 g/kg of 7- ^{14}C -Dibutyl Phthalate were given by oral intubation to Wistar rats.⁸ Urine and feces were collected at 4, 8, 24, and 48 h after dosing. At the same time points, selected animals were killed, and their tissues and organs were weighed and analyzed for metabolites of 7- ^{14}C -Dibutyl Phthalate. By 48 h after dosing, 80 to 90% of the test material administered had been excreted in urine or feces as phthalic acid, monobutyl phthalate, mono(3-hydroxybutyl) phthalate, or mono(4-hydroxybutyl) phthalate. Less than 0.001% of the dose was found in the tissues at 48 h. The radioactivity was cleared more rapidly from the body in the lower dose group (rates not reported). Furthermore, rats fed a diet of 1% Dibutyl Phthalate for 12 wk did not accumulate Dibutyl Phthalate or its monoester metabolites in the tissues or organs.

Dibutyl Phthalate was reported to absorb in the intestines of rats mostly as monobutyl phthalate following oral exposure.⁸ However, inhibition of intestinal epithelial esterases by S,S,S-tributylphosphorotrithioate (DEF) (8 mg/kg) did not significantly reduce circulating ^{14}C after oral administration of 7- ^{14}C -Dibutyl Phthalate at 6, 10, 14, 25, or 30 g/kg in rats. Blood concentrations of circulating ^{14}C -monobutyl phthalate were not measured. Approximately 4 h after a single dose of 837 mg/kg ^{14}C -Dibutyl Phthalate, rats that had been treated with 8 mg/kg DEF had a significantly lower ($p < 0.01$) concentration of the diester in the intestinal lumen, but there was no effect of DEF on the monoester in the lumen.

In both rats and hamsters, the major urinary metabolite of Dibutyl Phthalate or monobutyl phthalate was monobutyl phthalate-glucuronide following oral exposure.⁸ However, the levels of unconjugated monobutyl phthalate in rat urine were 3- to 4-fold higher than in hamster urine. The activities of β -glucuronidase in the livers of rats and hamsters was comparable, however the testicular β -glucuronidase activity was significantly higher in rats than in hamsters. Because β -glucuronidase deconjugates monobutyl phthalate-glucuronide to free monobutyl phthalate, this difference in enzyme activity between species may contribute to the quantitative differences in urinary metabolites and to the male developmental toxicity of Dibutyl Phthalate and monobutyl phthalate (see Reproductive and Developmental Toxicity section).

Pregnant Sprague-Dawley rats were given a single oral dose of radiolabeled Dibutyl Phthalate (up to 3 g/kg) on gestation day 14.⁸ Tissues from the dams and embryos were analyzed for Dibutyl Phthalate, monobutyl phthalate, and monobutyl phthalate-glucuronide for several time points up to 48 h after dosing. The percent of administered dose in the placenta and embryonic tissue peaked at 0.3 and 0.15%, respectively, at 2 h after dosing. By 48 h after dose administration,

no detectable phthalates were found in these tissues nor in the uterus, ovaries, or amniotic fluid. Although Dibutyl Phthalate was shown to cross the placenta into fetal tissues, neither Dibutyl Phthalate nor its metabolites accumulated in the fetus, placenta or in maternal tissues.

Dimethyl Phthalate was administered by gavage to male rats in a single dose of 0.1 ml, and the urine was collected for 24 h.² A total of 44.6% of the phthalate moiety was recovered in the urine, and it consisted of 77.5% monomethyl phthalate, 14.4% phthalic acid, and 8.1% intact Dimethyl Phthalate. Dimethyl Phthalate appeared to be metabolized only by hydrolysis of one or both ester groups.

CD-1 female mice were used in a study to determine the distribution of Dibutyl Phthalate (99.6% pure) after single or repeated doses via gavage.²⁹ Groups of 5 mice/treatment received 0, 1, 10, or 1000 mg/kg Dibutyl Phthalate in tocopherol-stripped corn oil either once or once daily for 10 d. The mice were killed at 2, 6, 12, or 24 h after the single dose or the final dose, and tissues, including liver and ovaries, and serum were collected for analysis of monobutyl phthalate. Regardless of duration of exposure (single or repeated), monobutyl phthalate was detected in the sera and tissues of Dibutyl Phthalate-treated mice. In single dose mice, monobutyl phthalate levels peaked at ≤ 6 h and fell close to background levels by 24 h post-treatment. Following the last repeated dose, monobutyl phthalate levels peaked at ≤ 2 h and fell to background levels by 12 h.

Parenteral

In a study in which male mice were administered ¹⁴C-Dibutyl Phthalate intravenously, the radioactivity accumulated in the liver and kidney within 1 h.² Diester and/or metabolic products were present in maternal blood, fetal tissue, amniotic fluid, and placentas after day 8 or day 11, respectively, and throughout gestation after ¹⁴C-Diethyl Phthalate was administered intravenously to pregnant rats on day 5 or day 10 of gestation.

In another study, Sprague-Dawley rats were given a single intravenous injection of 1 or 10 mg/kg ¹⁴C-Dibutyl Phthalate.⁸ After dosing, the animals were placed in individual metabolic cages for collection of urine and feces for 72 h. Arterial blood was collected at 10 s and 72 h after dose administration. The main route of excretion of the radiolabel was in urine (85% of the administered dose), followed by feces (9 - 10% of the administered dose). More than 95% of the radiolabel in urine was excreted within the first 24 h. Approximately 72 h after the dose administration, 0.6 - 2% of the administered dose remained in the exsanguinated rat carcass, and 1.9 - 2.3% remained in the blood. Unchanged Dibutyl Phthalate was undetectable in the urine. Monobutyl phthalate in urine accounted for $22 \pm 3.7\%$ of the administered dose. Monobutyl phthalate-glucuronide in urine accounted for $34.6 \pm 3.6\%$ of the administered dose.

Human

Dermal

Dimethyl Phthalate was reported to absorb through human skin, and its metabolites were detected in human urine.² (No further details available.)

The absorption, metabolism, and extraction of Dibutyl Phthalate after dermal application was studied in 26 Caucasian male subjects.³⁰ The study was single-blinded and the subjects were given daily whole-body topical applications of a basic cream (control; 2 mg/cm²) for 5 d and then a cream containing 2% (w/w) of Dibutyl Phthalate for another 5 d. Showering, bathing, or swimming were not allowed until 4 h after each application. For each subject, 24-h urine pools were collected in both weeks, except the first 24-h sample of the treatment week, where every single urine sampled was collected individually to prepare a 24-h excretion profile for the test material. The urine samples were analyzed for urinary total and unconjugated monobutyl phthalate metabolites by liquid chromatography-tandem mass spectroscopy. All 26 subjects had increased excretion of monobutyl phthalate after topical application. Total monobutyl phthalate excreted in the urine during the treated week was 11.8 ± 0.6 mg/24 h. On average, 1.82% of the applied Dibutyl Phthalate was recovered in urine as monobutyl phthalate. The urinary monobutyl phthalate excretion rate was relatively low in the first 4 h following application of the test cream but increased thereafter to a relatively constant rate over the next 20 h. The fraction of unconjugated monobutyl phthalate was 8.0%.

Dermal and Inhalation

In a controlled chamber study, 6 adult males wearing only shorts were exposed to airborne Dibutyl Phthalate and Diethyl Phthalate for two 6-h sessions.³¹ During one session, subjects wore breathing hoods that supplied clean air to quantify dermal exposure only ("hood-on"); during the other session, both dermal and inhalation exposure occurred ("hood-off"). Urine was collected before, during, and for 48 h after exposure and analyzed for key Diethyl Phthalate and Dibutyl Phthalate metabolites (e.g., mono-n-butyl phthalate and mono-3-hydroxy-n-butyl phthalate). The results demonstrated high levels of Diethyl Phthalate and Dibutyl Phthalate metabolite excretions while in the chamber and during the first 24 h once leaving the chamber under both exposure conditions.

Modeling of the urinary metabolite data was performed by linking inhalation and transdermal exposure modules to a simplified one-compartment pharmacokinetic model that predicted both the timing and magnitude of metabolite excretion. Each exposure route was simulated separately ("inhalation-only" and "dermal-only"), as well as jointly ("combined exposure"), using parameters derived independently from prior experiments. When the model was extrapolated to typical U.S. indoor air concentrations of Dibutyl Phthalate (approximately 0.2 $\mu\text{g}/\text{m}^3$), the estimated daily dermal uptake (0.14 $\mu\text{g}/\text{kg}$

bw/d) exceeded the inhalation intake (0.042 µg/kg bw/d) by roughly threefold. Based on these steady-state simulations, the combined dermal and inhalation uptake of airborne Dibutyl Phthalate was estimated to account for ~ 25% of total exposure to Dibutyl Phthalate in the general U.S. population.³¹

Oral

Groups of 8 volunteers received a single oral dose of 255 or 510 µg isotope-labeled Dibutyl Phthalate (type of isotope not reported).⁸ Approximately 24 h after the doses were given, the monobutyl phthalate content in the subjects' urine was determined. Subjects in the low and high dose groups respectively excreted 64 and 73% of the dose as monobutyl phthalate in the first 24 h. The amount of conjugated monobutyl phthalate was not measured.

In an oral study, an individual male ingested 60 µg/kg deuterium-labeled Dibutyl Phthalate (tetra substituted ring - ²H-3,4,5,6; 5.38 mg total) in decaffeinated coffee served in an edible cup to ensure full dose was consumed (no further description).³² Urine was collected at 20 time points up to 48 h post-dosing. Serum and saliva were collected no earlier than 15 min after eating or drinking at 7 time points up to 24 h post-dosing. Quantification of metabolites was performed with liquid chromatography-tandem mass spectrometry (LC-MS/MS). The majority of the dose (92.2%) was excreted in the first 24 h, with only < 1% of the dose excreted on day 2. The major metabolite was monobutyl phthalate (84%). Various side-chain oxidized metabolites accounted for approximately 8%. The Dibutyl Phthalate metabolites reached peak concentrations between 2 to 4 h post-exposure, followed by a monotonic decline. The elimination halftime of monobutyl phthalate was 2.6 h; oxidized metabolites had elimination halftimes estimated to be from 2.9 to 6.9 h. Monobutyl phthalate was detected in the saliva, and monobutyl phthalate plus side chain oxidized metabolites were detected in the serum.

TOXICOLOGICAL STUDIES

Acute Toxicity Studies

Dermal

The dermal LD₅₀ for Diethyl Phthalate in adult albino rats was > 10 ml/kg. The acute dermal LD₅₀ for Dimethyl Phthalate in rabbits was > 10 ml/kg.²

Oral

In mouse studies, the acute oral LD₅₀ values of Dibutyl Phthalate ranged from 9.77 to 17.0 g/kg, while in multiple rat studies, the LD₅₀ values ranged from approximately 8 g/kg to 23.0 g/kg.² The approximate lethal dose of commercial- grade Dibutyl Phthalate administered orally to male albino rats was found to be > 7500 mg/kg; no clinical signs of toxicity or pathological findings were observed 11 d after dosing.⁸ The LD₅₀ in rats of a nail polish formulation containing 9% Dibutyl Phthalate was > 5 ml/kg, while in another study of a nail preparations containing 6% Dibutyl Phthalate, the LD₅₀ was > 5 g/kg.² The oral LD₅₀ for Diethyl Phthalate was > 5 - 9 ml/kg in rats^{2,8} and 1.0 g/kg in rabbits.² Undiluted Dimethyl Phthalate had acute oral LD₅₀ values for mice, rats, guinea pigs, and rabbits as 7.2, 6.9, 2.4, and 4.4 ml/kg, respectively. Another study reported the acute oral LD₅₀ for Dimethyl Phthalate in mice to be 6281 mg/kg.⁸

Parenteral

The acute intraperitoneal LD₅₀ Dibutyl Phthalate in a study with female mice was 14.9 mmol/kg.² In other studies with mice, the LD₅₀ ranged from 3.57 g/kg (in males) to 4.00 g/kg (sex not specified). In female rats, the intraperitoneal LD₅₀ was reported to be 3.05 ml/kg. In acute intraperitoneal studies with Diethyl Phthalate, the LD₅₀ in mice ranged from 2.83 to 3.22 g/kg, and the LD₅₀ in a female rat study was 5.06 ml/kg. The intraperitoneal LD₅₀ in mice following administration of Dimethyl Phthalate ranged from 1.58 g/kg to 3.98 g/kg; an intraperitoneal study in female rats reported the LD₅₀ to be 3.38 mg/kg. The acute intravenous LD₅₀ of Dibutyl Phthalate to male mice was 0.72 g/kg. The intramuscular administration of Dibutyl Phthalate in a dose of 4 g/kg to 3 rats and 8 g/kg to 3 rats did not result in any deaths, and there was no effect on the growth of the rats. The subcutaneous LD₅₀ of Diethyl Phthalate in guinea pigs was ≥ 3 g/kg.

Acute toxicity studies are summarized in Table 4. In oral rat studies, the LD₅₀ for Dibutyl Phthalate was 6279 mg/kg bw for both sexes.⁵ The LD₅₀ for Diethyl Phthalate was > 5 ml/kg bw for both male and female rats.⁶ For Dimethyl Phthalate (85% pure), the combined LD₅₀ in male and female rats was 5.1 g/kg bw.⁷ In inhalation studies with rats, the LC₅₀ value for Dibutyl Phthalate in both sexes was estimated to be ≥ 15.68 mg/l, and the LC₅₀ for Diethyl Phthalate in both sexes was > 4.64 mg/l.^{5,6} *Danio rerio* (zebrafish) were exposed to Dibutyl Phthalate (up to 10 mg/l) for 96 h.⁵ The LC₀ was 1.3 mg/l and the LC₁₀₀ was 5 mg/ml.

Short-Term Toxicity Studies

Dermal

The National Toxicology Program (NTP) conducted a 4-wk study on the dermal effects of Diethyl Phthalate on mice and rats.⁸ The mice received up to 100 µl of the test material neat and the rats received up to 300 µl. There was no evidence of dermal toxicity and no adverse clinical signs in mice or rats at any dose tested. Feed consumption and weight gains were unaffected by Diethyl Phthalate treatment. The absolute and relative liver weights in female mice in the 25 and 50 µl groups and relative liver weights and relative kidney weights of male and female rats in the 100 µl groups were greater than those of control animals. No other adverse effects of Diethyl Phthalate were observed. The no-observed-adverse-effect level (NOAEL) was determined to be 12.5 µl in mice and 75 µl in rats.⁶

Oral

In a short-term study of 6 rats that received repeated oral doses of 7500 mg/kg Dibutyl Phthalate (commercial grade), 2 rats each died after the second and fourth dosing.⁸ The 2 surviving rats (10 doses total) showed clinical signs of toxicity and weight loss but had completely recovered 10 d after the last treatment. In a 14-d feed study in male and female mice exposed to up to 1% Dibutyl Phthalate, body weights were reduced ($p < 0.05$) in males in the 1% exposure group and liver weights were increased ($p < 0.05$) in males and females in the 1% group. There were no effects on feed or water consumption, mortality rate, or incidence of clinical signs due to Dibutyl Phthalate exposure in this study.

Relative weights of liver, kidney, and spleen were increased in male rats that were fed a powdered diet containing 0.5 or 5% Dibutyl Phthalate or 0.5 or 5% monobutyl phthalate for 34 - 36 d.⁸ Succinate dehydrogenase and pyruvate dehydrogenase activities in liver mitochondria were reduced in rats exposed to 5% Dibutyl Phthalate or 0.5 or 5% monobutyl phthalate, but glutamate dehydrogenase was not affected. Serum chemistry analysis revealed increases in ALP, GOT, GPT, and CPK in rats exposed to 5% Dibutyl Phthalate. Serum globulin was reduced, and A/G was increased at both concentrations of Dibutyl Phthalate and monobutyl phthalate. Both concentrations of monobutyl phthalate decreased TG and uric acid levels. Dibutyl Phthalate and monobutyl phthalate at 5% produced increased liver necrosis (single cells and zonal) and hypospermatogenesis. Microscopic examination of hepatocytes revealed that 5% Dibutyl Phthalate and 5% monobutyl phthalate caused increased peroxisomes, lysosomes, and mitochondria. Increased mitochondrial swelling and decreased liver glycogen was observed in the rats treated with 5% monobutyl phthalate.

In a 6-wk study in rats that received 1 ml/kg Dibutyl Phthalate in oil 2 times/wk, no adverse effects were reported.² Growth was inhibited in mice and leukocytosis was observed in rats that received 20 mg/kg Dibutyl Phthalate for over 11 wk. Rabbits that received 3 ml/kg/d Diethyl Phthalate for 8 d displayed no adverse effects for up to 2 wk after dosing.

Short-term oral toxicity studies are summarized in Table 5. In a 28-d study, no clinical signs of toxicity were observed in rats that received up to 7000 ppm Dibutyl Phthalate in feed, with or without the addition of 3000 ppm diethylhexyl phthalate.³³ Relative liver weights were increased in males and females of the 7000 ppm Dibutyl Phthalate group and in males of the 1000 ppm Dibutyl Phthalate group. In the groups that received both phthalates, an increase of absolute and relative liver weights were observed in all animals in all combined treatment groups. Mice that received 40 mg/kg bw/d Dibutyl Phthalate for 60 d via gavage had no significant differences in body weight gain or lung function parameters when compared to saline controls.³⁴ However, oxidative stress biomarkers in lung tissue were significantly altered. Serum total IgE levels showed a statistically significant increase compared to controls, whereas cytokine levels and inflammatory cell counts in bronchoalveolar lavage (BAL) fluid were not significantly different from controls. In a 9-wk oral gavage study, male rats treated with Dibutyl Phthalate at 750 mg/kg bw/d had marked histopathological changes in the lungs, including increased inflammatory cell infiltrations, epithelial cell shedding, and mucus secretion.³⁵ Hematological analysis showed significant elevations in white blood cells, neutrophils, lymphocytes, eosinophils, and platelets. Compared to controls, the Dibutyl Phthalate group exhibited a significant increase in inflammatory cells in rat BAL fluid, accompanied by elevated Th2 cytokines. Oxidative stress was evident, and increased apoptosis was observed in lung tissue.

The NOAEL for male and female rats that received up to 3170 mg/kg/d (5%) Diethyl Phthalate in feed for up to 16 wk was 150 mg/kg/d (0.2%).⁶ Male and female rats that received 5% Diethyl Phthalate for 16 wk had statistically significantly lower absolute weights of the brain, heart, spleen & kidneys than controls; reductions in the absolute weights of the female gonads and male heart, spleen, and kidneys were also observed in the 5% groups after 2 and 6 wk of treatment. Increased kidney weights were seen in males & lower pituitary weights in females in the 1% group after 2 wk of treatment. Increased gonad weights were also seen in females of the 1% group after 6 wk. No mortalities were observed in a 4-wk study in male rats that received Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate, the respective monoesters, or phthalic acid in corn oil via gavage (500 mg/kg bw/d for diesters, 250 mg/kg bw/d for monoester and acid).³⁶ Livers weights were significantly increased in groups treated with Dibutyl Phthalate and monobutyl phthalate compared to control group. Red blood cells and hematocrit were significantly lower and mean corpuscular hemoglobin concentration and platelet count were significantly higher in the Dibutyl Phthalate group (the monoester also had significantly higher platelet counts), and hemoglobin level was reduced only in the Dimethyl Phthalate group.

Inhalation

No significant effects were observed in male rats exposed to 1.5 mg/m³ of Dibutyl Phthalate vapor for 6 h/d and 6 d/wk for approximately 1 mo.² In another 1 mo study, no significant treatment-related effects were observed in rats exposed to up to 500 mg/m³ aerosolized Dibutyl Phthalate for 6 h/d and 5 d/wk.⁸ Particle size of the aerosolized Dibutyl Phthalate ranged from 1.5 to 2.0 μ m. The no-observed-effect concentration (NOEC) for inhalation exposure to aerosolized Dibutyl Phthalate was the highest concentration tested in this study, 500 mg/m³.

Other Routes

No paralysis or other adverse effects were observed in rabbits that received 2 ml/kg/d Diethyl Phthalate intraperitoneally for 8 d.² The intraperitoneal administration of 1.5 ml/kg/d Diethyl Phthalate to guinea pigs for 8 d did not result in any permanent ill effects during or after the experiment. An emulsion of Diethyl Phthalate administered intraperitoneally at 125 mg/kg/d for 6 wk to mice produced a slight decrease in weight gain and some evidence of peritonitis. No further adverse effects were observed.

Subchronic Toxicity Studies

Dermal

In a 90-d study, 0.5, 1, 2, or 4 ml/kg/d Dibutyl Phthalate or Dimethyl Phthalate were applied to clipped, intact skin over approximately 10% of the body surface.² For these ingredients, the subchronic dermal LD₅₀ was > 4 ml/kg/d. Dibutyl Phthalate was slightly irritating to skin and very irritating to rabbit penile mucosa. Slight dermatitis was observed, and in the 4 ml/kg dosed rabbits, slight renal damage (not further described) was observed. No dermal irritation or dermatitis was observed following application of Dimethyl Phthalate, although irritation was observed in the penile mucosa. Pulmonary edema and slight renal damage were observed in the rabbits that died during the study. Survivors had varying degrees of nephritis (not further described) at the two highest doses.

Oral

In a 3-mo study, rats received 0.12 or 1.2 g/kg/d Dibutyl Phthalate in olive oil.² Both doses produced a statistically significant increase in the rats mean liver weight; however, no histological evidence of any pathologic changes were found in the liver, kidneys, or spleen. No changes in behavioral patterns or clinical signs of toxicity were observed in rats that received 0.2, 1, or 5% Diethyl Phthalate in feed for up to 16 wk. Both sexes that received 5% and females that received 1% Diethyl Phthalate consumed less feed and gained less weight than controls. A reduction in absolute weight and an increase in relative weight of the brain, spleen, heart, kidneys, adrenal glands, gonads, and pituitary was observed in the 5% group. A pattern of increases in absolute and relative weights was observed in livers and various parts of the gastrointestinal tract in these rats. Both liver and kidneys were enlarged but histologically normal.

In another 3-mo dietary study, rats received 400, 2000, or 10,000 ppm (30, 152, or 752 mg/kg/d) Dibutyl Phthalate.⁸ No treatment-related effects were seen in the 400 and 2000 ppm groups. In the 10,000 ppm group the following observations were noted: decrease in triglyceride and triiodothyronine in both sexes; increase in cyano-intensive palmitoyl-CoA-oxidation in both sexes; decrease in number of red blood cells, hemoglobin, and hematocrit in the males; increase in glucose and albumin in the males; increased absolute and relative liver weights in both sexes; and decreased or missing lipid deposition in the hematocytes of both sexes.

The NTP conducted a 13-wk dietary study of Dibutyl Phthalate in mice and rats.⁸ The mice received 0, 1250, 2500, 5000, 10,000, or 20,000 ppm of the test material and rats received 0, 2500, 5000, 10,000, 20,000, or 40,000 ppm. All mice and rats survived to study completion. No clinical signs of toxicity attributed to Dibutyl Phthalate were observed in mice; however, male and female rats in the 40,000 ppm dose group were emaciated at the end of the 13-wk dosing period. No other clinical observations were reported in rats. In mice, reduced final body weights and body weight gains were observed starting at 5000 ppm in males and 20,000 ppm in females. Increased relative kidney weights were observed starting at 2500 ppm in females, and increased relative liver weights were observed started at 5000 ppm in both sexes. Female mice also had reduced hematocrit at 20,000 ppm. In male mice, increased serum testosterone and increased zinc in the testes were observed starting at 1250 and 5000 ppm, respectively. An increased incidence of cytoplasmic alteration in the liver was observed starting at 10,000 ppm in male mice and 20,000 ppm in female mice. In rats, reduced final body weights and body weight gains were observed starting at 10,000 ppm in males and 20,000 ppm in females. Reduced feed consumption was observed at 40,000 ppm in both sexes. Both sexes also had increased live and kidney weights (starting at 5000 ppm in males and 10,000 ppm in females), reduced cholesterol (starting at 20,000 ppm in both sexes), reduced triglycerides (starting at 2500 ppm in males and 10,000 ppm in females), increased palmitoyl-CoA oxidase (starting at 5000 ppm in both sexes), and increased incidence of cytoplasmic alteration in the liver (starting at 10,000 ppm in both sexes). Additionally, male rats had reduced hematocrit (starting at 20,000 ppm), reduced hemoglobin (starting at 5000 ppm), increased platelets (starting at 5000 ppm), increased incidence of germinal epithelium atrophy (starting at 10,000 ppm), and reduced serum testosterone, reduced testes weights, and reduced zinc in testes (starting at 20,000 ppm of all effects).

In a 120-d study, rats received 5% (v/v) ethanol, 50 ppm (w/v) Diethyl Phthalate, or 5% ethanol + 50 ppm Diethyl Phthalate in drinking water.⁸ No significant changes were seen in the body weights, liver weights, daily water consumption, and serum glucose levels of any treatment groups. Generally, Diethyl Phthalate caused changes in enzymes and biochemical endpoints in the liver and serum that were not affected by the presence of ethanol, thus there was no interaction between ethanol and Diethyl Phthalate at these doses, but Diethyl Phthalate alone appears to cause toxic injury to the liver under the study conditions.

Inhalation

Rats exposed to 0.5 mg/m³ and 50 mg/m³ of Dibutyl Phthalate mist for 6 h/d for 6 mo had reduced weight gains and increased brain and lung weights than control rats.² The higher concentration had a greater effect than the lower concentration.

Chronic Toxicity Studies

Oral

No adverse effects were observed in rats that received 2.5 mg/kg/d Dibutyl Phthalate for 6 mo.² In a year-long study, groups of 20 male and 20 female rats received 0 or 0.125% Dibutyl Phthalate in feed. Six rats died before study end, but no specific cause of death was determined. No remarkable alterations were observed upon gross and histological examination

of the liver, kidneys or spleen. In another year-long study, groups of 10 rats received 0, 0.01, 0.05, 0.25, or 1.25% Dibutyl Phthalate in feed. No effect of growth or survival was observed in rats that received up to 0.25% test material in their feed. At 1.25%, half of the rats died the first week of the study. The remaining rats gained weight comparable to the control group. No rats exhibited significant changes in the number or distribution of "elements" in the peripheral blood or specific gross pathological changes. No hematological effects on hematological parameters, organ weights, or other pathological changes were observed in an 18-mo study in rats that received 1ml/kg Dibutyl Phthalate in oil (type not specified) 2 times/wk. In a 2-yr study, groups of 10 female rats received 2, 4, or 8% Dimethyl Phthalate in feed. Mortality rates were similar to control rats, and no effects were observed in the 2% dose group. A slight but significant effect was observed on growth in the 4 and 8% dose groups. Chronic nephritis was observed in rats in the 8% group.

Other Routes

A series of doses of Dibutyl Phthalate, Diethyl Phthalate, or Dimethyl Phthalate were injected intraperitoneally into groups of male mice 5 d/wk.² The apparent LD₅₀ was calculated each week until it remained constant for 3 wk; this was the chronic LD₅₀. Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate reached chronic LD₅₀s in 25, 14, and 18 wk, respectively. The chronic LD₅₀ values for Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were 0.85, 1.39, and 1.18 ml/kg/d, respectively.

DEVELOPMENTAL AND REPRODUCTIVE TOXICITY STUDIES

The numerous developmental and reproductive toxicity studies from the original 1985 report and the data from the 2005 re-review have been summarized in Table 6. Studies available since these documents were published are summarized in Table 7, and extremely brief overviews of these studies are presented below.

Embryotoxic and Teratogenic Effects

The NOEC for malformations was 5.8 ppm for Dibutyl Phthalate (tested at up to 15 ppm) and 17 ppm for Diethyl Phthalate (tested at up to 200 ppm) in in vitro 96-h frog embryo teratogenesis assays using *Xenopus laevis* embryos.³⁷ The mean 96-h LC₅₀s for Dibutyl Phthalate and Diethyl Phthalate in this study were 12.88 and 64.5 ppm, respectively. Malformations were observed starting at 5 ppm in Dibutyl Phthalate-treated embryos and starting at 50 ppm in Diethyl Phthalate-treated embryos. Additional studies in *Xenopus laevis* with Dibutyl Phthalate also noted defects, with mean percent normal development decreased as concentration and exposure time increased.^{38,39} No significant effects on survival or hatching of zebrafish embryos were observed following exposure to up to 250 µg/l Dibutyl Phthalate; however, body length decrease, yolk sac abnormalities, and immune responses were observed.⁴⁰ An acute developmental toxicity of Dibutyl Phthalate (up to 0.2 mg/l), Diethyl Phthalate (up to 10 mg/l), and Dimethyl Phthalate (up to 8 mg/l) in zebrafish saw a malformation rate of 30% with 0.05 mg/l Dibutyl Phthalate (deaths were observed at higher concentrations).⁴¹ Dimethyl Phthalate was not correlated with larvae mortality. In a dermal study in mice, Diethyl Phthalate (tested at up to 5600 mg/kg/d) applied to pregnant mice on gestation days 0-17 had a NOAEL < 500 mg/kg/d in dams and 1600 mg/kg/d in offspring.⁶ The dams exhibited dose-related abnormal behavior and reduced thymus and spleen weights were observed at all doses; while significantly reduced fetal body weight and an increased incidence of cervical and lumbar rib variations were observed in 5600 mg/kg/d offspring. A 3-generation reproductive and developmental study in mice with Dibutyl Phthalate in olive oil up to 2000 mg/kg bw induced skeletal malformations in F₀ males, with significant mortality in postnatal life and skewing the sex ratio.⁴² An increased frequency of DNA damage in the germ cells was observed in F₁ males; exposure of F₀ males to Dibutyl Phthalate did not affect F₁ male fertility and pregnancy frequency. In multigeneration studies with Dibutyl Phthalate in rats at up to 10,000 ppm in dietary feed, no adverse effects to body weights and body weight gains were observed in parental dams.^{43,44} In one study, male offspring in the 10,000 ppm dose group had high incidence of small or absent reproductive organs and undescended testes, gross lesions correlated with microscopic lesions in the male reproductive organs.⁴⁴ In another feed study, anogenital distance was significantly decreased in male F₁ and F₂ offspring in the 10,000 ppm dose groups.⁴³ F₁ males of the 10,000 ppm dose group also had a statistically significant mean delay in preputial separation, testicular descent, and seminiferous tubular atrophy. In a 2-generation rat feed study with Diethyl Phthalate at up to 15,000 ppm, the NOAEL for parental animals was 15,000 ppm, and the NOAEL for development and growth of pups was 3000 ppm.^{6,45} Body weight gains before weaning were inhibited in F₁ and F₂ pups and vaginal openings were slightly delayed in F₁ females at 15,000 ppm.

Effects in Female Reproductive Organs

Mouse ovarian antral follicles treated with 1000 µg/ml Dibutyl Phthalate in DMSO for 24 h were significantly smaller than controls, and follicles treated with concentrations greater than 100 µg/ml were significantly reduced after 72 h.⁴⁶ No toxicity was observed to the follicles after treatment with monobutyl phthalate at the same concentrations for the same periods of time. Dibutyl Phthalate in tocopherol-stripped corn oil did not cause significant deviations in the expression of transcripts encoding insulin-like growth factor 1 (IGF) binding proteins in female mice that received up to 1000 mg/kg/d for 20 d when compared to the vehicle control.⁴⁷ Total number of ovarian follicles counted per ovary was reduced in the 100 µg/kg/d group when compared to controls; however no significant differences were observed in the 10 µg/kg/d or 1000 mg/kg/d dose groups. Significantly low primordial follicle counts observed in the 100 µg/kg/d group and in the 1000 mg/kg/d group compared to controls, also fewer primary follicles were observed in 100 µg/kg/d mice. In another female mouse study with Dibutyl Phthalate in tocopherol-stripped corn oil (up to 1000 µg/kg/d for 30 d), dose-dependent effects on

folliculogenesis and gene expression were observed.⁴⁸ At 1000 µg/kg, more atretic follicles were observed in the ovaries. In a dose-dependent manner, Dibutyl Phthalate significantly reduced the expression of genes responsible for homologous recombination, mismatch repair, and nucleotide excision repair. Pregnant rats that received 500 mg/kg Dibutyl Phthalate in corn oil every second day starting at gestation day 14.5 through postnatal day 6 exhibited no adverse effects and delivered exclusively live pups.⁴⁹ Offspring also exhibited no adverse effects, and female pups had no effects to anogenital distance or adverse effects in vaginal opening. In an oral uterotrophic assay, female rats that received 100 mg/kg Dibutyl Phthalate in corn oil from postnatal day 21 for 3 d had significantly decrease uterine wet weight and minor variations in ovary wet weight; no abnormal clinical signs or symptoms were observed in any of the treated or control animals.⁵⁰ Female rats from a pubertal onset assay that were treated with 10 or 100 mg/kg Dibutyl Phthalate from postnatal day 21 for 20 d had significantly reduced uterus and ovary weights in both treated groups. However, vaginal opening was not observed in any of the animals in the controls and treated groups until postnatal day 42, except in one animal each in vehicle control and the 100 mg/kg dose group. In an oral multigeneration study in which the F₀ pregnant mice received a mixture of phthalates that included 15% Dibutyl Phthalate and 35% Diethyl Phthalate in tocopherol-stripped corn oil at up to 500 mg/kg/d, female offspring of the F₂ generation had increased uterine weight, anogenital distance and body weight.⁵¹ Cystic ovaries and breeding and pregnancy complications were also observed. Similar effects were observed in the females of the F₃ generation.

Effects in Male Reproductive Organs

In an in vitro mechanistic study using TM3 (Leydig) and TM4 (Sertoli) cells, exposure to Dibutyl Phthalate at 0, 5, 10, 50, and 100 mg/l for 24 h produced dose-dependent evidence of ferroptosis.⁵² Dibutyl Phthalate treatment decreased cellular GSH and mitochondrial membrane potential, and increased malondialdehyde (MDA), reactive oxygen species (ROS), oxidized glutathione (GSSG), and ferrous iron (Fe²⁺) levels. Dibutyl Phthalate selectively upregulated peroxiredoxin-6 (PRDX6), a negative regulator of ferroptosis, while glutathione peroxidase-4 (GPX4) remained unchanged. Further, Dibutyl Phthalate upregulated SP1 expression, which can directly bound to the PRDX6 promoter and transcriptionally activated its expression. Rat testicular explants exposed in vitro to Dibutyl Phthalate at 10⁻⁶, 10⁻⁵, and 10⁻⁴ M for 24 h showed significant upregulation of Notch1, Dll4, and Hey1 at both mRNA and protein levels, while Hes1 expression remained unchanged.⁵³ Dibutyl Phthalate altered immunoexpression of activated NOTCH1, DLL4, HEY1 and HES5 both in seminiferous epithelium and interstitial tissue, with differential effects across cell types. In rat Leydig cells exposed to 50 mg/l Dibutyl Phthalate, Dibutyl Phthalate + prostaglandin E2, or flutamide for 24 h, expression of testosterone was significantly decreased in all groups.⁵⁴ A significant decrease of Cx43 was observed in the Dibutyl Phthalate group. Human sperm exposed to up to 134.7 µg/ml Dibutyl Phthalate for 30 min to up to 96 h had a concentration- and duration-dependent decrease in motility.⁵⁵ In another study with human sperm, Dibutyl Phthalate at 6 µM, monobutyl phthalate at 3 µM, and a mixture of both had adverse effects on sperm motility, penetration ability, and capacitation following exposure for 1 to 4 h.⁵⁶

Multiple oral studies have investigated the effects of Dibutyl Phthalate in the male rodents.^{6,36,44,49,52,57-68} The NOAEL was 50 mg/kg bw/d for developmental toxicity in a study in pregnant rats that received up to 500 mg/kg bw/d on gestation day 1 through postnatal day 21.⁵⁷ At doses of 250 mg/kg and higher, reduced birth weight, body weight gain, number of live pups per litter, anogenital distance in males, epididymis weight, and sperm count and motility were observed. Studies in which pregnant rats received up to 600 mg/kg bw/d Dibutyl Phthalate between gestation days 12-21 found cryptorchidism, infertility, hypospadias, decreases in anogenital distance, and/or testes abnormalities.^{49,59,61-67} Closer examination of the focal dysgenesis of the testes found adverse effects in the Leydig and Sertoli cells. One study of rats that received Dibutyl Phthalate (500 mg/kg/d) from gestation day 12 through gestation days 16 - 20 saw a significantly increased incidence of multinucleated gonocytes (MNG) in fetal rat testes following in utero exposure to the test material.⁶⁵ Sertoli cells exhibited retracted apical processes, disorganized vimentin cytoskeleton, and abnormal contacts with gonocytes. These morphological changes were no longer evident after birth and cessation of exposure. Pregnant dams that received Dibutyl Phthalate (up to 600 mg/kg/d) or Diethyl Phthalate (up to 900 mg/kg/d) on gestation days 8 - 18 experienced no adverse effects, but fetal testicular testosterone production was significantly reduced at doses of 300 mg/kg/d or higher in Dibutyl Phthalate male fetuses.⁶⁸ This effect was not observed with Diethyl Phthalate. A phthalate mixture containing Dibutyl Phthalate significantly increased fetal mortality at 40% of top dose (containing 120 mg/kg/d Dibutyl Phthalate) and above in dams that received the test material on gestations days 8 - 18, and testosterone production was reduced in a dose-additive manner starting at 20% of top dose (containing 60 mg/kg/d Dibutyl Phthalate) and above.⁶⁸ Phthalate-induced underdevelopment of the testes was observed at 100% of the top dose (300 mg/kg/d). Testes weights were significantly reduced in male rats that received Dibutyl Phthalate (500 mg/kg bw/d) for 4 wk, and Dibutyl Phthalate, monobutyl phthalate (250 mg/kg bw/d), and monoethyl phthalate (250 mg/kg bw/d) significantly lowered sperm counts and sperm motility of epididymal sperm. These effects were not observed with Diethyl Phthalate or Dimethyl Phthalate (500 mg/kg bw/d for each diester), monomethyl phthalate (250 mg/kg bw/d), or phthalic acid (250 mg/kg bw/d).³⁶

Adult male rats that received up to 600 mg/kg/d Dibutyl Phthalate in corn oil for 15 d had significantly decreased testicular weight at each dose tested, and sperm count and motility were significantly decreased in a dose-dependent manner.⁵⁸ Serum follicle-stimulating hormone, testosterone levels, and testicular lactate dehydrogenase activities were significantly decreased at all doses tested. In a study of male rats exposed to 250 mg/kg Dibutyl Phthalate, with or without BaP for up to 12 wk, no adverse effects were observed in the relative weights of the testes and epididymides, but vacuolization of Sertoli cells was observed in Dibutyl Phthalate-exposed rats after 12 wk.⁶⁰ Male rats orally exposed to

Dibutyl Phthalate (500 mg/kg/d) for 3 wk, had decreased germ cell layer thickness, and reduced sperm density in testicular tissue.⁵² These morphological changes were associated with increased testicular levels of MDA, Fe²⁺, and GSSG, together with decreased reduced GSH and serum testosterone.⁵² No gross lesion were observed in mice that received up to 1393 mg/kg bw/d Dibutyl Phthalate in a 2-yr study, but significantly increased incidences of germinal epithelium degeneration in the testes and exfoliated germ cells in the epididymal duct were observed.⁴⁴ In rats that received 2000 mg/kg/d Diethyl Phthalate in dietary feed for 7 d, lower levels of testosterone were measured in the testes and serum of the treated animals, but no testicular damage was observed by microscopic examination.⁶

Mice that received Dibutyl Phthalate (2.5 mg/kg/d), with or without diethylhexyl phthalate, for 40 d via implanted osmotic pumps had abnormal sperm morphology, particularly in the mixture group.⁶⁹ Significant differences in sperm concentration were observed between control and exposed groups. No clinical signs of toxicity were observed in rats that were injected subcutaneously with Dibutyl Phthalate (up to 20 mg/animal) from postnatal day 5 -14; and no alterations in testicular descent were observed in any treated rats. However, significantly reduced testes, seminal vesicles, levator ani plus bulbocavernosus muscles (LABC), and Cowper's glands weights were observed in the 20 mg dose group when compared to the controls.⁷⁰

In additional studies, 100 mg/kg Dibutyl Phthalate injected into chicken eggs significantly reduced hatching and increased late hatching when compared to controls.⁷¹ Gross malfunctions and severe moto dysfunctions were observed in treated animals. Zebrafish that received up to 1133 µg/l Dibutyl Phthalate, with or without diisobutyl phthalate had a greater imbalance of the testosterone to estradiol ratio and severe structural damage from the mixture than to just Dibutyl Phthalate.⁷² These effects were consistent with the testis transcriptome analysis for which 4570 genes were differentially expressed in the mixture exposure, while 2795 genes were differentially expressed in Dibutyl Phthalate.

GENOTOXICITY STUDIES

In Vitro

Dibutyl Phthalate was mainly not mutagenic in multiple Ames tests, with and without metabolic activation in several Salmonella typhimurium strains.² In additional studies, Dibutyl Phthalate (up to 2000 µg/plate with and up to 500µg/plate without metabolic activation) was mutagenic in S. typhimurium strain TA100 without activation at all doses tested, but mutagenicity was not observed in strains TA98, TA1535, TA1537, TA1538, or TA2637, with or without metabolic activation or in strain TA100 with activation.⁸ In the same study using the same concentrations of Diethyl Phthalate and Dimethyl Phthalate, mutagenicity was observed with both ingredients in strains TA100 and TA1535 without metabolic activation only. Diethyl Phthalate and Dimethyl Phthalates were mutagenic in Ames tests conducted with S. typhimurium strain TA100.² In a different Ames test, Dimethyl Phthalate (500 - 4000 µg/plate) was mutagenic in S. typhimurium strain TA100 in a dose-dependent manner without metabolic activation.⁸ Monobutyl phthalate (up to 4000 µg/plate) was not mutagenic to S. typhimurium strain TA100 without metabolic activation. No mutagenicity was reported in Ames tests performed by the NTP on Diethyl Phthalate (up to 10,000 µg/plate) and Dimethyl Phthalate (up to 6666 µg/plate), with and without metabolic activation. Dibutyl Phthalate and Diethyl Phthalate were not mutagenic in studies with Escherichia coli.² Dibutyl Phthalate had no mutagenic effect on the yeast Saccharomyces cerevisiae, with or without metabolic activation, in a reversion analysis.

Dibutyl Phthalate was negative for chromosome aberrations and sister chromatid exchanges (SCEs) in Chinese hamster cells in one study. In another study, the mitotic index was not appreciably decreased when the Chinese hamster cells were exposed to Dibutyl Phthalate in ethanol; however, a significant increase over the vehicle for the number of SCEs was found. No dose-response effect from Dibutyl Phthalate was observed, though. In a third study, Dibutyl Phthalate in a 0.2% bovine albumin solution led researchers to call the test material "a suspicious compound" even though the results did not conclusively prove that Dibutyl Phthalate caused chromosome aberrations. Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate (0.25 mg/ml) had no effect on chromatid aberrations in human leukocyte cultures when compared to controls.

Diethyl Phthalate (up to 10,000µg/plate) and Dimethyl Phthalate (up to 6666µg/plate) induced SCEs in Chinese hamster ovary (CHO) cells with metabolic activation, but not without.⁸ Neither of these ingredients induced chromosomal aberrations in CHO cells. Dibutyl Phthalate (up to 0.06 µl/ml) and Dimethyl Phthalate (up to 0.6 µl/ml) were found to be mutagenic in a L5178Y mouse lymphoma assay with metabolic activation, but not without. The genotoxicity of Dibutyl Phthalate (354 µmol/ml) was evaluated human mucosal cells from the upper respiratory tract. An increase in single strand breaks in DNA was observed, with cells of the nasal mucosa having more sensitivity than oropharyngeal epithelia.

Genotoxicity studies are summarized in Table 8. Dibutyl Phthalate (up to 2000 µg/plate) and Diethyl Phthalate (up to 5000 µg/plate) were not genotoxic in separate Ames tests, with and without metabolic activation.^{5,6} In a third Ames test, Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were mildly genotoxic in *S. typhimurium* strains TA100 and TA1535 without metabolic activation in a dose-dependent manner when tested at up to 2000 µg/plate.⁷³ Dibutyl Phthalate (up to 80 µM) was genotoxic in micronucleus tests using cultured bovine lymphocytes and CHO cells.^{74,75} In chromosome aberrations tests, Dibutyl Phthalate (up to 30 µM) caused a higher occurrence of chromosomal aberrations in CHO cells at all tested concentrations when compared to the vehicle control; however, Diethyl Phthalate (up to 1780 µg/ml) was not genotoxic to human lymphocytes.^{6,75} Diethyl Phthalate (up to 771 µg/ml) was also not genotoxic in a gene mutation assay in

mouse L5178 TK +/- lymphoma cells.⁶ Genotoxicity was observed to Dibutyl Phthalate when tested in human mucosal cells and human peripheral lymphocytes at up to 354 $\mu\text{mol/ml}$ and in cultured bovine lymphocytes at up to 80 μM .^{74,76}

CARCINOGENICITY STUDIES

Dermal

In a study performed by the NTP, mice were dermally exposed to 0, 9, 18, or 37 μg Diethyl Phthalate in 100 μl acetone 5 d/wk for 103 wk.⁸ The 2-yr survivability of treated mice was similar to that of controls. There were no adverse clinical signs and no signs of dermatotoxicity. An increase in liver neoplasms was observed in male and female treated mice. The incidence of adenoma or carcinoma (combined) was increased in 37 μg group males and in the 9 and 15 μg group females, but not in the 37 μg group females. The increased incidence of liver neoplasms was considered equivocal because the incidence of hepatic neoplasms was within the historical range and because there was no clear dose-response relationship in females.

A similar study was performed in rats with 0, 123, or 368 μg Diethyl Phthalate (neat).⁸ Survival of treated rats was similar to the controls for the first 15 mo; however, survival to 2 yr was reduced in male rats in all groups, including the controls. The incidence of anterior pituitary adenoma was higher in all groups of male and female rats than the historical incidence. This may have been partially responsible for the reduced survivability of the males. No adverse clinical effects or evidence of carcinogenic activity from treatment with Diethyl Phthalate were observed in male or female rats.

Oral

Carcinogenesis was not observed in rats that were fed Dibutyl Phthalate for 18-mo or longer.²

The carcinogenicity potential of Dibutyl Phthalate was evaluated in a 2-yr study in mice and rats performed by the NTP.⁴⁴ Groups of 50 male and 50 female B6C3F1/N mice received 0, 1000, 3000, or 10,000 ppm Dibutyl Phthalate in feed. The estimated average chronic chemical consumption was 105 - 112, 329 - 347, and 1306 - 1393 mg/kg bw/d Dibutyl Phthalate for the 1000, 3000, and 10,000 ppm dose groups, respectively. Feed consumption and body weights were measured and clinical observations were made through the study period. Complete necropsies and microscopic examinations were performed on all mice at the end of the study. No exposure-related effects on survival were observed. Mean body weights were lower only in the 10,000 ppm groups when compared to the control groups. Feed consumption was comparable to the controls. There was no exposure-related increase in neoplasms. Effects to the male reproductive system are summarized above in the Developmental and Reproductive Toxicity Studies section and in Table 7. Nonneoplastic lesions observed generally occurred only in the 10,000 ppm group. These included hepatocyte cytoplasmic alteration in male and female mice, multinucleated hepatocytes in male mice, and renal tubule hyperplasia in female mice. The authors of the study concluded that there is no evidence of carcinogenic activity of Dibutyl Phthalate in male or female mice at up to 10,000 ppm.

In the rat study, groups of 45 - 47 time-mated female Sprague Dawley rats were exposed to 0, 300, 1000, 3000, or 10,000 ppm Dibutyl Phthalate in feed starting on gestation day 6 and continuing through lactation.⁴⁴ Body weights and feed consumption were measured and clinical observations were made, and complete necropsies and microscopic examinations were performed on all F₁ rats at the end of the study. Post-weaning, groups of 50 male and 50 female F₁ offspring consumed diets with the same exposure concentrations as the dams for 2 yr. The estimated average chronic chemical consumption was 16 - 17, 54 - 57, 152 - 169, and 510 - 600 mg/kg bw/d Dibutyl Phthalate for the 300, 1000, 3000, and 10,000 ppm dose groups, respectively. No exposure-related effect in mortality between exposed and control groups was observed through the study period. Reproductive effects are summarized above in the Developmental and Reproductive Toxicity Studies section and in Table 7.

In the F₁ postweaning animals, body weights were approximately within 20% of the control animals. At 2 yr in male rats, the incidence of pancreatic acinus adenomas was slightly higher in the 10,000 ppm group compared to the control group. The marginal increase may be related to the test material as pancreatic acinus adenomas and carcinomas have been associated with peroxisome proliferator-activated receptor alpha activation, which has been observed with exposure to other phthalates. Again, nonneoplastic lesion observed generally occurred only in the 10,000 ppm group. These included hepatocyte cytoplasmic alteration in male and female rats and pars distalis hypertrophy in male rats. The authors of the study concluded that there was equivocal evidence of carcinogenic activity of Dibutyl Phthalate in male rats based on the marginal increased in the incidence of pancreatic acinus adenomas at up to 10,000 ppm; however, there was no evidence of carcinogenic activity of Dibutyl Phthalate at up to the same concentration in female rats.

Tumor Promotion

Dermal

In a 1-yr initiation/promotion study in male mice, Diethyl Phthalate and Dimethyl Phthalate were tested as tumor initiators, with and without promotion from 12-O-tetradecanoylphorbol-13-acetate.⁸ Diethyl Phthalate and Dimethyl Phthalate were also tested for the potential as promoters, with and without initiation from 7,12-diemthylbenzanthracene. Neither Diethyl Phthalate nor Dimethyl Phthalate initiated or promoted skin carcinogenesis.

In Vitro Cell Transformation

Normal human breast epithelial cells (MCF-10A) were co-cultured with fibroblasts derived from normal mammary tissue adjacent to estrogen receptor-positive primary breast cancers and exposed to Dibutyl Phthalate (100 nM).⁷⁷ Untreated MCF-10A and MCF-10A co-cultured with fibroblasts alone served as controls. Cell viability was assessed using a MTT assay, cell cycle distribution by flow cytometry, and the protein expression of cell cycle regulators and P13K/AKT/mTOR signaling components by Western blot analysis. The co-cultured cells treated with Dibutyl Phthalate exhibited a significant increase in cell viability when compared to the untreated control ($p < 0.001$) and to the co-cultured control ($p < 0.05$). PDK1 expression was significantly higher in the MCF-10A cells treated with Dibutyl Phthalate compared with the control cells ($p < 0.05$) and the co-cultured cells ($p < 0.05$). Significant higher expressions of P13K, p-AKT, and p-mTOR were observed in the MCF-10A cells treated with Dibutyl Phthalate ($p < 0.05$). Dibutyl Phthalate significantly increased cell percentages in the S and G2/M phases compared to the untreated controls ($p < 0.001$) and the co-cultured cells ($p < 0.05$). Elevated expression of cyclin/CDK complexes (cyclin D/DCK4, cyclin E/CDK2, cyclin A/CDK2, cyclin /CDK1, and cyclin B/CDK1) was also observed. The authors concluded that Dibutyl Phthalate may have a role in the development of estrogen receptor positive breast cancer.

In a related study by the same research group, MCF-10A normal breast cells were treated with Dibutyl Phthalate (10 or 100 nM) or 17 β -estradiol (10 nM) in co-culture with mammary fibroblasts.⁷⁸ Compared to controls, both Dibutyl Phthalate and estradiol significantly increased cell viability, decreased apoptosis (TUNEL assay), and increased cell numbers in the S and G2/M phases. Western blot analysis showed dose-dependent up-regulation of cyclins/CDKs, PDK1, P13K, p-AKT, p-mTOR, and BCL-2, with concomitant down-regulation of Bax protein, cytochrome C, caspase 8, and caspase 3.

The tumor promotion potential of Dibutyl Phthalate (up to 10 μ M) in DMSO was studied in normal bladder epithelial cells (SV-HUC-1) and human bladder cancer cells (T24 and UM-UC-3).⁷⁹ The 3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide (MTT) and Transwell assays were used to investigate cell proliferation and migration (cell motility through a porous membrane), respectively, and second-generation sequencing was used to determine any differences in gene expression before and after treatment with Dibutyl Phthalate. Differential gene expression was verified by quantitative polymerase chain reaction (q-PCR) and analyzed using bioinformatics. Dibutyl Phthalate exposure reduced migration of SV-UC-1 cells but enhanced migration of T24 and UC-3 cells in a concentration-dependent manner. This indicated that Dibutyl Phthalate enhanced migration in bladder cancer cells. FOSB, JUND, ATP6V1C2, and RHOQ genes were identified as differentially expressed following Dibutyl Phthalate treatment; q-PCR confirmed FOSB upregulation, and FOSB overexpression promoted both proliferation and invasion in bladder cancer cells.

In human breast epithelial R2d cells, exposure to Dibutyl Phthalate (1 μ M) for 24 h induced expression of mesenchymal markers vimentin and CD90 and decreased expression of epithelial markers cytokeratin 7 and E-cadherin.⁸⁰ After 72 h, the cells began to resemble mesenchymal-like morphological traits, such as spindle-like shape and scattered colony growth.

Analysis of MCF-7 and MDA-MB-231 breast cancer stem cells after exposure to Dibutyl Phthalate (10^{-9} M) revealed enhanced tumorsphere formation ability and higher levels of cancer stem cell markers (CD133, CD44, ALDH1A1, OCT-4, Nanog).⁸¹ The promotion of this stemness is proposed to be mediated by the Sonic hedgehog (Shh) signaling pathway and Δ Np63 α upregulation.

ANTI-CARCINOGENICITY STUDIES

Dibutyl Phthalate (50 μ g/ml) suppressed the growth of HL-60 cells.⁸ Human bone marrow granulocyte/macrophage progenitor and human myeloid leukemia assays were used to show that Dibutyl Phthalate has no effect on normal bone marrow cell growth, but growth of leukemic cells was dose- and time-dependently reduced in vitro. In another study, Dibutyl Phthalate (50 μ M) inhibited the incorporation of 3H-thymidine into the DNA of Ehrlich ascites tumor cells and inhibited cell proliferation. Diethyl Phthalate and Dimethyl Phthalate did not show these inhibitory effects. The ability of Dibutyl Phthalate to suppress leukemic cell growth may work through an apoptotic mechanism.

OTHER RELEVANT STUDIES

Effects on Gene Expression

The altered genetic expressions profiles were studied in fetal male rats exposed to 500 mg/kg/d Dibutyl Phthalate via daily gavage on gestation days 12 to 21.⁸ Corn oil was used as a negative control, and the antiandrogen flutamide (50 mg/kg/d) was administered as a positive control. Sample size was 18 dams per exposure condition. Testes of fetal males were collected and examined by microarray analysis on gestation days 16, 19, and 21 ($n = 6$ litters per group per examination day). Selected genes were quantified by reverse transcription-polymerase chain reaction (RT-PCR). Dibutyl Phthalate reduced the expression of the steroidogenic enzymes cytochrome P450c17 and steroidogenic acute regulatory protein. Testicular testosterone and androstenedione were decreased on gestation days 19 and 20, while progesterone was increased on gestation day 19 in rats of the Dibutyl Phthalate group. Dibutyl Phthalate exposure also caused an increase in the expression of the cell survival proteins testosterone-repressed prostate messenger-2 and bcl-2, which may be involved in

Dibutyl Phthalate-induced Leydig cell hyperplasia. A reduction in stem cell factor receptor in the Dibutyl Phthalate group may contribute to gonocyte degeneration.

In an ecotoxicological study, zebrafish were exposed to up to 2 mg/l Dibutyl Phthalate for up to 28 d.⁸² Feeding of the fish stopped for 24 h before the experiment and the fish were randomly selected on days 7, 14, 21, and 28 to determine various indicators, including antioxidant enzymes and gene expression. Oxidative stress, lipid peroxidation, and DNA damage occurred in zebrafish liver according to changes in antioxidant enzymes, MDA, and 8-hydroxydeoxyguanosine (8-OHdG) content. Acetylcholinesterase (AChE) activity was always active and negatively correlated with the Dibutyl Phthalate concentration. The expression of Cu/Zn-sod and gpx genes were similar to that of antioxidant enzymes from 7 to 21 d, while in the end, the inconsistent result appeared due to the time lag effect in protein modification, gene transcription and translation. The mRNA abundance of Caspase-3 and p53 were upregulated, showing a dose-response relationship. The integrated biomarker reaction indicated that the effects of exposure time on zebrafish liver was day 14 > day 28 > day 7 > day 21.

In another study used to assess gene expression, DNA damage, and oxidative stress biomarkers, zebrafish were exposed to Dibutyl Phthalate (up to 2 mg/l) for 7, 14, 21, or 28 d.⁸³ Dibutyl Phthalate significantly stimulated SOD and CAT activities, increasing MDA and 8-OHdG contents. On the day 28, AChE inhibition rates for 0.08, 0.4, and 2 mg/l were 13.4, 11.9, and 14.7%. The trend of Cu/Zn-sod gene variation was consistent with SOD activity, showing “inhibition-activation-inhibition”. Integrated biomarker response values showed a dose-response relationship on day 28.

CD-1 female mice were used in a study of potential alterations by Dibutyl Phthalate (99.6% pure) on metabolizing enzymes in the ovary and liver after single or repeated doses of the test material via gavage.²⁹ Groups of 5 mice/treatment mice received 0, 1, 10, or 1000 mg/kg Dibutyl Phthalate in tocopherol-stripped corn oil either once or once daily for 10 d. The mice were killed a 2, 6, 12, or 24 h after the single dose or the final dose, and tissues, including liver and ovaries, and serum were collected to extract RNA to generate cDNA for enzyme analysis. Hepatic and ovarian expression of *Lpl*, *ldhl1a1*, *Adhl*, *Ugt1a6a*, and *Cyp1b1* were altered in the treated mice in a time- and dose-specific manner.

Epigenetic Effects

The effects of Dibutyl Phthalate and its metabolite, monobutyl phthalate, on epigenetic parameters were studied in human peripheral blood mononuclear cells.⁸⁴ The cells were incubated with each test material at 0.001, 0.01, or 0.1 µg/ml for 24 h. Methylation in the promoter regions of tumor suppressor genes (P16, TP53) and proto-oncogenes (BCL2, CCND1) were assessed and the expression profile of these genes was analyzed. When compared to controls, a statistically significant reduction of global DNA methylation level was observed in the cells exposed to monobutyl phthalate at 0.001 and 0.01 µg/ml, but no statistically significant changes were observed for Dibutyl Phthalate. For TP53 gene promoter, statistically significant hypermethylation of monobutyl phthalate was observed at all concentrations tested, but for Dibutyl Phthalate, an increase was only observed at 0.001 µg/ml. With P16 gene promoter, Dibutyl Phthalate and its metabolite caused hypermethylation at the 2 highest concentrations. No significant changes were observed in methylation within the promoter of the CCND1 gene with monobutyl phthalate, but Dibutyl Phthalate caused statistically significant hypomethylation within this gene. Both forms of the phthalates caused a decrease in methylation within BCL2 gene promoter at 0.01 and 0.1 µg/ml. A statistically significant decrease in the expression of TP53 and P16 genes was found with monobutyl phthalate at all concentrations and with Dibutyl Phthalate at 0.01 (for TP53 only) and 0.1 µg/ml (for both genes). The greatest increase in CCND1 gene expression was observed with Dibutyl Phthalate at 0.001 and 0.01 µg/ml, while no changes were observed with monobutyl phthalate. An increase in BCL2 gene expression was observed with 0.1 µg/ml Dibutyl Phthalate, and Dibutyl Phthalate and its metabolite at 0.01 µg/ml caused a statistically significant increase in this examined parameter.

Endocrine Effects

In Vitro

The estrogenic effects of several phthalate esters were studied in a recombinant yeast assay and in estrogen-responsive human breast cancer cells; Dibutyl Phthalate and Diethyl Phthalate were shown to have extremely weak estrogenic activity.⁸ In another study, MCF-7 cells were transfected with human estrogen receptor construct associated with a luciferase reporter gene. The luciferase response of cells exposed to 10 µM Dibutyl Phthalate was 36% of the response to 10 nM 17-β-estradiol, suggesting a weak estrogenic activity.

*In recombinant yeast assay to determine the estrogenic potential of Diethyl Phthalate, a strain of *Saccharomyces cerevisiae* contained a DNA sequence for human estrogen receptor (hER) associated with the reporter gene lac-Z (encoding the enzyme β-galactosidase).⁸ Binding to and activation of the hER in this yeast assay results in an increase in the synthesis and secretion of β-galactosidase β-estradiol produced an estrogenic response with an EC₅₀ of 3.2×10^{-11} M. Spectrophotometry of the medium treated with Diethyl Phthalate at concentrations up to 10^{-3} M did not differ from that of negative control and untreated cultures.*

In a protein binding assay, several phthalates were evaluated for their potential for binding to sex hormone binding globulin (SHBG) in competition with radiolabeled H³-E₂. Dibutyl Phthalate and monobutyl phthalate (each at 200 µM) failed to displace appreciable amounts of H³-E₂, and thus, did not bind to SHBG. In a study evaluating the ability of phthalate esters to inhibit the binding of known (radiolabeled) ligands to endocrine receptors from healthy human uterus and

prostate tissues, Dibutyl Phthalate (100 μ M), monobutyl phthalate (10 μ M), Diethyl Phthalate (100 μ M), and monoethyl phthalate (10 μ M) did not significantly inhibit the binding of known ligands to the estrogen, progesterone, testosterone, or retinoic acid (uterus and prostate) receptors.

Dibutyl Phthalate exposure caused a slight increase in the number of progesterone receptors in rat uterine tissue, but the action was one-millionth the efficacy of β -estradiol.⁸ Dibutyl Phthalate at 1.0×10^{-3} M had weak binding to estrogen receptors in rat uterine tissue, however it showed no estrogenic activity). In another study, Dibutyl Phthalate did not interact with estrogen receptors as determined in three binding assays involving rat, human, and rainbow trout estrogen binding proteins.

Diethyl Phthalate (10^{-7} to 10^{-3} M) was evaluated in an estrogen receptor binding assay using rat uterine cytosol was acquired the uterine tissue of 10-wk-old Wistar rats.⁸ β -Estradiol was used as a positive control in a saturation binding assay, and caffeine was used as a negative control. Diethyl Phthalate or the positive or negative control material were allowed to compete with radiolabeled H^3 -E₂ to bind with the estrogen receptors in the rat uterine cytosol supernatant preparations. β -estradiol reduced the binding of H^3 -E₂ to the estrogen receptors with an EC₅₀ of 1.3×10^{-3} M. Caffeine and Diethyl Phthalate at concentration up to 10^{-3} M did not reduce H^3 -E₂ binding in this assay. A study found that Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate each had no binding affinity for the estrogen receptor and failed to prevent [3 H]-estradiol from binding to estrogen receptors obtained from the uteri of NCTR:SDN outbred rats.

Monobutyl phthalate, monoethyl phthalate, and monomethyl phthalate did not affect estradiol production in primary cultures of rat ovarian granulosa cells at concentrations up to 400 μ M. Monobutyl phthalate at 100 μ M did not affect aromatase RNA and aromatase protein levels in cultured granulosa cells.⁸

The effects of Diethyl Phthalate and monoethyl phthalate on the sex steroid hormone system were studied using human adrenocortical carcinoma (H295R) cells.⁸⁵ The cells were exposed to 0, 0.004, 0.04, or 0.4 mM of Diethyl Phthalate or monoethyl phthalate for 48 h. The media (Dulbecco's modified Eagle's) was collected for sex steroid hormone measurement using ELISA. Gene expression was analyzed from the total RNA extracted from the H295R cells. Diethyl Phthalate significantly reduced testosterone concentrations and significantly increased the 17 β -estradiol/testosterone ratio starting at 0.004 mM. These reductions and increases were observed with the metabolite, but statistical significance was observed at 0.04 mM with testosterone and 0.4 mM with the 17 β -estradiol/testosterone ratio. A statistically significant down-regulation of *Star* and *3 β HSD2* gene and up-regulation of *CYP19A* gene were also observed with both Diethyl Phthalate and monoethyl phthalate.

Animal

Monobutyl phthalate did not promote growth of the uterus in 20- to 22-d-old female rats (6 rats/dose group) that received oral doses of the test material at 0, 10, 100, 500, or 1000 mg/kg/d for 3 d.⁸ A positive control compound, estradiol benzoate (0.5 μ g/rat/d) did cause an increase in uterine weight.

Male zebrafish were exposed to Diethyl Phthalate (up to 10 mg/l) dissolved in DMSO for 14 d in accordance with OECD TG 204.⁸⁵ The fish were killed at the end of the exposure period, and liver and testis were dissected out of each fish. Blood samples were collected from the caudal vein. ELISA was used to measure sex steroid hormones and testosterone. Gene expression was analyzed from the total RNA extracted from fish organ samples. No mortalities were observed at any dose level. Diethyl Phthalate significantly decreased testosterone concentrations and 17 β -estradiol at 10 mg/l, decreases were non-significant at the lower doses. The test material also significantly up-regulated *cyp19a* gene expression at concentrations of 2 mg/l and higher and down-regulated *Star* (statistical significance at 10 mg/lg) and *3 β HSD* genes.

In a maternal toxicity study, pregnant Sprague-Dawley rats were administered streptozotocin (STZ; 40 mg/kg, intraperitoneally, gestation day 0) and/or Dibutyl Phthalate (750 mg/kg/d, by gavage, gestation days 1 - 3) to establish a gestational diabetes mellitus (GDM) model.⁸⁶ Dibutyl Phthalate alone caused unstable blood glucose, whereas combined exposure to STZ and Dibutyl Phthalate significantly increased fasting glucose and oral glucose tolerance test values to the GDM range and decreased serum insulin levels on gestation day 18 ($p < 0.05$). Histological and immunohistochemical examination of the pancreas demonstrated significant suppression of FoxM1 ($p < 0.001$) and increased pSTAT1 expression ($p < 0.05$). In vitro, primary islet β cells exposed to Dibutyl Phthalate exhibited reduced viability, suppressed FoxM1 expression, and increased apoptosis in a concentration-dependent manner, as confirmed by CCK-8 assays, flow cytometry, real-time PCR, and Western blot. Fludarabine, a STAT1 phosphorylation antagonist, partially restored β -cell viability, whereas IFN γ , a phosphorylation agonist, further suppressed FoxM1. These findings indicate that maternal Dibutyl Phthalate exposure can aggravate GDM through pSTAT1-mediated suppression of FoxM1, leading to impaired β -cell function, reduced insulin secretion, and hyperglycemia.

Evaluation of Mechanistic Evidence for Reproductive Toxicity

In a comprehensive evaluation conducted by a US Environmental Protection Agency (EPA) research group, mechanistic and toxicological data were systematically organized using the Adverse Outcome Pathway (AOP) framework to assess the mechanisms and species relevance of Dibutyl Phthalate-induced male reproductive toxicity.⁸⁷ Employing a structured weight-of-evidence approach, the authors performed extensive literature searches across multiple databases (PubMed, Web of Science, Toxline, and TSCA submissions) with no date or language restrictions, supplemented by manual screening of

regulatory citations. Eligible studies meeting predefined Population–Exposure–Comparator–Outcome criteria were compiled into a curated database capturing life stage, exposure conditions, and mechanistic endpoints. Evidence from rodent studies, ex vivo human fetal testis tissue cultures, and human fetal xenograft models was evaluated according to the mode-of-action/AOP framework, integrating molecular, cellular, tissue, and organism-level responses. Three relevant biological elements were identified: 1) fetal rats are more sensitive than other rodents and human fetal xenografts to Dibutyl Phthalate-induced anti-androgenic effects, 2) Dibutyl Phthalate-induced androgen-independent adverse outcomes are conserved among different mammalian models and human fetal testis xenografts, and 3) Dibutyl Phthalate-induced anti-androgenic effects are conserved in different mammalian species when exposure occurs during postnatal life stages.

Cytotoxicity

A 10 ppm solution of Dibutyl Phthalate in phosphate buffer at pH 7 decreased the percent survival of the yeast, Saccharomyces cerevisiae, throughout a 48-h incubation; a 20 ppm solution was even more toxic.² A concentration of 50 ppm of Dibutyl Phthalate completely inhibited the growth of cells of the protozoan, Tetrahymena pyriformis. Other phthalate esters were inhibitory as well.

The minimum inhibitory concentration of a 10% Diethyl Phthalate solution (in 95% ethanol) was 1000 ppm for Corynebacterium sp. and > 1000 ppm for Staphylococcus aureus and Escherichia coli.² The minimum inhibitory concentration of a 10% (w/v) solution of Diethyl Phthalate (in 95% ethanol) was 500 ppm for the fungus, Candida albicans. The concentration of neutralized Diethyl Phthalate that inhibited the multiplication of Pseudomonas putida was greater than 400 ppm. Neutralized Diethyl Phthalate inhibited the multiplication of the flagellate protozoan, Entosiphon sulcatum at a concentration of 19 ppm. Neutralized Diethyl Phthalate inhibited the multiplication of M. aeruginosa at a concentration of 15 ppm and inhibited the multiplication of the green alga, Scenedesmus quadricauda, at a concentration of 10 ppm.

In a validation study for a phototoxicity assay, cultured skin cells were exposed to 10, 30, 50, 70, or 100% Diethyl Phthalate and exposed to UVA and UVB radiation for 30 minutes.⁸ Controls were not irradiated. At concentrations of 30% and higher, virtually all cells in the culture system were killed, with or without radiation. At 10% Diethyl Phthalate, 70% viability of the cells were observed.

A concentration of 1000 ppm of Dimethyl Phthalate markedly inhibited the growth rate of Tetrahymena pyriformis. The growth of Pseudomonas aeruginosa was slightly inhibited by 1500 ppm Dimethyl Phthalate. The growth of the blue-green alga, Microcystis aeruginosa, was inhibited by 100 to 300 ppm of Dimethyl Phthalate and suppressed for 3 d by 400 ppm. After 4 d, cellular lysis was observed in the 400 ppm Dimethyl Phthalate culture, and concentrations of 500 ppm to 800 ppm completely destroyed the cells within 72 h.

The metabolism and toxicity of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate in cultures of mouse fibroblast have been investigated.² The ID_{50} (does required to inhibit growth by 50%) for Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate determined from dose-response curves were 1×10^{-4} , 3×10^{-3} , and 7×10^{-3} M, respectively. Toxicity to mouse fibroblasts was also investigated using a cell overlay method where pads containing 0.05 ml of a 50 mg/ml emulsion of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were placed on the agar surface (2.5 mg per pad), and the cells were observed for 48 h. Diethyl Phthalate and Dimethyl Phthalate were toxic to cells; however, Dibutyl Phthalate was not toxic to cells. In another study, a toxic response of mouse fibroblastic cells to 1, 5, 10, and 50% suspensions of Diethyl Phthalate and Dimethyl Phthalate was observed; Dibutyl Phthalate elicited a toxic response in the 10 and 50% suspensions only. A decrease in cellular ATP concentrations was observed over a 6-h incubation with Dibutyl Phthalate and Diethyl Phthalate.

The effects of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate on the outgrowth of nerve fibers and fibroblasts in primary cultures of rat cerebellum were investigated.² Dibutyl Phthalate and Diethyl Phthalate completely inhibited outgrowth at concentrations greater than or equal to 1.17×10^{-3} M and 1.53×10^{-3} M, respectively. Dimethyl Phthalate did not completely inhibit outgrowth at concentrations less than or equal to 3.05×10^{-3} M. Dibutyl Phthalate inhibited cell growth in addition to causing morphological changes, the appearance of lipid drops in the cytoplasm, and the accumulation of triacylglycerol in the cytosol of human embryonic lung cell cultures after the addition of 40 µg/ml of the test material in culture medium the cells. Dibutyl Phthalate and Diethyl Phthalate were inactive in a test in which the extent of membrane damage in human lung fibroblasts was determined by measuring the amount of a radioactively labeled cytoplasmic marker released into the media. In another study, the ID_{50} of Dibutyl Phthalate for human diploid cell strain WI-85 was 1.35×10^{-4} M. Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate killed and lysed human amnion and KB human cancer cells in culture. The 7-d IC_{50} (geometrical mean values between the totally inhibitory concentrations and the maximal completely non-injurious ones) were 3.1×10^{-2} , 6.3×10^{-2} , and 7.7×10^{-2} M for Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate, respectively, in HeLa cells.

Limb bud cells from embryonic rats (gestation day 12.5) were exposed in vitro to a range of concentrations of Dibutyl Phthalate or monobutyl phthalate for 96 h.⁸ Cytotoxicity was measured by uptake of neutral red, and inhibition of cell differentiation was measured by alcian blue staining. The IC_{50} values for cytotoxicity and cell differentiation were 25.54 µg/ml and 21.21 µg/ml, respectively, for Dibutyl Phthalate. Monobutyl phthalate had respective IC_{50} values for cytotoxicity and cell differentiation at 307.24 µg/ml and 142.61 µg/ml. Co-treatment with antioxidants, such as catalase or vitamin E,

protected the cells from Dibutyl Phthalate cytotoxicity and reduced the Dibutyl Phthalate-inhibition of cell differentiation. However, antioxidants did not have a protective effect in monobutyl phthalate-exposed cells.

The cytotoxicity potential of Dibutyl Phthalate was tested on bovine cultured lymphocytes using the MTT assay.⁷⁴ A concentration-dependent decrease in cell viability was observed when tested on concentrations ranging from 10 μ M to 100 μ M.

In another study using the MTT assay, human sperm were exposed to 13.47, 67.35, or 134.7 μ g/ml Dibutyl Phthalate for up to 96 h.⁵⁵ A concentration- and duration-dependent decrease in cell viability to Dibutyl Phthalate was observed. Cells exposed to low, medium and high concentrations at 24, 48, 72 and 96 h showed cell viability varies from 80 to 90%, 75 to 89%, 52 to 72% and 42 to 65%, respectively. Cytotoxicity was apparent only at highest dose level after 98 h exposure of Dibutyl Phthalate, where cell viability was observed to be 42%.

The human keratinocyte cell line, HaCaT cells, were utilized to assess the potential of cytotoxicity of Dibutyl Phthalate using the MTT assay, immunocytochemical, flow cytometric analysis, and western blotting.⁸⁸ Cells were exposed to up to 5 mg/ml Dibutyl Phthalate for 24 h in the cell viability assay and to 1 mg/ml Dibutyl Phthalate for 24 h in the other assays. HaCaT cell viability was 85, 81, 75, 62, and 37% after exposure to 0.1, 0.5, 1, 2, and 5 mg/ml Dibutyl Phthalate, respectively. The lowest-observed-effect concentration (LOEC) was 0.1 mg/ml. Dibutyl Phthalate at 1 mg/ml led to significantly decreased colony formation in HaCaT cells. Flow cytometric analysis showed 10% of the cells underwent apoptosis, with a 4.9-fold apoptotic fraction when compared to the vehicle control (ethanol).

The effects of Dibutyl Phthalate (0.0006 - 1.25 μ g/ml), Diethyl Phthalate (0.0002 - 0.5 μ g/ml), and Dimethyl Phthalate (0.0061 - 12.5 μ g/ml) were studied in an MTT cytotoxicity assay in human prostate cancer cell lines DU145 and PC3.⁸⁹ α -Lipoic acid was used as an antioxidant compound. A low dose proliferative effect was observed. In combination with the antioxidant, at IC₅₀ and lower doses an increase of cytotoxic effect was observed for Dibutyl Phthalate (IC₅₀ 27.32 ppb; 25.22% increase) in DU145 cells, while a decrease was observed with Diethyl Phthalate (IC₅₀ 1905.53 ppb; 8.12% decrease) and Dimethyl Phthalate (IC₅₀ 2390.48 ppb; 9.27% decrease). With the PC3 cells, a decrease was observed for Dibutyl Phthalate (IC₅₀ 77.21 ppb; 16.02%), Diethyl Phthalate (IC₅₀ 477.13 ppb; 17.01%), and Dimethyl Phthalate (IC₅₀ 1301.78 ppb; 26.58%)

Enzyme Effects

Rat liver incubated in vitro with 2×10^{-3} M Dibutyl Phthalate had no effect on epoxide hydratase or glutathione-S-transferase activities, decreased the monooxygenase activities, and decreased the conjugation of o-aminophenol and 4-methylumbelliferone with glucuronic acid.² In another study, Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate at concentrations of 5×10^{-5} to 1×10^{-3} M inhibited the respiration of isolated mitochondria from rat liver primarily by uncoupling oxidative phosphorylation rather than by inhibiting electron transport or energy transfer. Other researchers using the same concentrations have suggested that the contrary is probably true; Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate inhibited electron transport or energy transfer. In some studies, Dibutyl Phthalate and Dimethyl Phthalate inhibited the activities of succinate dehydrogenase and ATPase, enzymes of the rat liver inner mitochondrial membrane, after intraperitoneal administration, and in in vitro assays at concentrations of 1×10^{-4} to 1.5×10^{-3} M. Dibutyl Phthalate stimulated ATPase activity and induced swelling of rat liver mitochondria.

Administration of 0.7% Dibutyl Phthalate or 0.5% Dimethyl Phthalate in the diet of male rats for 21 d increased hepatic weights and reduced serum cholesterol concentrations.² Acetate incorporation into triglycerides and the steryl ester plus squalene and mevalonate incorporation into squalene plus sterols in liver minces were inhibited by dietary Dibutyl Phthalate; these results were not observed with Dimethyl Phthalate. Dimethyl Phthalate administration resulted in a decrease in total hepatic cholesterol and lipid that was not observed with Dibutyl Phthalate. The intraperitoneal administration of 20 mg/kg/d of Dibutyl Phthalate to mice for 16 d did not significantly lower serum cholesterol but did lower serum triglycerides. Dibutyl Phthalate, at a concentration of 2.5×10^{-6} M inhibited mouse liver homogenate acetyl-CoA synthetase, citrate lyase, and acetyl-CoA carboxylase but not fatty acid synthetase. These enzymes are involved in the cholesterol and triglyceride synthesis pathways. A 5×10^{-6} M concentration of Dibutyl Phthalate and Diethyl Phthalate inhibited in vitro human blood lecithin/cholesterol acyltransferase; whereas the enzyme was only slightly inhibited by the same concentration of Dimethyl Phthalate.

Single-dose intraperitoneal administration of 3.05 ml/kg of Dibutyl Phthalate or 3.6 mg/kg of Dimethyl Phthalates to rats inhibited the activity of hepatic aminopyrine N-demethylase and aniline hydroxylase and had no effect on glucose-6-phosphatase, NADPH-cytochrome c reductase, and tyrosine aminotransferase activity.² The activities of these enzymes were not decreased when the test materials were administered intraperitoneally every day for 7 d. Results of another study indicated that Dibutyl Phthalate weakly enhanced the activity of aminopyrine N-demethylase from rat hepatic 10,000 g supernatant. The oral administration of 5 mmol/kg/d of Dibutyl Phthalate for 6 d to male rats increased the hepatic cytochrome P-450, had no effects on glutathione-S-transferase activity or the monooxygenase activities dependent on cytochrome P-450, increased the epoxide hydratase activity, and increased the conjugation of o-aminophenol and 4-methylumbelliferone with glucuronic acid.

Dibutyl Phthalate and monobutyl phthalate were reported to inhibit mitochondrial respiration by uncoupling energy-dependent processes and inhibiting succinate dehydrogenase activity, causing a diminished energy supply, altered ion transport, and mitochondrial swelling.⁸ Additionally, Dibutyl Phthalate was reported to inhibit the metabolism of arachidonic acid by cyclooxygenase and lipoxygenase in rat peritoneal leukocytes. Aliquots of the rat leukocytes were pre-incubated with 0, 1, or 10, or 100 µg/ml Dibutyl Phthalate and then incubated with 1.7 nM radiolabeled arachidonic acid for 5 min before reactions were stopped with methanol and formic acid. Cells pre-incubated with Dibutyl Phthalate showed reduced concentrations of prostaglandins, LTB₄, 5-HETE, 11-HETE, and 15-HETE.

To evaluate the effects of Dibutyl Phthalate on mitochondrial enzymes in the liver, male Wistar rats were fed a diet containing 5% Dibutyl Phthalate for 35 to 45 d.⁸ The treated rats had reduced body weights; increased liver, kidney, and spleen organ weights; and reduced testicular weights ($p < 0.01$) when compared to control rats. Respiration in liver mitochondria and succinate dehydrogenase and pyruvate dehydrogenase activities were reduced in Dibutyl Phthalate-treated rats. In in vitro preparations of liver cells, less than 100 µM Dibutyl Phthalate inhibited respiration, and 150 µM Dibutyl Phthalate inhibited dehydrogenase activity.

The inhalation of Dibutyl Phthalate (0.5, 2.5, or 7.0 ppm) for 5 d dose-dependently decreased lung microsomal cytochrome P-450, and was accompanied by a reduction in n-hexane and benzo[a]pyrene (BaP) metabolism.⁸ Liver cytochrome P-450 was not affected, but n-hexane and BaP metabolism in liver was increased. Additionally, Dibutyl Phthalate was a weak inducer of liver cytochrome P-450 in male Sprague-Dawley rats that were dosed with 0, 0.01, 0.1, or 1.0 mmol/kg via gavage for 5 d. Cytochrome P-450 and NADPH-cytochrome-c-reductase activities in the liver were increased at the lowest Dibutyl Phthalate dose, and the microsomal enzyme activity did not return to normal after a 4-wk recovery period. Liver microsomal metabolism of n-hexane was increased by Dibutyl Phthalate, but the increased n-hexane metabolism did not persist. Metabolism of BaP was not affected by Dibutyl Phthalate.

Oral intake of Diethyl Phthalate (5.4 mmol/kg/d for 3 d) caused a slight (~1.6 fold) increase and Dibutyl Phthalate (3 mmol/kg/d) caused a 5-fold increase in the activity of hepatic laurate 12-hydroxylase.⁸ Likewise, palmitoyl-CoA oxidase activity was only slightly increased by Diethyl Phthalate, but increased 6-fold in animals dosed with Dibutyl Phthalate (no further details available).

In an in vitro study using recombinant human microsomes, Dibutyl Phthalate and its monoester metabolite monobutyl phthalate were evaluated for their effects on CYP2C9*1 and CYP2C19*1 enzyme activities.⁹⁰ The assays were performed using 5 µM diclofenac (for CYP2C9*1) and voriconazole (for CYP2C19*1) as probe substrates, with metabolite formation quantified by ultra-high-performance liquid chromatography (UHPLC) or UHPLC–MS/MS. Dibutyl Phthalate produced concentration-dependent inhibition of both CYP2C9*1 and CYP2C19*1 activities across the tested concentration range (0.1 - 100 µM). At 100 µM, CYP2C9*1 activity was inhibited by 67.3% ($IC_{50} = 29.63$ µmol/l; $K_i = 7.06$ µmol/l), while CYP2C19*1 activity was completely inhibited (100%) with an IC_{50} of 2.63 µmol/l and K_i of 7.01 µmol/l. Kinetic analyses demonstrated that Dibutyl Phthalate was a competitive inhibitor of the respective probe substrates and a noncompetitive inhibitor with respect to the cofactor NADPH. In contrast, monobutyl phthalate had little effect on either enzyme, producing 1.34% inhibition of CYP2C9*1 and 23.58% inhibition of CYP2C19*1 at 100 µM.

Hepatotoxicity

Mice that received up to 50 mg/kg/d Dibutyl Phthalate in Tween 80 via gavage for 28 d had disrupted hepatic architecture, inflammatory cell infiltration, and hepatocellular disorganization.⁹¹ Serum alanine aminotransferase (ALT) and aspartate aminotransferase (AST) activities were elevated, with the AST/ALT ratio significantly increased. Sirius red staining demonstrated collagen deposition in Dibutyl Phthalate-treated livers, and protein expression of fibrosis markers (α -SMA, TGF- β , Col-I) was significantly upregulated. In a HepG2 and LX-2 co-culture, treatment with inhibitors targeting p38MAPK (SB203580), NF- κ B (PDTC), and NLRP3 (MCC950) attenuated LX-2 activation, suppressed p38MAPK/NF- κ B signaling, and reduced Dibutyl phthalate-induced pyroptosis.

The effects of Dibutyl Phthalate on hepatic structure and function were studied using adult albino male mice.⁹² Groups of 10 mice received the corn oil, Dibutyl Phthalate (250 mg/kg), diethylhexyl phthalate (300 mg/kg), or Dibutyl Phthalate + diethylhexyl phthalate (250 + 300 mg/kg) in corn oil (0.2 ml) via gavage for 28 d. An untreated group served as a negative control. All animals were killed at the end of the treatment period and blood was collected for serum lipid profiles and liver function tests. Body weights and liver weights were measured. Liver samples were also collected for histopathological examination, and microphotographs of hematoxylin and eosin-stained sections were evaluated with micrometry.

In the Dibutyl Phthalate group, the mean body weights of the mice were not significantly affected, but a significant ($p < 0.05$) increase was observed in the diethylhexyl phthalate and Dibutyl Phthalate + diethylhexyl phthalate groups. All phthalate treatment groups had significant increases in liver enzymes, liver proteins, and altered serum cholesterol, triglycerides, LDL, and VLDL levels. In the mice treated with the phthalates, marked alterations leading to hepatic hypertrophy were observed. Additionally, a significant ($p < 0.01$; $p < 0.05$) increase in the mean number of micronucleated cells and binucleated cells per unit area was observed following treatment with Dibutyl Phthalate. Treatment with the phthalates caused disorganized structure of trabeculae arising from the hepatic vein, deshaped mononucleated hepatocytes, binucleated hepatocytes with swollen nuclei, dilated sinusoidal spaces, regeneration zones cellular hypertrophy, and intracellular vacuolation in the liver, when compared to the control group. Dibutyl Phthalate and Dibutyl Phthalate +

diethylhexyl phthalate caused a highly significant ($p < 0.001$) increase in the relative nucleo-cytoplasmic index of mononuclear hepatocytes when compared to the control group. The researchers concluded that Dibutyl Phthalate and diethylhexyl phthalate alone or in combination, had toxic potential to hepatic tissue.⁹²

A 28-d oral gavage study in male Kunming mice (8 per group) was conducted with Dibutyl Phthalate at 50 mg/kg/d.⁹³ Dibutyl Phthalate exposure resulted in significant increases in the AST/ALT ratio and decreases in albumin (ALB) levels, indicating hepatocellular injury. Histological evaluation showed hepatocyte edema, unclear hepatic cord structure, sinusoidal narrowing, and central vein dilation. Dibutyl Phthalate treatment also increased hepatic ROS, MDA, and phosphorylated ERK1/2. Co-administration of Vitamin E (antioxidant) or PD98059 (ERK inhibitor) attenuated hepatic biochemical and histological alterations, indicating that Dibutyl Phthalate-induced hepatotoxicity is mediated, at least in part, through oxidative stress and ERK1/2 activation.

Renal Toxicity

A 28-d oral gavage study in Kunming mice (8 per group) also assessed renal function following Dibutyl Phthalate exposure at 50 mg/kg/d.⁹³ Significant increases in serum creatinine and urea nitrogen were observed compared with controls. Histopathological examination demonstrated glomerular and tubular injury, characterized by cellular swelling and structural disorganization. Mechanistic evaluations revealed elevated renal ROS and MDA levels, and activation of ERK1/2 signaling. Co-treatment with vitamin E or PD98059 mitigated these biochemical and morphological alterations.

Another study investigated the effects of Dibutyl Phthalate (50 mg/kg/d) on renal function of male Kunming mice (7 per group) and the potential protective effects of curcumin (2.5 mg/kg/d).⁹⁴ Dibutyl Phthalate was administered orally, while curcumin was administered intraperitoneally for 28 d. Mice were divided into a control group, Dibutyl Phthalate group, curcumin group, and Dibutyl Phthalate + curcumin group. Kidney function, oxidative stress biomarkers, and apoptosis factors as well as Bcl-2 and Bax were investigated. A marked increase of renal dysfunction, oxidative stress and apoptosis level after Dibutyl Phthalate exposure was observed when compared to the control. Administration of curcumin to Dibutyl Phthalate-treated mice appeared to reduce these adverse biochemical changes when compared with the group that received Dibutyl Phthalate alone.

Renal tubular epithelial cells (HK-2) were exposed to Dibutyl Phthalate (100 μ M) to investigate mechanisms of renal fibrosis.⁹⁵ Dibutyl Phthalate exposure significantly upregulated connexin 43 (Cx43) expression at mRNA and protein levels ($p < 0.01$), as measured by RT-PCR and Western blot. Knockdown of Cx43 with siRNA attenuated Dibutyl Phthalate-induced epithelial-mesenchymal transition, indicated by restored E-cadherin ($p < 0.01$) and decreased N-cadherin ($p < 0.05$) and Snail1 expression ($p < 0.01$). Inhibitor studies further demonstrated that Dibutyl Phthalate increased Cx43 expression through activation of the Ang II/AMPK α 2 signaling pathway.

Splenic Toxicity

The effect of Dibutyl Phthalate (> 99% pure) on the spleen and potential protective effects of antioxidants were studied in male BALB/c mice.⁹⁶ Groups of 6 mice received saline (control; oral), 50 mg/kg/d Dibutyl Phthalate (oral), 50 mg/kg/d Vitamin E (oral), 50 mg/kg/d Dibutyl Phthalate + 50 mg/kg/d Vitamin E (intragastric), 2.5 mg/kg/d curcumin (intraperitoneal injection), or 50 mg/kg/d Dibutyl Phthalate + 2.5 mg/kg/d curcumin (intragastric) daily for 3 wk. After the treatment period ended, the mice were killed, and the spleens were collected for tissues sectioning for histological examination and preparation of a homogenate. The homogenate was used to determine the levels of ROS, glutathione, IL-1 β , necrosis factor- α (TNF- α), cytochrome C, and caspase-3 and -9. Histopathological changes of the spleen were observed in the 50 mg/kg/d Dibutyl Phthalate group, including inflammation and apoptosis associated with oxidative stress. The antioxidants were observed to alleviate spleen tissue damage; reduce oxidative stress, IL-1 β , and TNF- α ; and reduce the levels of apoptosis-associated factors in the mice exposed to Dibutyl Phthalate.

In a 90-day oral study, Sprague-Dawley rats were divided into four groups ($n = 5$ /group): control, BaP (1 mg/kg bw/d), Dibutyl Phthalate (50 mg/kg bw/d), and combined BaP + Dibutyl Phthalate.⁹⁷ Test substances were dissolved in corn oil and administered once daily by gavage. Dibutyl Phthalate alone induced moderate splenic alterations, including blurred boundaries between the white pulp and marginal zone and an enlarged marginal. Serum cytokine levels of IL-1 β and IL-18 (~ 1.3-fold increase) were elevated compared to controls. In spleen tissue, Western blot analyses indicated increased expression of the pyroptosis-related proteins IL-1 β , ASC, and GSDMD (~ 1.3–1.5-fold increase). Dibutyl Phthalate exposure also led to reduced Arg1 expression (~ 0.8-fold decrease; M2 macrophage marker) and increased iNOS expression (~ 1.4-fold increase; M1 macrophage marker), indicating a shift toward pro-inflammatory macrophage polarization. Co-exposure to BaP and Dibutyl Phthalate produced more severe splenic injury and a stronger inflammatory response than either chemical alone. Compared to Dibutyl Phthalate or BaP exposure, the combined group showed synergistic increases in iNOS (1.79-fold), NLRP3 (2.11-fold), and cathepsin B (1.9-fold) protein levels, accompanied by a further reduction in Arg1 (0.8-fold decrease). In vitro studies using primary splenic macrophages demonstrated that Dibutyl Phthalate alone reduced cell viability and activated caspase-1, consistent with the induction of pyroptosis observed in vivo.

Insulin Effects

In Vitro

A culture system of rat insulinoma cells was used to assess the effect of Dibutyl Phthalate on insulin synthesis and secretion.⁹⁸ The cells were treated with 15, 30, 60, or 120 μ M Dibutyl Phthalate or just the vehicle, DMSO, alone for 24 h. The contents of insulin in the intracellular and extracellular fluid of the cells were measured, and the apoptosis rate and mitochondrial membrane potential of the cells were measured by flow cytometry with annexin V-fluorescein isothiocyanate (FITC) conjugate and PI and JC-1, respectively. Western blot was used to measure oxidative stress. Insulin synthesis and secretion in the insulinoma cells were significantly decreased ($p < 0.01$) in the high dose group. Dibutyl Phthalate also caused a significant increase ($p < 0.01$) in apoptosis rate at 30 and 60 μ M, and a significant decrease ($p < 0.01$) in the mitochondrial membrane potential in the insulinoma cells starting at 60 μ M. The expression of Bax/Bcl-2, caspase-3, caspase-9, and Cyt-C were significantly increased, starting at 30 μ M for caspase 9 and at 60 μ M for the remaining proteins. Starting at 30 μ M, oxidative stress levels (MDA, GSH/GSSG) were increased, and the antioxidant index (SOD) levels were decreased.

Animal

The role of Dibutyl Phthalate (> 99% pure) on the development of type 2 diabetes was studied in male Balb/c mice.⁹⁹ Groups of 8 mice received 0, 0.5, 5, or 50 mg/kg/d Dibutyl Phthalate orally for 7 wk in combination with a high fat diet and injections of a low dose of streptozotocin. One group received 50 mg/kg/d Dibutyl Phthalate without the high fat diet and injections. Exposure to 50 mg/kg/d Dibutyl Phthalate alone induced a marked decrease in insulin secretion and glucose intolerance, but had no influence on insulin resistance. Combined with a high fat diet and treatment with streptozotocin, Dibutyl Phthalate exposure markedly aggravated glucose intolerance, insulin tolerance, and insulin resistance. Additionally, this combination induced lesions in the pancreas and kidney. Further investigation of the role of Dibutyl Phthalate on the insulin signaling pathway found that Dibutyl Phthalate could disrupt phosphatidylinositol 3 kinase (PI3K) expression and AKT (protein kinase B) phosphorylation, and decrease the level of glucose transporter 2 (GLUT2) in the pancreas.

The effects of Dibutyl Phthalate (98% pure) on glucose homeostasis and insulin secretion were studied in groups of 10 male Sprague-Dawley rats.¹⁰⁰ The rats received corn oil (solvent control), Dibutyl Phthalate (500 mg/kg), diethylhexyl phthalate (750 mg/kg), or both Dibutyl Phthalate and diethylhexyl phthalate (500 + 750 mg/kg) via gavage 5 d/wk for 8 wk. Body weights were determined weekly and fasting blood glucose was measured in tail vein blood at 0, 2, 4, 6, and 8 wk. At the end of the study, an oral glucose tolerance test and insulin resistance test were performed. The rats were killed at study. Blood samples were collected from the heart, and the pancreas, perigenital fat, and one side of the gastrocnemius muscle were collected and weighed. During the study, there were no clinical signs of toxicity or mortality. Starting at 2 wk, rats that received Dibutyl Phthalate + diethylhexyl phthalate had significantly lower ($p < 0.05$) body weights than control rats, and non-significantly lower ($p > 0.05$) body weights than the individual phthalate groups. The organ coefficient did not differ significantly between the treatment groups for pancreatic tissue, perigenital fat, and gastrocnemius tissue. Fasting blood glucose was elevated in the Dibutyl Phthalate + diethylhexyl phthalate group at 2 and 4 wk, and in all treatment groups at 6 and 8 wk. Dibutyl Phthalate alone and with diethylhexyl phthalate caused an increase in insulin resistance index, a decrease in insulin sensitivity, and impairment of the function of pancreatic islet β -cells. Dibutyl Phthalate, with and without diethylhexyl phthalate, triggered an increase in plasma MDA and reduction in SOD activity; a reduction in the phosphorylation of PI3K and phospho-protein kinase B (p-Akt473) proteins; an increase in the relative expression of Bax, Caspase-8, cleaved-Caspase-9, and cleaved-Caspase-3; and a reduction in the relative expression of Bcl-2-related Bax in pancreatic tissue and of gastrocnemius GLUT4 in the gastrocnemius muscle.

Cardiovascular Effects

In Vitro

Human endothelial EA.hy926 cells were exposed to Dibutyl Phthalate at 10^{-6} , 10^{-5} , and 10^{-4} M to evaluate its effect on angiogenesis.¹⁰¹ Dibutyl Phthalate at 10^{-4} M significantly enhanced tube formation without affecting cell viability and upregulated angiogenesis-related genes (VEGFA, KDR, and CDH5). The pro-angiogenic effect was abolished by ERK1/2 inhibitor (U0126), PI3K-Akt inhibitor (wortmannin), and nitric oxide (NO) synthase inhibitor (L-NAME). Short-term exposure (15 - 120 min) to 10^{-4} M Dibutyl Phthalate induced phosphorylation of ERK1/2, Akt, and eNOS and increased NO production after 24 - 48 h. Co-treatment with nuclear estrogen receptor (ER) antagonist ICI 182,780 or G-protein-coupled estrogen receptor (GPER) antagonist G-15 eliminated these effects. This study demonstrated that Dibutyl Phthalate promoted angiogenesis in human endothelial cells through ER- and GPER-dependent activation of ERK1/2, PI3K-Akt, and nitric oxide signaling pathways.

Animal

Groups of 6 Wistar rats received Dibutyl Phthalate (25 mg/kg) or Dibutyl Phthalate plus rutin (50 mg/kg) orally for 28 d.¹⁰² A control group received 0.1% DMSO. Cardiac lipid peroxidation, antioxidants, and inflammatory markers activities were measured. Dibutyl Phthalate reduced catalase (CAT) activity, GSH, and Nrf2. Additionally, Dibutyl Phthalate increased tissue C-reactive protein (CRP) and tissue nuclear factor kappa B (NF- κ B). However, rutin efficiently reduced CRP and NF- κ B, increasing GSH and Nrf2 levels in Dibutyl Phthalate-exposed rats.

In a mechanistic study evaluating the effects of Dibutyl Phthalate on hypertension, male C57BL/6 strain mice (n = 9/group) were orally administered Dibutyl Phthalate by gavage at doses of 0.1, 1, or 10 mg/kg/d for 6 wk, and their blood pressure was monitored using the tail-cuff plethysmography method.¹⁰³ Across all Dibutyl Phthalate dose groups, no significant changes in blood pressure were observed compared to vehicle controls. Dibutyl Phthalate exposure caused a significant increase in serum estradiol levels in the 1 and 10 mg/kg/d groups ($p < 0.01$). Co-treatment with ICI182780, an estrogen receptor antagonist, inhibited the Dibutyl Phthalate-induced elevation in NO levels.

Immunotoxicity

Animal

Zebrafish embryos (30 per treatment; 3 strains used: transgenic, albino, and AB line) were exposed to 0.02, 0.2, or 2 μ M Dibutyl Phthalate.¹⁰⁴ Immune cell formation and phagocytosis ability were measured after 6 - 96 h post-fertilization. Respiratory burst activity was measured to assess the immune-toxic effects of the test material. Analysis of T/B relevant gene expression was also performed. No statistically significant differences in hatching and survival rates among the treatment groups; however, some developmental abnormalities, such as yolk sac resorption and swim bladder inflation were inhibited at 2 μ M. Exposure to Dibutyl Phthalate inhibited neutrophil and macrophage formation in a concentration-dependent manner. The ability of macrophage phagocytosis was decreased after Dibutyl Phthalate exposure. Respiratory burst was induced, and transcription levels of T/B cell-related genes *rag1/2* were up-regulated.

In a murine model of contact hypersensitivity induced with the hapten FITC, Dibutyl Phthalate was identified as an adjuvant that promotes TSLP expression in skin keratinocytes.¹⁰⁵ FITC was resuspended in vehicle (1:1 mixture of Dibutyl Phthalate and acetone) to a concentration of 0.5%. Mice were sensitized with 100 μ l of 0.5% FITC or vehicle applied to the shaven or unshaven back or abdomen on days 0 and 2, or days 0 and 1. On day 6, baseline ear thickness was measured with calipers, followed by challenge with 20 μ l of 0.5% FITC on one ear and vehicle on the contralateral ear. On day 7 (24 h post challenge), ear swelling was assessed to calculate the change in thickness. Epicutaneous exposure to Dibutyl Phthalate, either alone or in combination with FITC, significantly upregulated TSLP mRNA and protein in keratinocytes. Functional studies in TSLP receptor-deficient (TSLPR^{-/-}) mice and in mice treated with neutralizing anti-TSLP antibodies confirmed that TSLP is required for the development of Th2-type contact hypersensitivity responses. In a similar mouse study of FITC induced contact hypersensitivity, Dibutyl Phthalate exhibited clear adjuvant activity.¹⁰⁶ Epicutaneous co-exposure to FITC and Dibutyl Phthalate significantly increased the number of FITC-positive CD11c⁺/CD8 α ⁺ dendritic cells in draining lymph nodes ($p < 0.01$), and a smaller but detectable increase was observed in FITC-positive CD11c⁺/B220⁺ plasmacytoid dendritic cells. Dibutyl Phthalate exposure also markedly enhanced IL-4 production, with lymph node cell cultures yielding ~100 - 150 pg/ml compared to undetectable levels (< 3 pg/ml) in FITC/acetone controls. By contrast, Diethyl Phthalate showed relatively weak activity on FITC-positive dendritic cell migration but still enhanced IL-4 production (~100 pg/ml) in draining lymph nodes.

In a related study with an FITC-induced contact hypersensitivity mouse model, female BALB/c mice were sensitized epicutaneously with 0.5% FITC (in acetone) on shaved forelimbs on days 0 and 7 in the presence of Dibutyl Phthalate (2%), and monobutyl phthalate (2%).¹⁰⁷ On day 14, mice were challenged on the ear with 0.5% FITC in acetone-Dibutyl Phthalate (1:1), and ear swelling was measured at 24, 48, and 72 h. In parallel, transient receptor potential ankyrin 1 (TRPA1) activation was evaluated in vitro using CHO cells expressing human TRPA1. Dibutyl Phthalate significantly enhanced ear swelling responses to FITC compared to acetone controls ($p < 0.001$), whereas monobutyl phthalate did not. In vitro, Dibutyl Phthalate activated TRPA1 with an EC₅₀ of 7.3 μ M, monobutyl phthalate required > 40-fold higher concentrations (EC₅₀ ~313 μ M). These results indicate that Dibutyl Phthalate itself, rather than its metabolite, enhances FITC-induced skin sensitization, at least in part via TRPA1 activation.

In another murine model of FITC-induced contact hypersensitivity, long-term dermal exposure to Dibutyl Phthalate enhanced Th2-type immune responses and aggravated atopic dermatitis.¹⁰⁸ Male BALB/c mice (n = 8/group) were dermally exposed to saline or Dibutyl Phthalate (0.4, 4, or 40 mg/kg/d) once daily for 40 consecutive days by topical application (0.1 mL) to the shaved backs. On days 41 and 42, mice were sensitized with 120 μ l of 0.5% FITC (in acetone/Dibutyl Phthalate 1:1 v/v) or saline alone. On day 47, baseline ear thickness was measured and animals were challenged with saline or 0.5% FITC on the right ear and vehicle on the left ear, followed by measurement of ear swelling 24 h later (day 48). High-dose Dibutyl Phthalate (40 mg/kg/d) significantly increased ear swelling and epidermal thickness ($p < 0.01$), with histopathologic evidence of inflammatory cell infiltration and mast cell degranulation. Co-exposure to Dibutyl Phthalate and FITC elevated serum total IgE and skin levels of Th2 cytokines (IL-4, IL-5, IL-13) and Th17 cytokine (IL-17A), while IFN- γ and TNF- α remained unchanged. Eosinophil accumulation, indicated by increased eosinophil cationic protein staining, was also observed. Dermal exposure to Dibutyl Phthalate markedly upregulated TSLP expression in skin, suggesting a TSLP-mediated mechanism in the Dibutyl Phthalate-induced enhancement of Th2-type hypersensitivity.

In a similar FITC-induced contact hypersensitivity model, topical exposure to Dibutyl Phthalate during the sensitization phase promoted Th2-type immune responses via TSLP signaling.¹⁰⁹ Female BALB/c mice were topically sensitized on days 0 and 1 with 0.5% FITC in acetone/Dibutyl Phthalate (1:1 v/v, 400 μ l on shaved abdomen or 20 μ l on ear), and challenged on the right ear with 0.5% FITC/Dibutyl Phthalate more than 5 days later. Ear thickness was measured before and 24 h after the elicitation. FITC/Dibutyl Phthalate sensitization significantly increased ear swelling ($p < 0.001$), and was associated with

elevated IL-4 and TNF- α expression in the ears, and increased plasma total IgE levels. Dibutyl Phthalate alone, but not FITC or acetone, induced TSLP expression in the skin during the sensitization phase. Removal of Dibutyl Phthalate from either the sensitization or elicitation phase abolished ear swelling, which was restored by recombinant mouse TSLP injection during sensitization. Conversely, TSLP knockdown using intradermal siRNA significantly suppressed ear swelling and IL-4 production without affecting IL-12p40 levels. Dibutyl Phthalate also enhanced the migration of FITC⁺/CD11c⁺ dendritic cells to draining lymph nodes, although TSLP did not independently induce this migration.

In a murine model of contact hypersensitivity, the adjuvant effect of Dibutyl Phthalate on skin sensitization to multiple haptens was evaluated.¹¹⁰ Female BALB/c mice (n = 7-10/group) were epicutaneously sensitized on shaved forelimbs with a suboptimal dose of each hapten dissolved in acetone or in an acetone/Dibutyl Phthalate mixture (1:1, v/v). For the isothiocyanate-containing haptens, such as 0.5 % (w/v) eosin 5-isothiocyanate (EITC) and 0.5 % (w/v) rhodamine B isothiocyanate (RITC), mice were treated twice (days 0 and 7); for 2.5 % (v/v) phenethyl isothiocyanate (PEITC), mice were treated four times (days 0, 1, 7, and 8), each time with 160 μ l of the hapten solution applied to the shaved forelimbs. On day 14, mice were challenged by applying 20 μ l of the same hapten solution (0.5 % for EITC and RITC, 2.5 % for PEITC) in acetone/Dibutyl Phthalate (1:1, v/v) to the right ear, while the left ear received vehicle only. For Th1-type haptens, mice were sensitized with 0.04 % (w/v) 2,4-dinitrofluorobenzene (DNFB) on days 0 and 1 or with 0.1 % (w/v) oxazolone (OXA) on day 0. For Th2-type haptens, sensitization was performed on day 0 using 2.5 % (w/v) trimellitic anhydride (TMA), 0.01 % (w/v) methylenediphenyl diisocyanate (MDI), or 0.05 % (w/v) toluene 2,4-diisocyanate (TDI). Mice were challenged with the corresponding hapten solutions on day 5, and ear thickness was measured before (0 h) and at 24 and 48 h post-challenge to evaluate sensitization. Dibutyl Phthalate did not enhance sensitization to Th1-type haptens (DNFB, OXA) or Th2-type haptens (TMA, MDI, TDI). In contrast, Dibutyl Phthalate significantly enhanced sensitization to isothiocyanate-containing haptens, including EITC, RITC, and PEITC. Mice co-exposed to Dibutyl Phthalate during sensitization exhibited markedly greater ear-swelling responses upon hapten challenge compared with those sensitized without Dibutyl Phthalate ($p < 0.01$).

In an allergic asthmatic mouse model, male Balb/c mice were sensitized and challenged with ovalbumin (OVA), exposed to Dibutyl Phthalate (40 mg/kg bw/d) by gavage on days 1 - 53, and subsequently exposed to OVA aerosol (30 min/d) on days 54 - 60.³⁴ Compared to OVA alone, co-exposure to OVA and Dibutyl Phthalate produced significantly greater lung inflammatory cell infiltration, higher serum IgE, and increased Th2 (IL-4, IL-5, IL-13) and Th17 (IL-17A) cytokine concentrations in bronchoalveolar (BAL) fluid. These effects were accompanied by alterations in lung histology and airway hyperresponsiveness ($p < 0.01$), consistent with allergic asthma. Oxidative stress was also enhanced, as evidenced by decreased GSH and elevated MDA and 8-OHdG levels ($p < 0.01$), along with significantly increased lung calcitonin gene-related peptide (CGRP) expression. Co-treatment with melatonin attenuated oxidative stress and CGRP expression and alleviated asthma-like symptoms. Collectively, these results indicate that Dibutyl Phthalate exacerbates allergic airway responses through oxidative stress-mediated up-regulation of CGRP.

Human

In a double-blinded, randomized crossover human exposure study, 16 allergen-sensitized adults (sensitized to grass, n = 5; birch, n = 5; house dust mite, n = 6) were exposed for 3 h in an environmental chamber to filtered air or Dibutyl Phthalate (mean concentration $\sim 150 \mu\text{g}/\text{m}^3$), immediately followed by an inhalation challenge with their specific sensitizing allergen.¹¹¹ Dibutyl Phthalate exposure significantly enhanced the early allergen response, resulting in a 21.4% greater decline in FEV1 (forced expiratory volume in 1 second, a standard measure of lung function) area under the curve ($p = 0.03$), and increased airway responsiveness by 48.1% in participants without baseline hyperresponsiveness ($p = 0.01$). Dibutyl Phthalate exposure modestly increased the total number of macrophages in BAL (4.6%; $p = 0.07$), but significantly increased the proportion of M2-polarized macrophages (46.9%; $p = 0.04$). Effects on immune mediator levels were generally limited; among 65 analytes assessed, only a few were altered, with more pronounced changes in bronchial wash compared to BAL. Collectively, these findings indicate that Dibutyl Phthalate exposure may exacerbate allergen-induced impairments in lung function and exerts immunomodulatory effects in the airways of sensitized individuals. In a follow-up analysis, Dibutyl Phthalate also altered systemic immune endpoints, increasing circulating CD4⁺ T helper cells and reducing regulatory T cells at both 3 and 20 h post-exposure, with reductions in non-classical monocytes.¹¹² Notably, the cells and mediators affected by Dibutyl Phthalate were largely distinct from those altered by allergen inhalation alone, suggesting that Dibutyl Phthalate exerted systemic immunomodulatory effects independent of simple enhancement of allergen responses.

In Vitro Keratinocyte Monolayer Penetration Model

A monolayer of human keratinocytes cultured on a Transwell insert (24-mm diameter, 0.4- μm polycarbonate membrane; effective diffusion area 4.52 cm²) was used to evaluate the percutaneous penetration of Diethyl Phthalate and Dibutyl Phthalate.¹¹³ A complete culture medium containing 200 or 500 ng/ml of test compound was applied to the apical side (1.5 ml) for 540 min, and 2.5 ml Hanks' balanced salt solution served as the receptor phase. Samples of receptor fluid, apical medium, and cell layers were collected and analyzed by gas chromatography – mass spectrometry (GC–MS). The 9-h absorption factor (AF) values for Diethyl Phthalate were 62 and 39%, and for Dibutyl Phthalate were 53 and 34%, in the low- and high-dose groups, respectively. The corresponding 1-h AF-values were 6.9 and 4.3% for Diethyl Phthalate, and 5.8 and 3.7% for Dibutyl Phthalate. The permeability coefficient values were inversely correlated with the log K_{ow} , while cumulative

amounts of phthalates in the cells increased with log K_{ow} , indicating phthalates with higher hydrophobicity were more likely to remain on the skin.

DERMAL IRRITATION AND SENSITIZATION STUDIES

Animal

Slight to moderate dermal irritation was observed in rabbit studies with Dibutyl Phthalate (concentrations not reported).^{2,8} Diethyl Phthalate also caused slight to moderate skin irritation in several rabbit studies, including a study with undiluted material and in a shampoo formulation at 1%.⁸ Dimethyl Phthalate was not irritating in rabbits or guinea pigs (concentrations not reported).^{2,8}

No sensitization was observed in rabbits that received up to 4 ml/kg/d Dibutyl Phthalate or Dimethyl Phthalate (concentration not reported) for 90 d.² Diethyl Phthalate (50%) and Dimethyl Phthalate (concentration not reported) was not sensitizing to guinea pigs.⁸

Human

Dibutyl Phthalate at 5% in petrolatum was not irritating in 48-h closed patch tests with 53 subjects.² One positive reaction was reported in 1532 subjects that underwent patching with 2% Dibutyl Phthalate, Diethyl Phthalate, or Dimethyl Phthalate in petrolatum in a 48-h closed patch test (no further details provided). A nail polish containing 9% Dibutyl Phthalate was slightly irritating in a 23-h patch test on 13 subjects and not irritating in a 48-h patch test on 25 subjects when tested neat. Another nail preparation containing 6% Dibutyl Phthalate was not irritating in a 4 wk controlled use study with 47 subjects. A deodorant containing 4.5% Dibutyl Phthalate was not irritating in an antiperspirant efficacy test on 43 subjects. The same formulation was slightly irritating in a 21-d cumulative irritancy test on 12 subjects when tested neat. Diethyl Phthalate was not irritating in a primary irritation study with 10 subjects (concentration not reported); little to no irritation was observed to undiluted Diethyl Phthalate in another primary irritation study with 45 subjects.⁸

In a modified maximization test in 25 subjects, no contact sensitization was observed to a nail polish containing 9% Dibutyl Phthalate.² A deodorant containing 4.5% Dibutyl Phthalate was considered non-sensitizing in a modified human repeated-insult patch test (HRIPT) in 159 subjects. No sensitization was observed in HRIPTs with nail enamels that contained 2.74 or 5.32% Dibutyl Phthalate (105 and 25 subjects, respectively).⁸ No sensitization was observed in several HRIPTs with up to 100 subjects and undiluted Diethyl Phthalate. Sensitization was also not reported in an HRIPT with Dimethyl Phthalate (concentration not reported) in 16 subjects.

Dermal irritation and sensitization studies are summarized in Table 9. Undiluted Dibutyl Phthalate and Diethyl Phthalate were not irritating to rabbit skin.^{5,6} When induced topically at 75% and challenged at 50%, Dibutyl Phthalate was not sensitizing in a guinea pig maximization test (GPMT).⁵ Diethyl Phthalate at up to 100% was not sensitizing in guinea pigs.⁶ Diethyl Phthalate was also not sensitizing in a local lymph node assay (LLNA) when tested at up to 100%. A mouse ear swelling test was used to determine the adjuvant effect of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate (1:1 v/v with acetone; no further details).¹¹⁴ The mice were sensitized with FITC. A strong enhancement of ear-swelling response was observed in mice treated with Dibutyl Phthalate, while the response was not as strong with Diethyl Phthalate and Dimethyl Phthalate.

Photosensitization/Phototoxicity

Human

Dibutyl Phthalate (6% in a nail preparation) was non-irritating, non-sensitizing, and non-photosensitizing in a prophetic patch test with 99 subjects and in a modified HRIPT with irradiation at challenge in 48 subjects.² While some primary irritation was observed, 25% Diethyl Phthalate in a fragrance material was not sensitizing or photoallergenic in photoallergy studies with 23-35 subjects.⁸

OCULAR IRRITATION STUDIES

In Vitro

Dibutyl Phthalate was used in a test to determine the utility of using human corneal endothelial cell line B4G12 to evaluate an in vitro model for eye irritancy testing.¹¹⁵ Cell proliferation and toxicity were assessed after exposure with up to 200 $\mu\text{mol/l}$ Dibutyl Phthalate ($\geq 98\%$ pure). Additionally, gene expression and secretion of inflammatory cytokines were evaluated. Decreased cell proliferation was observed (mean $\text{IC}_{50} = 57.8 \pm 14.8 \mu\text{mol/l}$ with 5% fetal calf serum supplement; $30.6 \pm 12.7 \mu\text{mol/l}$ without 5% fetal calf serum), as well as cell toxicity (lactate dehydrogenase leakage observed at $\geq 50 \mu\text{mol/l}$ without fetal calf serum; at 200 $\mu\text{mol/l}$ with fetal calf serum). A significant ($p < 0.05$) increased gene expression and cytokine cell secretion were observed for interleukin- 1β and interleukin-8. Increased interleukin-6 secretion was also observed.

Animal

Overall, undiluted Dibutyl Phthalate, undiluted Diethyl Phthalate, and nail formulations containing up to 9% Dibutyl Phthalate were relatively non-irritating to rabbit eyes.² An undescribed material containing 12.5% Diethyl Phthalate caused

severe conjunctival irritation, including chemosis and discharge in 3 rabbit eyes.⁸ Undiluted Dimethyl Phthalate may cause severe injury to the eyes of rabbits after prolonged contact.²

In an ocular study in 3 Vienna White rabbits (2 males and 1 female), 0.1 ml of undiluted Dibutyl Phthalate was instilled into the conjunctival sac of the right eye while the untreated left eye served as control.⁵ This study was performed in accordance with OECD TG 405. The eyes were not rinsed and were observed for 72 h. Minimal irritation was observed (only conjunctivae redness) and were reversible within 72 h. Dibutyl Phthalate was considered not an ocular irritant in this study.

CLINICAL STUDIES

Retrospective and Multicenter Studies

In 3 different studies, patch testing was conducted in subjects with allergies to glues and/or plastics (n = 173, 357, and 839, respectively).⁸ No allergic reactions were observed in subjects that were tested with 5% (w/w) Dibutyl Phthalate.

The German Contact Dermatitis Research Group performed a retrospective study on 251 metalworkers that had suspected metalworking fluid dermatitis in 2002 and 2003.¹¹⁶ Of the 251 patients, 199 were tested with a metalworking fluid series that included Dibutyl Phthalate in 5% pet. Negative results were observed in 197 patients and 2 had questionable reactions.

Case Reports

A chemical worker accidentally swallowed approximately 10 g of Dibutyl Phthalate.² The worker's symptoms included nausea, vomiting, dizziness, headache, pain and irritation in the eyes, conjunctivitis, and toxic nephritis. He recovered completely after 2 wk. A patient that exhibited acute swelling and erythema of the groin and genitals after using a cream to treat intertrigo had a positive patch test to Dibutyl Phthalate.⁸ Another patient exhibited anaphylactic shock from a drug capsule containing Dibutyl Phthalate; undiluted Dibutyl Phthalate induced a 3 cm diameter wheal that expanded over the entire arm in a prick test. Proper treatment of a human corneal burn caused by Dimethyl Phthalate resulted in healing within 48 h and no loss of vision.²

In a case report, a 65-yr-old man presented with an acute erythematous rash in his groin and anogenital region following use of a cream product to treat itching in the peri-anal region.¹¹⁷ There was no personal, family, or occupational history of note. Patch testing using the European stand and medicament series, the cream product, and constituents of the cream product was performed. Positive reactions were observed on day 4 to 5% Dibutyl Phthalate pet. (++), the cream product (++), 0.1% benzalkonium chloride aq. (+), balsam of Peru (++), and fragrance mix (++).

Occupational Exposure

The health status of 147 workers subjected to prolonged occupational exposure to mixtures of phthalate plasticizers (including Dibutyl Phthalate) was investigated; many workers had a moderately pronounced toxic polyneuritis.²

In a human biomonitoring study performed in Europe, assessed the exposure of electronic waste processing workers from 10 companies to phthalates, including Dibutyl Phthalate.¹¹⁸ This assessment including analyzing urine samples from 106 workers at the beginning and at the end of the work week, comparing the urine samples to those of 63 non-occupationally exposed controls, and analyzing settled floor dust collected on the waste processing plants premises. The median concentration of Dibutyl Phthalate metabolite in the workers was 14.1 µg/g creatinine, compared to 7.54 µg/g creatinine in the controls. However, no significant differences were observed between pre- and post-shift concentrations in the workers. The concentration of Dibutyl Phthalate in dust was correlated with the corresponding urinary metabolite in the workers. Workers that used respiratory protective equipment had significantly lower urinary concentration Dibutyl Phthalate.

Occupational exposure to Dibutyl Phthalate was assessed in 37 manicurists.¹¹⁹ Pre- and post-shift urine samples were obtained on a single workday, and the urine was analyzed for monobutyl phthalate, mono-3-carboxypropyl phthalate, and mono-isobutyl phthalate, metabolites of Dibutyl Phthalate. Work environments varied, with open windows and doors in less than half of the salons and exhaust hoods used in 22% of salons. Approximately 41% of workers used paper masks and 22% wore gloves during work. More than 75% of workers ate in the salon. All 3 metabolites were detected in workers at baseline, with the highest concentration detected in monobutyl phthalate (58.5 ng/ml overall, n = 37), followed by mono-isobutyl phthalate (10.7 ng/ml overall). Glove users had higher monobutyl phthalate concentration at baseline (91.6 ng/ml, n = 7) than those who did not use gloves, n = 27). A significant increase (p = 0.05) in specific gravity-adjusted monobutyl phthalate concentrations was observed across the work shift (after work concentration overall = 87.2 ng/ml). A small, but significant increase (p = 0.05) was observed in mono-3-carboxypropyl phthalate, but no significant changes were noted in mono-iso-butyl phthalate. In linear regression modeling, individual cross-shift change in urinary monobutyl phthalate concentration was not related to whether doors and/or windows were open. Manicurists in salons without local exhaust had a 54% increase in urinary monobutyl phthalate concentration across shift compared with a 7% decrease across shift for those with exhaust ventilation. Glove use was associated with a significant reduction (p = 0.04) in urinary cross-shift monobutyl phthalate concentration. The cross-shift mono-iso-butyl phthalate concentration increased 6% among manicurists who did not wear gloves and decreased 15% in manicurists wearing gloves; however, these findings did not reach statistical

significance. No associations between cross-shift mono-3-carboxypropyl phthalate concentration and exhaust or glove use were detected.

Other Clinical Reports

A high incidence of premature breast development (thelarche) in young girls (younger than 8-yr-old) of Puerto Rico.⁸ Serum samples from 41 thelarche patients and 35 control subjects in Puerto Rico were analyzed by GC-MS for detection of phthalate esters and other possible contaminants. No pesticides or pesticide metabolites were detected in the samples tested. Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were detected in the serum of thelarche patients but not the control subjects. There was no statistically significant evidence that the premature breast development was a consequence of phthalate exposure.

In a study of 5 healthy college students, semen samples were analyzed by GC-MS for concentrations of Dibutyl Phthalate and Dimethyl Phthalate, for which the respective average concentrations were 10 ± 1.5 and $59.0 \pm 6.5\%$ at the level of 10 ppm. Phthalate exposures were presumably from ambient levels in the environment. The concentrations of Dibutyl Phthalate were compared to the sperm densities of the semen samples. The plot of Dibutyl Phthalate concentration vs. sperm density gave two clusters of points. The two clusters were surmised to represent the subjects' differing abilities to metabolize Dibutyl Phthalate. For those with a lesser ability to metabolize Dibutyl Phthalate, the average concentration was 129 ppb, the correlation coefficient between sperm density and Dibutyl Phthalate concentration was -0.4, and the slope of the regression line was -0.7. The subjects with apparent greater ability to metabolize Dibutyl Phthalate had an average semen concentration of Dibutyl Phthalate of 51 ppb. The correlation coefficient between sperm density and Dibutyl Phthalate concentration for the latter cluster was -0.4, and the slope of the regression line was -0.6. The investigators concluded that in both clusters there was a negative correlation between sperm density and semen Dibutyl Phthalate concentration.

Semen samples from "normal" men (number not reported) were collected, maintained in culture, and exposed to various ranges of concentrations of Dibutyl Phthalate, Diethyl Phthalate, or Dimethyl Phthalate for 0, 2, or 18 h.⁸ After the phthalate exposures, sperm motility was measured by a computerized sperm analysis system. Results were presented as EC_{25} , or the concentration of each phthalate to reduce sperm motility parameter by 25%, compared to untreated control sperm. The respective EC_{25} values for 0, 2, and 18 h of Dibutyl Phthalate exposure on sperm motility were 1, 30, and 8 mM. The EC_{25} values for 0, 2, and 18 h of Diethyl Phthalate exposure on sperm motility were >3.3, 2.5, 0.7mM, respectively. The respective EC_{25} values for 0, 2, and 18 h of Dimethyl Phthalate exposure on sperm motility were >0.94, >.94, and 0.4 mM. Likewise, acute exposure to each of the phthalates produced dose-dependent reductions in sperm curvilinear velocity, straight line velocity, and linearity of sperm motion, and Dibutyl Phthalate was the most potent of the three phthalate diesters in inhibiting these parameters.

In a crossover-crossback prospective study, 73 men with inflammatory bowel disease treated with mesalamine medications were studied to determine if Dibutyl Phthalate (approximately 350 $\mu\text{g/kg/d}$) in mesalamine coating was associated with decreased semen parameters.¹²⁰ Men ($n = 26$) taking mesalamine without Dibutyl Phthalate at baseline (i.e. background exposure) crossed-over for 4 mo to Dibutyl Phthalate-containing mesalamine and then crossed-back for 4 mo, to the mesalamine without Dibutyl Phthalate. Another group of men ($n = 47$) that were taking mesalamine with Dibutyl Phthalate at baseline followed the opposite procedure. All participants provided up to 6 semen samples during the study: 2 at baseline, 2 at crossover, and 2 at crossback. Semen parameters in the men with no Dibutyl Phthalate in the baseline decreased after exposure to the Dibutyl Phthalate mesalamine exposure (crossover versus baseline), especially motility parameters. This continued to decrease further even after crossback to mesalamine without Dibutyl Phthalate. Cumulative carryover effect of Dibutyl Phthalate exposure (crossback versus baseline) was a decrease of percent total sperm motility by 7.61, percent progressive sperm motility by 4.23, and motile sperm count by 26.0%. No significant change during crossover or crossback was observed in men that started with the Dibutyl Phthalate-containing mesalamine.

ENVIRONMENTAL EXPOSURE

The semen samples of 58 men were analyzed for phthalates by gas-liquid chromatography.⁸ These samples were found to contain concentrations of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate at 0.89, 0.40, and 0.22 mg/kg, respectively. The phthalates apparently entered the body through general environmental exposure.

In a study done by the Centers for Disease Control and Prevention, phthalate monoesters were measured in urine samples from 289 adult humans aged 20 - 60-yr-old.⁸ The analyses were performed after the samples were treated with β -glucuronidase to release the phthalate monoesters from the glucuronide conjugates. High concentrations of monobutyl phthalate (mean = 41.5 ng/ml, max = 4670 ng/ml) and monoethyl phthalate (mean = 345 ng/ml, max = 16200 ng/ml). Women of child-bearing age, 20 to 40-yr-old, were found to have higher urinary monobutyl phthalate concentrations (46.9 $\mu\text{g/g}$ creatinine) than any other age/sex group (31.4 $\mu\text{g/g}$ creatinine). These data suggest that the subjects had been exposed to diester phthalates, but the source of the phthalate exposure was uncertain.

In a pilot study, urine samples were collected from 19 young children (17 Hispanic, 2 Caucasian; 14 boys, 5 girls; ages 11.8 to 16.5 mo).⁸ The samples were analyzed for the presence of phthalate monoesters after deconjugation by β -glucuronidase. Monobutyl phthalate and monoethyl phthalate were detected in the children's urine samples at respective

mean (\pm SD) concentrations of 117.4 ± 287.6 ng/ml and 184.1 ± 246.9 ng/ml, respectively. The levels of monobutyl phthalate fall into the 75th to 100th percentile of the levels found in adult urine in the previously described study. The primary care givers of the children were questioned about the parents' cosmetic and perfume use, the presence of a farm or pesticide worker in the household, use of plastic toys, the child's health, and diet habits. None of these parameters correlated significantly with urinary phthalate monoester levels.

The reproducibility of detecting urinary phthalate metabolites in first morning urine samples was studied in 46 women (ages 35-49) that provided the urine samples on 2 consecutive weekdays.⁸ The samples were analyzed for detection of 7 phthalate monoesters. Of the 7 phthalate monoesters sought, 4 were detected in samples from all subjects. The mean concentrations of monobutyl phthalate was 78 ng/ml (range 0.7 - 251 mg/ml); monoethyl phthalate was 260 ng/ml (range 58 - 1043 ng/ml); monobenzyl phthalate was 39 ng/ml (range 6 - 135 ng/ml); and monoethylhexyl phthalate was 16.5 ng/ml (range 1 - 143 ng/ml). Phthalate levels did not differ between the two sampling days, as determined by Pearson coefficient correlation tests.

The intrauterine exposure to Dibutyl Phthalate from environmental exposure was measured in the cord blood of 50 postpartum women in China using ultra-HPLC-tandem mass spectrometry and solid-phase extraction for the detection of the metabolites.¹²¹ The mean concentration of Dibutyl Phthalate metabolites was 24.7 μ g/, with a detection rate of 95.7%.

RISK ASSESSMENT

Margin of exposure (MOE) is a quantitative ratio calculated for cosmetic ingredients by dividing the point of departure (PoD) obtained for an ingredient in an animal experiment by the estimated systemic exposure dose (SED) for the ingredient in humans, generally according to US EPA and EC SCCS guidelines. An MOE value greater than 100 has traditionally been considered an indication of safety. The basis for this MOE value of 100 comes from two multiplication factors: a 10-fold factor for extrapolating data from test animals to human beings (interspecies extrapolation), and an additional 10-fold for differences among the human population (intraspecies extrapolation). Notably, the MOE value is sometimes referred to as the margin of safety (MOS) despite the parameters being definitionally different.

The EC SCCP performed an MOE calculation for Dibutyl Phthalate at the trace levels found in tested perfume products in 2007.²² The result of the calculation was 10,000. The MOE was based on a LOAEL of 2 mg/kg bw/d (from a rat developmental toxicity study) and an SED of 0.0002 mg/kg/d (calculated from 17.79 g/d (exposure to all cosmetic products) x 0.0014% (maximum trace level detected in perfume) x 1000 (unit conversion) x 5% (estimated worst-case maximum absorption rate)/60 (default human weight)). When the calculation was performed at a concentration of 100 ppm (0.01%) Dibutyl Phthalate, the MOE was calculated to be 1350.

CIR staff has prepared an MOE calculation for Diethyl Phthalate using the 2025 concentration of use survey performed by the Council (Table 10). The maximum reported use concentration was 0.15% for leave-on face and neck skin care products and 0.1% for rinse-off cleansing products. Using a conservative NOAEL of 150 mg/kg bw/d from a 16-wk oral study in rats^{6,8} and a dermal absorption rate of 5.5% determined from an in vitro human skin study,¹²² the calculated MOE was 28,301. When an NOAEL of 200 mg/kg bw/d for DART was applied, the resulting MOE was 37,735. Both values exceed the safety threshold of 100, indicating an adequate margin of safety.

The following parameters were used in the calculation:

- Σ Sum of each of the separate exposures = 5.25 mg/d + 0.481 mg/d = 5.731 mg/d
- Body weight (adult) = 60 kg
- Skin absorption = 5.5 % (Absorption of Diethyl Phthalate through human skin reached $3.9 \pm 1.2\%$ (mean \pm SD, n = 4) of the applied dose over 72 h when the skin was occluded and $4.8 \pm 0.7\%$ (mean \pm SD, n = 3) when the skin was unoccluded)
- NOAEL - 150 mg/kg bw/d (repeated dose toxicity): Diethyl Phthalate was administered in the diet of rat (dose at 0, 0.2, 1.0, and 5% in feed, groups of 15 rats of each sex) for 16 wk.
 - 200 mg/kg bw/d (maternal & developmental toxicity): Diethyl Phthalate was administered in the diet of rats (dose at 0, 0.25, 2.5 and 5% in feed) during gestation days 6 -15 (following OECD TG 414 - Prenatal Developmental Toxicity Study).
 - 222 mg/kg bw/d (reproductive toxicity): dose at 0 (control) 600, 3000 & 15,000 ppm; approximately 15 wk for male and 17 wk for female parents of the F₀ & F₁ generations (following OECD TG 416 – Two Generation Reproduction Toxicity Study)

$$\text{SED}_{\text{dermal}} = \frac{5.731 \text{ mg/d} \times 5.5 \%}{60 \text{ kg}} = 0.0053 \text{ mg/kg bw/d}$$

$$\text{MoE}_{\text{dermal}} = \frac{\text{NOAEL}_{\text{repeated dose}}}{\text{SED}_{\text{dermal}}} = \frac{150 \text{ mg/kg bw/d}}{0.0053 \text{ mg/kg bw/d}} = 28,301$$

EPIDEMIOLOGICAL STUDIES

Phthalate Exposure and Metabolic Health

In a prospective cohort study of 618 women in Mexico City, urinary concentrations of Dibutyl Phthalate metabolites (Σ DBP) measured during the second and third trimesters of pregnancy were associated with adverse long-term maternal metabolic outcomes assessed 4 - 5 and 6 - 8 yr after delivery.¹²³ Higher Σ DBP concentrations were positively associated with elevated plasma glucose increased glucose and insulin levels, insulin resistance, and glycosylated hemoglobin (HbA1c). These findings suggest that prenatal exposure to Dibutyl Phthalate may contribute to persistent metabolic dysfunction in mothers, including impaired glucose regulation and adverse lipid alterations, years after pregnancy.

Phthalate Exposure and Neurodevelopment

Several epidemiological studies have explored the relationship between phthalate exposure and neurodevelopmental outcomes in children. While numerous studies suggesting a potential link between gestational/childhood phthalate exposure and neurodevelopment, a review of the broader epidemiological literature concluded that the associations between specific phthalates and behavioral domains are inconsistent, likely due to differences in the behavioral and cognitive domains assessed at different ages and the varied timing of exposure measurements across studies.¹²⁴

A prospective cohort study of 417 Korean mother-infant pairs found that higher maternal urinary concentrations of diethylhexyl phthalate and Dibutyl Phthalate metabolites during late pregnancy were associated with lower Mental and Psychomotor Developmental Index scores in infants at 6 mo of age.¹²⁵ These effects were especially prominent in boys.

In a Cincinnati-based cohort study of 350 mother-infant pairs, higher maternal urinary concentrations of Dibutyl Phthalate metabolites at 26 wk of gestation were associated with improved behavioral organization in 5-wk-old infants.¹²⁶ This effect was reflected by decreased arousal, increased self-regulation, and reduced handling needs on the NICU Network Neurobehavioral Scale.

A prospective study assessed the exposure of phthalates to 30 preterm neonates who received total parenteral nutrition (TPN) during their stay in the neonatal intensive care unit and the risk of neurodevelopmental delays at the age of 2 mo.¹²⁷ Urine samples for analyzing phthalate metabolites were obtained at admission and then daily until the last day of receiving TPN. The phthalate content in the daily TPN was also analyzed. The neurodevelopment of the neonates was assessed using a developmental questionnaire at 2 mo of age. Approximately 26% of the samples (n = 195) of TPN contained Dibutyl Phthalate above the method detection limit of 0.169 $\mu\text{g/l}$. In the neonates, all urine samples collected at admission and before discharge contained monobutyl phthalate above the method detection limit of 1.758 $\mu\text{g/l}$. The mean urinary level of monobutyl phthalate was 66.7 (standard deviation 101.8) $\mu\text{g/l}/\mu\text{mol/l}$ at admission and 44.5 $\mu\text{g/l}/\mu\text{mol/l}$ (standard deviation 40.7) before discharge. The mean molar sum of the metabolites of Dibutyl Phthalate was calculated to be 0.615 $\mu\text{mol/l}$ (standard deviation 1.254) at admission and 0.398 $\mu\text{mol/l}$ (standard deviation 0.181) before discharge. The daily dose of Dibutyl Phthalate and several other phthalates in the preterm neonates measured from their TPN was much lower than the recommended tolerable daily limit. However, these levels were not correlated with the phthalate parent compounds in the TPN solutions, suggesting other sources of exposure in the neonatal intensive care unit, such as devices, tubings, or staff/family members. The contribution of Dibutyl Phthalate in TPN to the preterm neonates' daily dose ranged between 0.002 and 1.427 $\mu\text{g/kg bw}$, much lower than the oral reference dose of 100 $\mu\text{g/kg bw}$. While negative association of neurodevelopment was observed with some of the phthalates tested, this association was not observed with Dibutyl Phthalate.

In another study assessing the effects of prenatal exposure of phthalates on neurodevelopment in infants age 4.5 and 7.5 mo, 5 first-morning urine samples were collected from pregnant women in Illinois.¹²⁸ Biomarkers included in the assessment were sum of Dibutyl Phthalate (Σ DBP), monoethyl phthalate, sum of all phthalate metabolites (Σ All), and sum of anti-androgenic metabolites (Σ AA). Overall, anti-androgenic phthalates were associated with higher (i.e., better) scores; however, a one-unit increase in Σ DBP was associated with a 12% increase in problem solving scores in 4.5-mo-old females but a 85% decrease for 7.5-mo-old females. A one-unit increase in monoethyl phthalate was associated with poorer scores on several outcomes, which were sex- and timepoint-specific: a 52% decrease in personal-social score in 7.5-mo-old males, a 39% decrease in fine motor scores in 7.5-mo-old males, and a 6% decrease in fine motor scores in 4.5-mo-old females. A one-unit increase in Σ All was associated with a 4% increase in personal-social scores in 4.5-mo-old males, but a 17% decrease in 7.5-mo-old males.

In a cross-sectional study of 667 Korean school-aged children, urinary concentrations of diethylhexyl phthalate metabolites were significantly and inversely associated with Full Scale and Verbal IQ scores, whereas Dibutyl Phthalate metabolites showed no such relationship.¹²⁹ Both diethylhexyl phthalate and Dibutyl Phthalate metabolites were negatively associated with children's vocabulary subscores, but after adjusting for maternal IQ, only the association with diethylhexyl phthalate metabolites remained significant.

In the Mount Sinai Children's Environmental Health birth cohort in New York City, 188 children aged 4 - 7 yr were evaluated for behavioral and executive functioning.¹³⁰ Higher maternal urinary concentrations of Diethyl Phthalate and Dibutyl Phthalate metabolites during late pregnancy were associated with increased externalizing behaviors (aggression, conduct, and attention problems) and poorer emotional control and executive function scores. These behavioral domains are consistent with patterns observed in attention-deficit/hyperactivity disorder-like profiles. A follow-up study in the same cohort observed more autistic-like behaviors in children aged 7 - 9 yr with higher prenatal exposure to Diethyl Phthalate and Dibutyl Phthalate.¹³¹ In a cohort of 277 New York City mother-child pairs, higher prenatal urinary concentrations of Dibutyl Phthalate and butyl benzyl phthalate metabolites were associated with more internalizing behaviors (e.g., withdrawn and anxious/depressed symptoms) in 3-yr-old children.¹³²

A systematic review and meta-analysis of human epidemiological studies evaluated the potential association between prenatal and early-life phthalate exposure and neurodevelopmental outcomes.¹³³ Six phthalates, including Dibutyl Phthalate and Diethyl Phthalate, were reviewed. In total, 53 epidemiological studies were identified and evaluated for risk of bias and study sensitivity using predefined domain-based criteria covering exposure assessment, outcome measurement, confounding control, and statistical analysis. Evidence was synthesized separately by phthalate compound and neurodevelopmental outcome, and for studies assessing cognitive and motor development in children aged ≤ 4 yr, a random-effects meta-analysis was conducted. The primary outcome domains included cognition (14 studies), motor development (9 studies), behavioral outcomes including attention-deficit/hyperactivity disorder (20 studies), infant behavior (3 studies), and social behavior, including autism spectrum disorder (7 studies). Overall, the review concluded that there is no consistent pattern linking prenatal or early-life phthalate exposure with adverse neurodevelopmental outcomes. The authors noted that null or inconsistent findings may reflect methodological limitations, including exposure misclassification, differences in susceptibility windows, potential sex-specific responses, and confounding by phthalate mixtures.

Phthalate Exposure and Reproductive Health

The relationship between phthalate levels and semen parameters was analyzed in semen samples collected from 52 men in Shanghai, China.¹³⁴ Samples were analyzed for Dibutyl Phthalate, Diethyl Phthalate, and diethylhexyl phthalate. The 3 phthalates were detected in most of the samples, with levels ranging from 0.09 - 0.57 mg/l (mean 0.16 mg/l) for Dibutyl Phthalate and 0.13 - 1.32 mg/l (mean 0.47 mg/l) for Diethyl Phthalate. A significant positive association between liquefied time of semen and phthalate concentrations of semen was observed. The correlation coefficient was 0.475 for Dibutyl Phthalate and 0.456 for Diethyl Phthalate. There was no significant difference between phthalate concentrations of semen and sperm density or livability, though the correlation coefficients were negative.

In a study comparing semen of fertile ($n = 50$) versus infertile ($n = 130$) men in Lucknow, India, Dibutyl Phthalate was detected in greater than 80% of the samples from infertile subjects.⁵⁵ In men that were oligoasthenospermic (low sperm count and low sperm motility), the maximum concentration of Dibutyl Phthalate in semen observed was 13.47 $\mu\text{g/ml}$. The maximum concentration in semen observed in asthenospermic (low sperm motility alone) men was 4.11 $\mu\text{g/ml}$. Fertile men had a maximum concentration of Dibutyl Phthalate in semen at 0.80 $\mu\text{g/ml}$. A significant negative association between sperm motility and Dibutyl Phthalate levels was observed in oligoasthenospermic and asthenospermic men ($r = -0.25$, -0.20 ; $p < 0.01$, $p < 0.01$).

A study evaluated the relationship of preconception phthalate exposure (measured as metabolites) with fecundability and pregnancy loss in women.¹³⁵ Data were collected from the Effects of Aspirin in Gestation and Reproduction trial of 1228 women attempting pregnancy for up to 6 menstrual cycles and throughout pregnancy if one occurred. Urine samples collected at enrollment were tested for 20 phthalate metabolites, including monobutyl phthalate, monoethyl phthalate, and mono-methyl phthalate. Monobutyl phthalate had a fecundity odds ratio (FOR) of 0.82 (95% confidence interval (CI) 0.70-0.96), while the FOR mono-ethyl phthalate and mono-methyl phthalate were 0.99 (95% CI 0.86-1.13) and 0.92 (95% CI 0.84-1.01), respectively. No consistent associations were observed with pregnancy loss.

Phthalate Exposure and Pubertal Development

A large prospective cohort study of 1151 girls examined the association between childhood phthalate exposure and pubertal development one year later.¹³⁶ Girls with the highest urinary concentrations of low-molecular-weight phthalates, including Diethyl phthalate and Dibutyl Phthalate, had a slightly higher prevalence of stage 2+ breast/pubertal hair development compared with those with the lowest exposure (prevalence ratio = 1.06). In contrast, a cross-sectional study of 725 Danish girls reported delayed onset of pubertal hair development among those with higher urinary concentrations of phthalate metabolites, particularly from Dibutyl Phthalate and butyl benzyl phthalate.

Phthalate Exposure and Breast Cancer

A case-control study in northern Mexico (233 breast cancer cases and 221 controls) analyzed urinary concentrations of nine phthalate metabolites.¹³⁷ Monoethyl phthalate, a metabolite of Diethyl Phthalate, was significantly higher in cases than controls and was positively associated with breast cancer risk (OR = 2.20; 95% CI 1.33 - 3.63), particularly among premenopausal women (OR = 4.13; 95% CI 1.60 - 10.70). In contrast, monobenzyl phthalate and mono(3-carboxypropyl) phthalate showed inverse associations, suggesting higher exposure was linked to reduced breast cancer risk.

A nested case-control study within the Women's Health Initiative (419 post-menopausal breast cancer cases and 838 controls) measured 13 urinary phthalate metabolites, including monoethyl phthalate, collected over 1–3 years.¹³⁸ Multivariable analyses showed no significant associations between phthalate biomarker concentrations and breast cancer, regardless of hormone therapy use, body mass index, or tumor subtype. Overall, phthalate exposure was not linked to increased postmenopausal breast cancer risk.

In a cohort of 102 women, serum levels of several phthalates, including Dibutyl Phthalate, Diethyl Phthalate, butyl benzyl phthalate, and diethylhexyl phthalate, were detected in all participants.¹³⁹ Women with breast cancer—particularly those from Mexico City and those with HER2⁺ or triple-negative subtypes—showed significantly higher Dibutyl Phthalate and butyl benzyl phthalate concentrations than controls or ER⁺ patients, indicating a positive association between phthalate exposure and breast cancer incidence.

Phthalate Exposure and Genotoxic Effect

Semen and urine samples of 168 men recruited from the Massachusetts General Hospital Andrology Laboratory analyzed to investigate whether environmental phthalate exposure was associated with DNA damage in human sperm.¹⁴⁰ Eight phthalate metabolites, including monobutyl phthalate, monoethyl phthalate, and monomethyl phthalate were measured in urine by using HPLC and tandem mass spectrometry (HPLC-MS). Sperm DNA integrity was assessed using the neutral single-cell microgel electrophoresis assay (comet assay). VisComet image analysis software was used to quantify comet parameters, including comet length, percent DNA in tail (tail%), and tail distributed moment (an integrated measure reflecting both comet length and DNA intensity in the tail). Results showed that for each interquartile range increase in specific gravity-adjusted urinary monoethyl phthalate, comet extent increased significantly by 3.6 μm (95% CI: 0.74 - 6.47), and increased by 1.2 μm (95% CI: -0.05 - 2.38). No significant associations were observed for monobutyl or monomethyl phthalates with any comet assay parameters.

Phthalate Exposure and Endocrine Effect

The association of maternal urinary phthalates with markers of the thyroid system was studied in 1996 women.¹⁴¹ Urinary concentration of phthalate metabolites, including monobutyl phthalate, and serum concentration of thyroid stimulating hormone, free and total thyroxine (FT4 and TT4), and free and total triiodothyronine (FT3 and TT3) were measured in pregnant women in early pregnancy (median gestational week 10). Monobutyl phthalate was associated with lower T4/T3 (free and total) ratios and higher FT4/TT4 and FT3/TT3 ratios.

A prospective Danish-Finnish cohort study conducted from 1997 to 2001 analyzed individual breast milk samples of mothers to male infants collected 1 - 3 mo postnatally for 6 phthalate monoesters, including monobutyl phthalate, monoethyl phthalate, and monomethyl phthalate.¹⁴² Of the 130 mothers studied, 62 were cryptorchid males infants and 68 were healthy males. Serum samples of 74% of all males were also analyzed for gonadotropins, sex-hormone binding globulin, testosterone, and inhibin B. The phthalate monoester content in breast milk ranged from 0.6 - 10,900 $\mu\text{g/l}$ for monobutyl phthalate, 0.07 - 41.4 $\mu\text{g/l}$ for monoethyl phthalate, and < 0.01 - 5.53 $\mu\text{g/l}$ for monomethyl phthalate. No association was found between phthalate monoester levels and cryptorchidism. Positive correlation with sex-hormone binding globulin was observed with monoethyl phthalate and monobutyl phthalate ($r = 0.323$, $p = 0.002$ and $r = 0.272$, $p = 0.01$, respectively); and monomethyl phthalate, monoethyl phthalate, and monobutyl phthalate was positively correlated with luteinizing hormone:free testosterone ratio ($r = 0.21$ – 0.323 , $p = 0.002$ – 0.044). Monobutyl phthalate was negatively correlated with free testosterone ($r = -0.22$, $p = 0.033$).

SUMMARY

According to the *Dictionary*, Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate are reported to function as fragrance ingredients, plasticizers, and solvents in cosmetic formulations. Diethyl Phthalate also functions as a denaturant. The Panel first reviewed the safety of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate in a report that was published in 1985, with the conclusion that these ingredients are “safe for topical application in the present practices of use and concentration in cosmetics.” This conclusion was reaffirmed in re-reviews that were published in 2005 and 2017. This report was placed on the 2024 CIR Priorities List following nomination by the US FDA for cause.

According to RLD that CIR received in 2024, Diethyl Phthalate is used in 168 formulations, with most of the uses reported in fragrance preparations. Additionally, the RLD reported Dibutyl Phthalate is used in 2 manicuring preparations and Dimethyl Phthalate had no uses. The results of the concentration of use survey conducted by the Council in 2025 indicate Diethyl Phthalate has a maximum concentration of use range of 0.1-0.15%, with 0.15% reported in leave-on face and neck products. No concentrations of use were reported for Dibutyl Phthalate or Dimethyl Phthalate; however, responses to the survey indicated that Dibutyl Phthalate and Diethyl Phthalate may be present in cosmetics as impurities.

When determining whether to re-open this safety assessment, the Panel considered FDA VCRP survey data submitted to CIR in 2023 as compared to that stated in the previous reports. In 2023, Diethyl Phthalate was reported to be used in 1 skin care formulation; no uses were reported for Dibutyl Phthalate or Dimethyl Phthalate. When comparing the VCRP data received in 2023 to that received in 2001, the frequencies of use for these phthalate ingredients have greatly decreased since the 2005 re-review was published; Dibutyl Phthalate was reported to have 150 uses (most in manicuring preparations), Diethyl Phthalate was reported to have 73 uses (most in fragrance preparations), and Dimethyl Phthalate was reported to have

12 uses (most in non-coloring hair preparations). In the 2005 re-review, the maximum concentration of use range for Dibutyl Phthalate was 0.0038 - 15% (15% reported in manicuring preparations). Diethyl Phthalate was reported to have a maximum concentration of use range of 0.00003 - 11% (11% was reported in perfumes), and Dimethyl Phthalate was reported to have a maximum concentration of use range of 0.00002 - 2% (2% was reported in hair spray).

Under European regulations for cosmetic ingredients, Dibutyl Phthalate is listed in Annex II, list of substances prohibited in cosmetic products. Diethyl Phthalate and Dimethyl Phthalate are not restricted from use in any way under the rules governing cosmetic products in the European Union. In 2004, the SCCNFP determined that Dibutyl Phthalate should not be intentionally added to cosmetic products. Dibutyl Phthalate has been classified as a Category 2 toxic substance to reproduction that “may cause harm to the unborn child.” In 2002, the SCCNFP opined that the safety profile for Diethyl Phthalate supported its use in cosmetic products; this opinion was reaffirmed by the SCCNFP in 200321 and the SCCP in 2007.

Globally, Dibutyl Phthalate has been detected in commercial PET bottled water, and in the US, Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate have been detected in feminine hygiene products (pads, panty liners, and tampons). In 2012, the FDA issued guidance for industry on limiting the use of Dibutyl Phthalate as an excipient in regulated drug and biologic products, recommending that use of this ingredient be avoided if an alternative is available. Dibutyl Phthalate (> 1000 ppm or 0.1%) has been prohibited in the manufacturing, sale, distribution, or importation of children’s toys and childcare articles. Diethyl Phthalate is a plasticizer in food packaging material and an authorized denaturant.

In an in vitro study, the percutaneous absorption of ^{14}C -Dibutyl Phthalate in the receptor fluid over 24 h was $27.1 \pm 1.9\%$, with $29.2 \pm 2.4\%$ remaining in the skin (total skin penetration over 24 h = $56.3 \pm 2.7\%$). Over 72 h, the amount of Dibutyl Phthalate in the receptor fluid was $59.9 \pm 3.2\%$, with $13.7 \pm 3.1\%$ remaining in the skin (total skin penetration over 72 h = $75.6 \pm 4.7\%$). A percutaneous absorption study of Dimethyl Phthalate and 2 other phthalates using different carriers, determined the permeation rate of the ointment, base cream, and lotion containing Dimethyl Phthalate to be 1.391, 1.216, and $1.120 \mu\text{g}/\text{cm}^2/\text{h}$, respectively. Of the 3 phthalates tested, Dimethyl Phthalate had the fastest permeation rate, regardless of carrier type. The 9-h percutaneous penetration of 200 or 500 ng/ml Diethyl Phthalate and Dibutyl Phthalate in human keratinocytes for Diethyl Phthalate was 62 and 39%, and for Dibutyl Phthalate were 53 and 34%, for the low- and high-dose groups, respectively. The dermal absorption of ^{14}C -Dibutyl Phthalate in hairless guinea pigs was $62.0 \pm 2.0\%$ of the applied dose. Most of the Dibutyl Phthalate was found in the urine ($60.4 \pm 1.8\%$) and the amount systemically absorbed in the ovaries, kidneys, and liver were less than 2%. Approximately 2% of Dibutyl Phthalate remained on the skin.

Female mice were used in a study to determine the distribution of 0, 1, 10, or 1000 mg/kg Dibutyl Phthalate after single or repeated doses of via gavage either once or once daily for 10 d. Regardless of duration of exposure (single or repeated), monobutyl phthalate was detected in the sera and tissues of Dibutyl Phthalate-treated mice. In single dose mice, monobutyl phthalate levels peaked at < 6 h and fell close to background levels by 24 h post-treatment. Following the last repeated dose, monobutyl phthalate levels peaked at < 2 h and fell to background levels by 12 h.

In human dermal studies, increased excretion of monobutyl phthalate was observed after topical application of a cream containing 2% Dibutyl Phthalate given as daily whole-body topical applications for 5 d. Total monobutyl phthalate excreted in the urine during the treated week was $11.8 \pm 0.6 \text{ mg}/24 \text{ h}$. On average, 1.82% of the applied Dibutyl Phthalate was recovered in urine as monobutyl phthalate. High levels of Diethyl Phthalate and Dibutyl Phthalate metabolite excretions were observed in a controlled chamber study with males exposed to the airborne phthalates for two 6-h sessions (one “hood-on” and the other “hood-off”). Modeling of the urinary metabolite data estimated that the daily dermal uptake ($0.14 \mu\text{g}/\text{kg bw}/\text{d}$) exceeded the inhalation intake ($0.042 \mu\text{g}/\text{kg bw}/\text{d}$) by roughly threefold. In an oral study, an individual male ingested $60 \mu\text{g}/\text{kg}$ of labeled Dibutyl Phthalate in decaffeinated coffee served in an edible cup. The majority of the dose (92.2%) was excreted in the first 24 h, with only < 1% of the dose excreted on day 2. The major metabolite was monobutyl phthalate (84%). Various side chain oxidized metabolites accounted for approximately 8%.

In oral rat studies, the LD_{50} for Dibutyl Phthalate was 6279 mg/kg bw for both sexes. The LD_{50} for Diethyl Phthalate was > 5 ml/kg bw for both male and female rats. For Dimethyl Phthalate, the combined LD_{50} in male and female rats was $5.1 \text{ g}/\text{kg bw}$. In inhalation studies with rats, the LC_{50} value for Dibutyl Phthalate in both sexes was estimated to be > 15.68 mg/l, and the LC_{50} for Diethyl Phthalate in both sexes was > 4.64 mg/l.

In a 28-d study, no clinical signs of toxicity were observed in rats that received up to 7000 ppm Dibutyl Phthalate in feed, with or without the addition of 3000 ppm diethylhexyl phthalate. Relative liver weights were increased in males and females of the 7000 ppm Dibutyl Phthalate group and in males of the 1000 ppm Dibutyl Phthalate group. In the groups that received both phthalates, an increase of absolute and relative liver weights were observed in all animals in all combined treatment groups. Mice that received $40 \text{ mg}/\text{kg bw}/\text{d}$ Dibutyl Phthalate for 60 d via gavage had no significant differences in body weight gain or lung function parameters when compared to saline controls. However, oxidative stress biomarkers in lung tissue were significantly altered. Serum total IgE levels showed a statistically significant increase compared to controls, whereas cytokine levels and inflammatory cell counts in BAL fluid were not significantly different from controls. In a 9-wk oral gavage study, male rats treated with Dibutyl Phthalate at $750 \text{ mg}/\text{kg bw}/\text{d}$ had marked histopathological changes in the lungs, including increased inflammatory cell infiltrations, epithelial cell shedding, and mucus secretion. Hematological analysis showed significant elevations in white blood cells, neutrophils, lymphocytes, eosinophils, and platelets. Compared

to controls, the Dibutyl Phthalate group exhibited a significant increase in inflammatory cells in rat BAL fluid, accompanied by elevated Th2 cytokines. Oxidative stress was evident, and increased apoptosis was observed in lung tissue.

The NOAEL for male and female rats that received up to 3170 mg/kg/d (5%) Diethyl Phthalate in feed for up to 16 wk was 150 mg/kg/d (0.2%). Male and female rats that received 5% Diethyl Phthalate for 16 wk had statistically significantly lower absolute weights of the brain, heart, spleen & kidneys than controls; reductions in the absolute weights of the female gonads and male heart, spleen, and kidneys were also observed in the 5% groups after 2 and 6 wk of treatment. Increased kidney weights were seen in males & lower pituitary weights in females in the 1% group after 2 wk of treatment. Increased gonad weights were also seen in females of the 1% group after 6 wk. No mortalities were observed in a 4-wk study in male rats that received Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate, the respective monoesters, or phthalic acid in corn oil via gavage (500 mg/kg bw/d for diesters, 250 mg/kg bw/d for monoester and acid). Livers weights were significantly increased in groups treated with Dibutyl Phthalate and monobutyl phthalate compared to control group. Red blood cells and hematocrit were significantly lower and mean corpuscular hemoglobin concentration and platelet count were significantly higher in the Dibutyl Phthalate group (the monoester also had significantly higher platelet counts), and hemoglobin level was reduced only in the Dimethyl Phthalate group.

Zebrafish were exposed to Dibutyl Phthalate (up to 10 mg/l) for 96 h. The LC_0 was 1.3 mg/l and the LC_{100} was 5 mg/ml.

The NOEC for malformations was 5.8 ppm for Dibutyl Phthalate (tested at up to 15 ppm) and 17 ppm for Diethyl Phthalate (tested at up to 200 ppm) in in vitro 96-h frog embryo teratogenesis assays using *Xenopus laevis* embryos. The mean 96-h LC_{50} s for Dibutyl Phthalate and Diethyl Phthalate in this study were 12.88 and 64.5 ppm, respectively. Malformations were observed starting at 5 ppm in Dibutyl Phthalate-treated embryos and starting at 50 ppm in Diethyl Phthalate-treated embryos. Additional studies in *Xenopus laevis* with Dibutyl Phthalate also noted defects, with mean percent normal development decreased as concentration and exposure time increased. No significant effects on survival or hatching of zebrafish embryos were observed following exposure to up to 250 μ g/l Dibutyl Phthalate; however, body length decrease, yolk sac abnormalities, and immune responses were observed. An acute developmental toxicity of Dibutyl Phthalate (up to 0.2 mg/l), Diethyl Phthalate (up to 10 mg/l), and Dimethyl Phthalate (up to 8 mg/l) in zebrafish saw a malformation rate of 30% with 0.05 mg/l Dibutyl Phthalate (deaths were observed at higher concentrations). Dimethyl Phthalate was not correlated with larvae mortality. In a dermal study in mice, Diethyl Phthalate (tested at up to 5600 mg/kg/d) applied to pregnant mice on gestation days 0 - 17 had a NOAEL < 500 mg/kg/d in dams and 1600 mg/kg/d in offspring. The dams exhibited dose-related abnormal behavior and reduced thymus and spleen weights were observed at all doses; while significantly reduced fetal body weight and an increased incidence of cervical and lumbar rib variations were observed in 5600 mg/kg/d offspring. A 3-generation reproductive and developmental study in mice with Dibutyl Phthalate in olive oil up to 2000 mg/kg bw induced skeletal malformations in F_0 males, with significant mortality in postnatal life and skewing the sex ratio. An increased frequency of DNA damage in the germ cells was observed in F_1 males; exposure of F_0 males to Dibutyl Phthalate did not affect F_1 male fertility and pregnancy frequency. In multigeneration studies with Dibutyl Phthalate in rats at up to 10,000 ppm in dietary feed, no adverse effects to body weights and body weight gains were observed in parental dams. In one study, male offspring in the 10,000 ppm dose group had high incidence of small or absent reproductive organs and undescended testes, gross lesions correlated with microscopic lesions in the male reproductive organs. In another feed study, anogenital distance was significantly decreased in male F_1 and F_2 offspring in the 10,000 ppm dose groups. F_1 males of the 10,000 ppm dose group also had a statistically significant mean delay in preputial separation, testicular descent, and seminiferous tubular atrophy. In a 2-generation rat feed study with Diethyl Phthalate at up to 15,000 ppm, the NOAEL for parental animals was 15,000 ppm, and the NOAEL for development and growth of pups was 3000 ppm. Body weight gains before weaning were inhibited in F_1 and F_2 pups and vaginal openings were slightly delayed in F_1 females at 15,000 ppm.

Mouse ovarian antral follicles treated with 1000 μ g/ml Dibutyl Phthalate in DMSO for 24 h were significantly smaller than controls, and follicles treated with concentrations greater than 100 μ g/ml were significantly reduced after 72 h. No toxicity was observed to the follicles after treatment with monobutyl phthalate at the same concentrations for the same periods of time. Dibutyl Phthalate in tocopherol-stripped corn oil did not cause significant deviations in the expression of transcripts encoding IGF binding proteins in female mice that received up to 1000 mg/kg/d for 20 d when compared to the vehicle control. Total number of ovarian follicles counted per ovary was reduced in the 100 μ g/kg/d group when compared to controls; however no significant differences were observed in the 10 μ g/kg/d or 1000 mg/kg/d dose groups. Significantly low primordial follicle counts observed in the 100 μ g/kg/d group and in the 1000 mg/kg/d group compared to controls, also fewer primary follicles were observed in 100 μ g/kg/d mice. In another female mouse study with Dibutyl Phthalate in tocopherol-stripped corn oil (up to 1000 μ g/kg/d for 30 d), dose-dependent effects on folliculogenesis and gene expression were observed. At 1000 μ g/kg, more atretic follicles were observed in the ovaries. In a dose-dependent manner, Dibutyl Phthalate significantly reduced the expression of genes responsible for homologous recombination, mismatch repair, and nucleotide excision repair. Pregnant rats that received 500 mg/kg Dibutyl Phthalate in corn oil every second day starting at gestation day 14.5 through postnatal day 6 exhibited no adverse effects and delivered exclusively live pups. Offspring also exhibited no adverse effects, and female pups had no effects to anogenital distance or adverse effects in vaginal opening. In an oral uterotrophic assay, female rats that received 100 mg/kg Dibutyl Phthalate in corn oil from postnatal day 21 for 3 d

had significantly decrease uterine wet weight and minor variations in ovary wet weight; no abnormal clinical signs or symptoms were observed in any of the treated or control animals. Female rats from a pubertal onset assay that were treated with 10 or 100 mg/kg Dibutyl Phthalate from postnatal day 21 for 20 d had significantly reduced uterus and ovary weights in both treated groups. However, vaginal opening was not observed in any of the animals in the controls and treated groups until postnatal day 42, except in one animal each in vehicle control and the 100 mg/kg dose group. In an oral multigeneration study in which the F₀ pregnant mice received a mixture of phthalates that included 15% Dibutyl Phthalate and 35% Diethyl Phthalate in tocopherol-stripped corn oil at up to 500 mg/kg/d, female offspring of the F₂ generation had increased uterine weight, anogenital distance and body weight. Cystic ovaries and breeding and pregnancy complications were also observed. Similar effects were observed in the females of the F₃ generation.

In an in vitro mechanistic study using TM3 and TM4 cells, exposure to Dibutyl Phthalate at 0, 5, 10, 50, and 100 mg/l for 24 h produced dose-dependent evidence of ferroptosis. Dibutyl Phthalate treatment decreased cellular GSH and mitochondrial membrane potential, and increased MDA, ROS, GSSG, and Fe²⁺ levels. Dibutyl Phthalate selectively upregulated PRDX6, a negative regulator of ferroptosis, while GPX4 remained unchanged. Further, Dibutyl Phthalate upregulated SP1 expression, which can directly bind to the PRDX6 promoter and transcriptionally activated its expression. Rat testicular explants exposed in vitro to Dibutyl Phthalate at 10⁻⁶, 10⁻⁵, and 10⁻⁴ M for 24 h showed significant upregulation of Notch1, Dll4, and Hey1 at both mRNA and protein levels, while Hes1 expression remained unchanged. Dibutyl Phthalate altered immunoexpression of activated NOTCH1, DLL4, HEY1 and HES5 both in seminiferous epithelium and interstitial tissue, with differential effects across cell types. In rat Leydig cells exposed to 50 mg/l Dibutyl Phthalate, Dibutyl Phthalate + prostaglandin E2, or flutamide for 24 h, expression of testosterone was significantly decreased in all groups. A significant decrease of Cx43 was observed in the Dibutyl Phthalate group. Human sperm exposed to up to 134.7 µg/ml Dibutyl Phthalate for 30 min to up to 96 h had a concentration- and duration-dependent decrease in motility. In another study with human sperm, Dibutyl Phthalate at 6 µM, monobutyl phthalate at 3 µM, and a mixture of both had adverse effects on sperm motility, penetration ability, and capacitation following exposure for 1 to 4 h.

Multiple oral studies have investigated the effects of Dibutyl Phthalate in the male rodents. The NOAEL was 50 mg/kg bw/d for developmental toxicity in a study in pregnant rats that received up to 500 mg/kg bw/d on gestation day 1 through postnatal day 21. At doses of 250 mg/kg and higher, reduced birth weight, body weight gain, number of live pups per litter, anogenital distance in males, epididymis weight, and sperm count and motility were observed. Studies in which pregnant rats received up to 600 mg/kg bw/d Dibutyl Phthalate between gestation days 12 - 21 found cryptorchidism, infertility, hypospadias, decreases in anogenital distance, and/or testes abnormalities. Closer examination of the focal dysgenesis of the testes found adverse effects in the Leydig and Sertoli cells. One study of rats that received Dibutyl Phthalate (500 mg/kg/d) from gestation day 12 through gestation days 16 - 20 saw a significantly increased incidence of MNG in fetal rat testes following in utero exposure to the test material. Sertoli cells exhibited retracted apical processes, disorganized vimentin cytoskeleton, and abnormal contacts with gonocytes. These morphological changes were no longer evident after birth and cessation of exposure. Pregnant dams that received Dibutyl Phthalate (up to 600 mg/kg/d) or Diethyl Phthalate (up to 900 mg/kg/d) on gestation days 8 - 18 experienced no adverse effects, but fetal testicular testosterone production was significantly reduced at doses of 300 mg/kg/d or higher in Dibutyl Phthalate male fetuses. This effect was not observed with Diethyl Phthalate. A phthalate mixture containing Dibutyl Phthalate significantly increased fetal mortality at 40% of top dose (containing 120 mg/kg/d Dibutyl Phthalate) and above in dams that received the test material on gestations days 8 - 18, and testosterone production was reduced in a dose-additive manner starting at 20% of top dose (containing 60 mg/kg/d Dibutyl Phthalate) and above. Phthalate-induced underdevelopment of the testes was observed at 100% of the top dose (300 mg/kg/d). Testes weights were significantly reduced in male rats that received Dibutyl Phthalate (500 mg/kg bw/d) for 4 wk, and Dibutyl Phthalate, monobutyl phthalate (250 mg/kg bw/d), and monoethyl phthalate (250 mg/kg bw/d) significantly lowered sperm counts and sperm motility of epididymal sperm. These effects were not observed with Diethyl Phthalate or Dimethyl Phthalate (500 mg/kg bw/d for each diester), monomethyl phthalate (250 mg/kg bw/d), or phthalic acid (250 mg/kg bw/d).

Adult male rats that received up to 600 mg/kg/d Dibutyl Phthalate in corn oil for 15 d had significantly decreased testicular weight at each dose tested, and sperm count and motility were significantly decreased in a dose-dependent manner. Serum follicle-stimulating hormone, testosterone levels, and testicular lactate dehydrogenase activities were significantly decreased at all doses tested. In a study of male rats exposed to 250 mg/kg Dibutyl Phthalate, with or without BaP for up to 12 wk, no adverse effects were observed in the relative weights of the testes and epididymides, but vacuolization of Sertoli cells was observed in Dibutyl Phthalate-exposed rats after 12 wk. Male rats orally exposed to Dibutyl Phthalate (500 mg/kg/d) for 3 wk, had decreased germ cell layer thickness, and reduced sperm density in testicular tissue. These morphological changes were associated with increased testicular levels of MDA, Fe²⁺, and GSSG, together with decreased reduced GSH and serum testosterone. No gross lesion were observed in mice that received up to 1393 mg/kg bw/d Dibutyl Phthalate in a 2-yr study, but significantly increased incidences of germinal epithelium degeneration in the testes and exfoliated germ cells in the epididymal duct were observed. In rats that received 2000 mg/kg/d Diethyl Phthalate in dietary feed for 7 d, lower levels of testosterone were measured in the testes and serum of the treated animals, but no testicular damage was observed by microscopic examination.

Mice that received Dibutyl Phthalate (2.5 mg/kg/d), with or without diethylhexyl phthalate, for 40 d via implanted osmotic pumps had abnormal sperm morphology, particularly in the mixture group. Significant differences in sperm concentration were observed between control and exposed groups. No clinical signs of toxicity were observed in rats that were injected subcutaneously with Dibutyl Phthalate (up to 20 mg/animal) from postnatal day 5 -14; and no alterations in testicular descent were observed in any treated rats. However, significantly reduced testes, seminal vesicles, LABC, and Cowper's glands weights were observed in the 20 mg dose group when compared to the controls.

In additional studies, 100 mg/kg Dibutyl Phthalate injected into chicken eggs significantly reduced hatching and increased late hatching when compared to controls. Gross malfunctions and severe moto dysfunctions were observed in treated animals. Zebrafish that received up to 1133 µg/l Dibutyl Phthalate, with or without diisobutyl phthalate had a greater imbalance of the testosterone to estradiol ratio and severe structural damage from the mixture than to just Dibutyl Phthalate. These effects were consistent with the testis transcriptome analysis for which 4570 genes were differentially expressed in the mixture exposure, while 2795 genes were differentially expressed in Dibutyl Phthalate.

Dibutyl Phthalate (up to 2000 µg/plate) and Diethyl Phthalate (up to 5000 µg/plate) were not genotoxic in separate Ames tests, with and without metabolic activation. In a third Ames test, Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate were mildly genotoxic in *Salmonella typhimurium* strains TA100 and TA1535 without metabolic activation in a dose-dependent manner when tested at up to 2000 µg/plate. Dibutyl Phthalate (up to 80 µM) was genotoxic in micronucleus tests using cultured bovine lymphocytes and CHO cells. In chromosome aberrations tests, Dibutyl Phthalate (up to 30 µM) caused a higher occurrence of chromosomal aberrations in CHO cells at all tested concentrations when compared to the vehicle control; however, Diethyl Phthalate (up to 1780 µg/ml) was not genotoxic to human lymphocytes. Diethyl Phthalate (up to 771 µg/ml) was also not genotoxic in a gene mutation assay in mouse L5178 TK +/- lymphoma cells. Genotoxicity was observed to Dibutyl Phthalate when tested in human mucosal cells and human peripheral lymphocytes at up to 354 µmol/ml and in cultured bovine lymphocytes at up to 80 µM. In an in vivo micronuclei study with tilapia, mean micronuclei frequencies in both Dibutyl Phthalate at 10 mg/l and the positive control groups were significantly different with respect to control and solvent control groups following both 24 and 96 h exposures.

The potential for Dibutyl Phthalate (up to 10 µM) to promote tumors was studied in normal bladder epithelial cells and human bladder cancer cells. A decrease in migration of the bladder epithelial cells was observed as the Dibutyl Phthalate concentration increased, while migration of the bladder cancer cells increased with increased Dibutyl Phthalate. FOSB, JUND, ATP6V1C2, and RHOQ were differentially expressed in the second-generation sequencing before and after Dibutyl Phthalate treatment. FOSB expression was confirmed, and overexpression of FOSB increased both proliferation and invasion in the bladder cancer cells. In human breast epithelial cells and breast cancer stem cells, exposure to Dibutyl Phthalate (up to 1 µM) induced expression of mesenchymal markers vimentin and CD90 and decreased expression of epithelial markers cytokeratin 7 and E-cadherin and enhanced tumorsphere formation ability and higher levels of cancer stem cell markers (CD133, CD44, ALDH1A1, OCT-4, Nanog).

In a 2-yr dietary study in mice and rats performed by the NTP, no evidence of carcinogenic activity of Dibutyl Phthalate in male or female mice at up to 10,000 ppm. However, there was equivocal evidence of carcinogenic activity of Dibutyl Phthalate in male rats based on the marginal increase in the incidence of pancreatic acinus adenomas at up to 10,000 ppm; there was no evidence of carcinogenic activity of Dibutyl Phthalate at up to the same concentration in female rats.

Dibutyl Phthalate treatment of MCF-10A normal breast cells co-cultured with mammary fibroblasts led to increased cell viability, decreased apoptosis, and enhanced cell-cycle progression (S and G2/M phases), accompanied by upregulation of cyclins/CDKs and PI3K/AKT/mTOR/BCL-2 signaling and downregulation of pro-apoptotic proteins.

Oxidative stress, lipid peroxidation, and DNA damage occurred in zebrafish liver according to changes in antioxidant enzymes, MDA, and 8-OHdG content following exposure to up to 2 mg/l Dibutyl Phthalate for up to 28 d. AChE activity was always active and negatively correlated with the Dibutyl Phthalate concentration. The expression of Cu/Zn-sod and gpx genes were similar to that of antioxidant enzymes from 7 to 21 d, while in the end, the inconsistent result appeared due to the time lag effect in protein modification, gene transcription and translation. In another zebrafish study using to assess gene expression, DNA damage, and oxidative stress biomarkers following exposure to Dibutyl Phthalate (up to 2 mg/l) for up to 28 d, Dibutyl Phthalate significantly stimulated SOD and CAT activities, increasing MDA and 8-OHdG contents. On day 28, AChE inhibition rates for 0.08, 0.4, and 2 mg/l were 13.4, 11.9, and 14.7%. The trend of Cu/Zn-sod gene variation was consistent with SOD activity, showing "inhibition-activation-inhibition." Hepatic and ovarian expression of Lpl, Ldh1a1, Adh1, Ugt1a6a, and Cyp1b1 were altered in a time- and dose-specific manner in mice treated with up to 1000 mg/kg Dibutyl Phthalate either once or once daily for 10 d. Dibutyl Phthalate exposure induced oxidative stress, lipid peroxidation, and DNA damage in zebrafish liver, as evidenced by increased MDA and 8-OHdG levels and altered antioxidant enzyme activities (SOD, CAT, GPx) and Cu/Zn-sod gene expression. In zebrafish, AChE activity was inhibited in a concentration-dependent manner, while in mice, hepatic and ovarian expression of metabolic genes (Lpl, Ldh1a1, Adh1, Ugt1a6a, and Cyp1b1) was altered in a time- and dose-dependent manner following Dibutyl Phthalate exposure.

Monobutyl phthalate significantly reduced global DNA methylation and altered promoter-specific methylation and gene expression (TP53 and P16) in human blood mononuclear cells, while Dibutyl Phthalate showed more limited but distinct

methylation changes across genes of TP53, P16, CCND1 and BCL2. Both phthalates modulated epigenetic patterns and gene expression in a concentration-dependent and gene-specific manner.

Diethyl Phthalate significantly reduced testosterone concentrations and significantly increased the 17 β -estradiol/testosterone ratio in human adrenocortical carcinoma cells starting at 0.004 mM, with similar effects observed for its metabolite monoethyl phthalate at higher concentrations. Both Diethyl Phthalate and monoethyl phthalate downregulated steroidogenic genes StAR and 3 β HSD2, while upregulating CYP19A,

Diethyl Phthalate (up to 10 mg/l) reduced testosterone and 17 β -estradiol levels in male zebrafish and altered steroidogenic gene expression. In pregnant rats, combined exposure to Dibutyl Phthalate (750mg/kg/d, oral, GD 1-3) and STZ induced gestational diabetes-like effects with hyperglycemia, reduced insulin, and pancreatic FoxM1 suppression. In primary islet β cells, Dibutyl Phthalate decreased viability and increased apoptosis via pSTAT1 activation, reversible by STAT1 inhibition. Using the AOP framework, three key elements were identified for Dibutyl Phthalate-induced male reproductive toxicity: 1) fetal rats are most sensitive to anti-androgenic effects, 2) androgen-independent adverse outcomes are conserved across mammalian models and human fetal testis xenografts, and 3) anti-androgenic effects are also conserved across species when exposure occurs postnatally.

Dibutyl Phthalate caused concentration-dependent cytotoxicity across multiple cell types. In bovine lymphocytes and human sperm, MTT assays showed reduced viability with increasing concentration (10 - 100 μ M in lymphocytes; 13.47 - 134.7 μ g/ml in sperm) and exposure duration (up to 96 h). In human keratinocytes, viability declined from 85% to 37% across 0.1- 5 mg/ml (LOEC = 0.1 mg/ml), with decreased colony formation and apoptosis at \geq 1 mg/ml. In prostate cancer cells, Dibutyl Phthalate exhibited higher cytotoxicity (IC₅₀ = 27.32 ppb in DU145; 77.21 ppb in PC3) compared with Diethyl and Dimethyl Phthalates.

In recombinant human microsomes, Dibutyl Phthalate inhibited CYP2C9*1 and CYP2C19*1 activities in a concentration-dependent manner (0.1 - 100 μ M), acting as a competitive inhibitor toward probe substrates and a noncompetitive inhibitor with respect to the cofactor NADPH. In contrast, its metabolite monobutyl phthalate showed minimal inhibition.

Mice that received up to 50 mg/kg/d Dibutyl Phthalate in Tween 80 via gavage for 28 d had disrupted hepatic architecture, inflammatory cell infiltration, and hepatocellular disorganization. Serum ALT and AST activities were elevated, with the AST/ALT ratio significantly increased. Sirius red staining demonstrated collagen deposition in Dibutyl Phthalate-treated livers, and protein expression of fibrosis markers was significantly upregulated. Mice orally exposed to Dibutyl Phthalate (250 mg/kg/d), Diethylhexyl Phthalate (300 mg/kg/d), or their combination for 28 d showed elevated liver enzymes and proteins, altered lipid profiles, and histological signs of hepatic hypertrophy and structural disorganization. Dibutyl Phthalate (50 mg/kg/d) specifically increased the AST/ALT ratio, reduced albumin, and induced hepatocyte edema, sinusoidal narrowing, and central vein dilation. Increased hepatic ROS, MDA, and phosphorylated ERK1/2 levels, and the protective effects of Vitamin E and PD98059, indicate that DBP-induced hepatotoxicity involves oxidative stress and ERK1/2 activation.

In mice, 28-d oral exposure to Dibutyl Phthalate (50 mg/kg/d) caused glomerular and tubular injury, increased renal ROS and MDA, and activated ERK1/2 signaling, effects mitigated by Vitamin E or PD98059. Co-treatment with curcumin (2.5 mg/kg/d) similarly alleviated DBP-induced renal dysfunction, oxidative stress, and apoptosis. In HK-2 cells, Dibutyl Phthalate (100 μ M) upregulated Cx43 via Ang II/AMPA α 2 signaling, and Cx43 knockdown attenuated epithelial-mesenchymal transition.

In mice, 3-wk exposure to Dibutyl Phthalate (50 mg/kg/d) caused splenic inflammation and apoptosis linked to oxidative stress, which were alleviated by antioxidants. In rats, Dibutyl Phthalate (\leq 50 mg/kg/d) induced moderate splenic alterations and increased IL-1 β , ASC, and GSDMD expression, with co-exposure to BaP producing greater injury and inflammation than either chemical alone.

Dibutyl Phthalate (\geq 30 μ M) induced oxidative stress, mitochondrial dysfunction, and apoptosis in rat insulinoma cells, evidenced by increased MDA, altered GSH/GSSG ratio, elevated pro-apoptotic markers and reduced SOD activity. At 120 μ M, insulin synthesis and secretion were markedly decreased.

In male mice, Dibutyl Phthalate (50 mg/kg/d in the diet for 7 wk) alone reduced insulin secretion and induced glucose intolerance without affecting insulin resistance. When combined with a high-fat diet and low-dose STZ, Dibutyl Phthalate markedly aggravated glucose intolerance, insulin resistance, and pancreatic and renal lesions, accompanied by disrupted PI3K expression, reduced AKT phosphorylation, and decreased pancreatic GLUT2 levels.

In rats, an 8-wk exposure to Dibutyl Phthalate (500 mg/kg/d, gavage) alone or with Diethylhexyl Phthalate (750 mg/kg/d) elevated fasting glucose, increased insulin resistance, and impaired pancreatic β -cell function. Both treatments increased oxidative stress, suppressed PI3K and p-Akt phosphorylation, upregulated pro-apoptotic markers (Bax, Caspase-8, -9, -3), and reduced Bcl-2 and GLUT4 expression.

Dibutyl Phthalate (100 μ M) promoted angiogenesis in human endothelial cells by enhancing tube formation and upregulating angiogenic genes, accompanied by ERK1/2, Akt, and eNOS phosphorylation and increased NO production—

effects blocked by ER and GPER antagonists. In rats, Dibutyl Phthalate (25 mg/kg, oral, 28 d) reduced CAT, GSH, and Nrf2 while increasing CRP and NF- κ B, changes reversed by rutin. In mice Dibutyl Phthalate (≤ 10 mg/kg/d, oral, 6 wk), elevated serum estradiol and NO levels, effects inhibited by the estrogen receptor antagonist ICI182780.

Dibutyl Phthalate (≤ 0.2 μ M) inhibited neutrophil and macrophage formation and reduced macrophage phagocytic activity in zebrafish. In mice, Dibutyl Phthalate (≤ 40 mg/kg/d) acted as an in murine FITC-induced contact hypersensitivity by promoting TSLP expression in keratinocytes, whereas Diethyl Phthalate showed weaker effects but enhanced IL-4 production. In an asthmatic mouse model, Dibutyl Phthalate increased airway inflammation, serum IgE, and Th2/Th17 cytokines. In allergen-sensitized humans, 3-h Dibutyl Phthalate inhalation (~ 150 μ g/m³) enhanced early allergen responses, airway reactivity, M2 macrophage polarization, and circulating CD4⁺ T cells while reducing regulatory T cells and non-classical monocytes. Undiluted Dibutyl Phthalate and Diethyl Phthalate were not irritating to rabbit skin. When induced topically at 75% and challenged at 50%, Dibutyl Phthalate was not sensitizing in a GPMT. Diethyl Phthalate at up to 100% was not sensitizing in guinea pigs. Diethyl Phthalate was also not sensitizing in a LLNA when tested at up to 100%. A mouse ear swelling test was used to determine the adjuvant effect of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate (1:1 v/v with acetone; no further details). The mice were sensitized with FITC. A strong enhancement of ear-swelling response was observed in mice treated with Dibutyl Phthalate, while the response was not as strong with Diethyl Phthalate and Dimethyl Phthalate.

Dibutyl Phthalate (up to 200 μ mol/l) was used in a test to determine the utility of using human corneal endothelial cell line B4G12 to evaluate an in vitro model for eye irritancy testing. Decreased cell proliferation was observed, as well as cell toxicity. In rabbit eyes, Dibutyl Phthalate was not considered an ocular irritant when tested undiluted.

In a retrospective study on 251 metalworkers that had suspected metalworking fluid dermatitis, 199 were tested with a metalworking fluid series that included Dibutyl Phthalate in 5% pet. Negative results were observed in 197 patients and 2 had questionable reactions. A 65-yr-old man presenting with an acute erythematous rash following use of a cream product to treat itching in the peri-anal region had positive patch reactions to 5% Dibutyl Phthalate pet. (++), the cream product (++), 0.1% benzalkonium chloride aq. (+), balsam of Peru (++), and fragrance mix (++).

The median concentration of Dibutyl Phthalate metabolite in electronic waste processing workers was 14.1 μ g/g creatinine, compared to 7.54 μ g/g creatinine in the controls. However, no significant differences were observed between pre- and post-shift concentrations in the workers. The concentration of Dibutyl Phthalate in dust was correlated with the corresponding urinary metabolite in the workers. Workers that used respiratory protective equipment had significantly lower urinary concentration Dibutyl Phthalate. Metabolites of Dibutyl Phthalate were detected in manicurists at baseline, with the highest concentration detected in monobutyl phthalate (58.5 ng/ml overall, n = 37), followed by mono-iso-butyl phthalate (10.7 ng/ml overall). A significant increase in specific gravity-adjusted monobutyl phthalate concentrations was observed across the work shift. Manicurists in salons without local exhaust had a 54% increase in urinary monobutyl phthalate concentration across shift compared with a 7% decrease across shift for those with exhaust ventilation. Glove use was associated with a significant reduction (p = 0.04) in urinary cross-shift monobutyl phthalate concentration.

In a crossover-crossback prospective study of men with inflammatory bowel disease treated with Dibutyl Phthalate-containing mesalamine medications, semen parameters in the men with no Dibutyl Phthalate in the baseline decreased after exposure to the Dibutyl Phthalate mesalamine exposure (crossover versus baseline), especially motility parameters. This continued to decrease further even after crossback to mesalamine without Dibutyl Phthalate. Cumulative carryover effect of Dibutyl Phthalate exposure (crossback versus baseline) was a decrease of percent total sperm motility by 7.61, percent progressive sperm motility by 4.23, and motile sperm count by 26.0%. No significant change during crossover or crossback was observed in men that started with the Dibutyl Phthalate-containing mesalamine.

An MOE calculation performed by the EC SCCP for Dibutyl Phthalate at trace levels found in perfume products resulted in a value of 10,000. The MOE was based on a LOAEL of 2 mg/kg bw/d and a SED of 0.0002 mg/kg/d. When the calculation was performed at a concentration of 100 ppm (0.01%) Dibutyl Phthalate, the MOE was calculated to be 1350. CIR staff prepared an MOE for Diethyl Phthalate using concentrations from 2025 concentration of use survey. The resulting value was 28,301 when using a conservative NOAEL of 150 mg/kg/bw/d, and the value was 37,735 when using a NOAEL of 200 mg/kg bw/d.

Epidemiology studies have been performed on the associations of phthalate exposure to maternal metabolic outcomes, neurodevelopmental in children, reproductive health, pubertal development in girls, breast cancer, genotoxic effects, and endocrine effects. Mixed results have been found.

DISCUSSION FROM THE 1985 REPORT

A comparison of the chemical structures of the phthalates suggests that Dibutyl Phthalate may have the greatest toxicological significance.² Data are limited for both Dimethyl Phthalate and Diethyl Phthalate, and, in particular, there are clinical phototoxicity and photosensitivity data only for a preparation containing Dibutyl Phthalate. However, the Panel believes that the information contained in this report is adequate for a safety assessment of all 3 phthalates.

Dibutyl Phthalate but not Dimethyl Phthalate and Diethyl Phthalate caused testicular injury in laboratory animals. The combined teratogenic test data available to the Expert Panel are not adequate to conclude that Dibutyl Phthalate, Dimethyl Phthalate, or Diethyl Phthalate are proven teratogens. The concentrations used in cosmetic products and the rapid metabolism and elimination of these ingredients, as indicated by experimental studies, minimize the significance of the observations of testicular damage by Dibutyl Phthalate and the conflicting teratogenic test results. The Panel notes that the information provided in the literature on the carcinogenicity of Dibutyl Phthalate is limited and does not permit an evaluation of the assays performed and the results obtained. The results of mutagenesis studies, however, are essentially negative.

DISCUSSION FROM THE 2017 RE-REVIEW

The Panel reviewed new studies that focused on the potential for endocrine disruption/reproductive and developmental toxicity on Dibutyl, Dimethyl, and Diethyl Phthalate and butyl benzyl phthalate.⁴ One study of children aged 5 - 9, who were part of a Manhattan-Bronx cohort, revealed detectable, although varied, levels of phthalates in the urine of all 244 study participants. Higher levels of both Diethyl Phthalate and butyl benzyl phthalate were associated with airway inflammation.

Two studies addressed diabetes and phthalates. In 1 study, there were 1015 men and women 70 yr of age from Uppsala, Sweden. One sample per participant was collected from 2001 to 2004 and analyzed 5 - 8 yr later. In this study, blood levels for Dimethyl Phthalate, Diethyl Phthalate, diisobutyl phthalate, and diethylhexyl phthalate were measured and correlated with measures of insulin resistance and poor insulin secretion in nondiabetic participants.

In the second study, urinary concentrations of phthalate metabolites measured by the Centers for Disease Control and Prevention and self-reported diabetes in 2350 women aged 20 to < 80 participating in the National Health and Nutrition Examination Survey (2001 - 2008) were used. The odds ratio for diabetes in women with higher levels of n-butyl phthalate, isobutyl phthalate, benzyl phthalate, 3-carboxypropyl phthalate, and the sum of diethylhexyl phthalate metabolites was greater than the odds ratio for women with the lowest concentrations of these phthalates.

The Panel noted that all of these studies identified associations between phthalate metabolites and either diabetes or airway inflammation. Such studies did not suggest a causal link between phthalates and any adverse outcome. The possibility that phthalate metabolites may impact peroxisome proliferation pathways was suggested in the diabetes studies, but that mechanism is not established as a mode of action. The Panel agreed that there is a need for further study of the reported association between phthalate exposures and diabetes and to investigate possible causal links.

DISCUSSION

To be developed.

CONCLUSION

To be determined.

TABLES**Table 1. Definitions, reported functions, and structures of the ingredients in this safety assessment.¹**

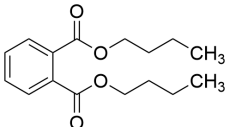
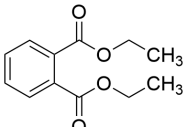
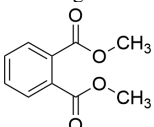
| Ingredient & CAS No. | Definition | Function(s) |
|--------------------------------|--|--|
| Dibutyl Phthalate 84-74-2 | Dibutyl Phthalate is the aromatic diester of butyl alcohol and phthalic acid. It conforms to the following structure:  | Fragrance ingredient; plasticizer; solvent |
| Diethyl Phthalate 84-66-2 | Diethyl Phthalate is the aromatic diester of ethyl alcohol and phthalic acid. It conforms to the following structure:  | Denaturant; fragrance ingredient; plasticizer; solvent |
| Dimethyl Phthalate 131-11-3 | Dimethyl Phthalate is the diester of methyl alcohol and phthalic acid. It conforms to the following structure:  | Fragrance ingredient; plasticizer; solvent |

Table 2. Chemical properties

| Property | Value | Reference |
|---|---|-----------|
| <i>Dibutyl Phthalate</i> | | |
| Physical Form | colorless oily liquid | 2 |
| Odor | odorless | 2 |
| Molecular Weight (g/mol) | 278.34 | 2 |
| Density (g/ml @ 20°C) | 1.05 | 2,5 |
| Viscosity (mm ² /s) @ 20°C) | 18.8 | 5 |
| Vapor pressure (mmHg@ 150°C) | 1.1 | 2 |
| (mmHg@ 20°C) | 7.5 x 10 ⁻⁵ | 5 |
| Melting Point (°C) | -35 | 2 |
| Boiling Point (°C) | 340 | 2,5 |
| Water Solubility (mg/l @ 25°C & pH 6) | 11.4 | 5 |
| Other Solubility | alcohol, ether, and other common organic solvents | 2 |
| log P _{ow} (@ 30 °C) | 4.46 | 5 |
| <i>Diethyl Phthalate</i> | | |
| Physical Form | colorless or pale yellow oily liquid | 2,6 |
| Odor | odorless to slight odor | 2,20 |
| Molecular Weight (g/mol) | 222.23 | 2 |
| Density (g/ml @ 20°C) | 1.12 | 6 |
| Viscosity (mm ² /s) @ 20°C) | 11.53 | 6 |
| Vapor pressure (mmHg@ 16 °C) | 14 | 2 |
| (mmHg@ 25°C) | 0.0021 | 6 |
| Melting Point (°C) | -40.5 | 2,20 |
| | -60 | 6 |
| Boiling Point (°C) | 295-298 | 2,6 |
| Water Solubility (mg/l @ 20°C & pH 7.2) | 932 | 6 |
| Other Solubility | alcohol, ether, and other common organic solvents | 2 |
| log P _{ow} | 2.47 | 20 |
| (@ 41°C) | 2.2 | 6 |
| <i>Dimethyl Phthalate</i> | | |
| Physical Form | colorless oily liquid | 2 |
| Odor | odorless to slight odor | 2 |
| Molecular Weight (g/mol) | 194.19 | 2 |
| Specific Gravity (@ 20°C) | 1.19 | 7 |
| Vapor pressure (mmHg @ 20°C) | < 0.1- < 0.01 | 2 |
| | 7.5 x 10 ⁻⁴ | 7 |
| Melting Point (°C) | 5.5 | 2 |
| | 0.36 | 7 |
| Boiling Point (°C) | 282 | 2 |
| | 283.1 | 7 |
| Water Solubility (mg/l @ 25°C) | 4000 | 7 |
| Other Solubility | alcohol, ether, and other common organic solvents | 2 |
| log P _{ow} (@ 25°C) | 1.54 | 7 |

Table 3. Frequency and concentration of use according to likely duration and exposure and by product category

| | Dibutyl Phthalate | | Diethyl Phthalate | |
|---|--------------------------|------------------------|--------------------------|------------------------|
| | # of Uses | Max Conc of Use | # of Uses | Max Conc of Use |
| | RLD (2024) ¹¹ | % (2025) ¹² | RLD (2024) ¹¹ | % (2025) ¹² |
| Totals* | 2 | NR | 168 | 0.1-0.15 |
| summarized by likely duration and exposure** | | | | |
| <i>Duration of Use</i> | | | | |
| <i>Leave-On</i> | 2 | NR | 235 | 0.15 |
| <i>Rinse-Off</i> | NR | NR | NR | 0.1 |
| <i>Diluted for (Bath) Use</i> | NR | NR | NR | NR |
| <i>Permanent Tattoo Ink</i> | NR | NR | NR | NR |
| <i>Unknown</i> | NR | NR | NR | NR |
| <i>Exposure Type</i> | | | | |
| Baby Products | NR | NR | NR | NR |
| Children's Makeup | NR | NR | NR | NR |
| Eye Area | NR | NR | NR | NR |
| Incidental Ingestion | NR | NR | NR | NR |
| Mucous Membrane | NR | NR | NR | NR |
| Incidental Inhalation-Spray | NR | NR | 231; 2 ^a | NR |
| Incidental Inhalation-Airbrush | NR | NR | NR | NR |
| Incidental Inhalation-Powder | NR | NR | 2 ^a | 0.15 ^b |
| Dermal Contact | NR | NR | 233 | 0.1-0.15 |
| Deodorant (underarm) | NR | NR | NR | NR |
| Hair - Non-Coloring | NR | NR | 2 | NR |
| Hair-Coloring | NR | NR | NR | NR |
| Nail | 2 | NR | NR | NR |
| Tattoo Preparations | NR | NR | NR | NR |
| Other Preparations (Unknown Exposure Type) | NR | NR | NR | NR |
| as reported by product category | | | | |
| <i>Fragrance Preparations</i> | | | | |
| Cologne and Toilet Water | | | 70 | NR |
| Perfumes | | | 114 | NR |
| Other Fragrance Preparation | | | 47 | NR |
| <i>Hair Preparations (non-coloring)</i> | | | | |
| Tonics, Dressings, Other Hair Grooming Aids | | | 1 | NR |
| Other Hair Preparations | | | 1 (l.o.) | NR |
| <i>Manicuring Preparations</i> | | | | |
| Nail Polish and Enamel | 2 | NR | | |
| <i>Shaving Preparations</i> | | | | |
| Aftershave Lotions | | | 2 | NR |
| <i>Skin Care Preparations</i> | | | | |
| Cleansing | | | NR | 0.1 |
| Face and Neck (excluding shaving preps) | | | NR | 0.15 (l.o.) |

NR – not reported

l.o. – leave-on

*The sum of all exposure types or for all product categories may not equal the sum of total uses because each ingredient may be used in cosmetics with multiple *exposure* types and because each formulation may be reported for multiple product categories.

**Likely duration and exposure are derived from survey data based on product category (see Use Categorization <https://www.cir-safety.org/cir-findings>)

^a Not specified whether a spray or a powder, but it is possible the use can be as a spray or a powder, therefore the information is captured in both categories

^b It is possible these products are powders, but it is not specified whether the reported uses are powders.

Table 4. Acute toxicity studies

| Test Article | Vehicle | Animals/Group | Concentration/Dose | Protocol | LD ₅₀ /LC ₅₀ /Results | Reference |
|------------------------------|--------------|--|-----------------------------------|---|---|--------------|
| ORAL | | | | | | |
| Dibutyl Phthalate | olive oil | male and female Sprague-Dawley rats, number not reported | 50%; no further details available | Acute oral study performed in accordance with OECD TG 401; rats observed for 7 d before being killed and necropsied; no further details available | 6279 mg/kg bw for both sexes; no further details available | ⁵ |
| Diethyl Phthalate | neat | groups of 5 male and 5 female Wistar rats | 0.5, 1, 2, or 5 ml/kg | Acute oral study performed in accordance with OECD TG 401; rats observed for 14 d before being killed and necropsied | > 5 ml/kg bw; 1 male rat in the 5 ml/kg was killed on day 5 after dosing due to mild depression and body weight loss | ⁶ |
| Dimethyl Phthalate; 85% pure | not reported | groups of 5 male and 5 female Sprague-Dawley rats | 2, 3.5, 5, or 8 ml/kg bw | Acute oral study performed in accordance with OECD TG 401; no further details available | Combined LD ₅₀ = 5.1 g/kg bw, male LD ₅₀ = 5.7 g/kg bw; female LD ₅₀ = 4.4 g/kg bw; signs of toxicity included unsteady gait, sluggishness, lacrimation, kyphosis, red crust around nose and eyes, wetness on periurogenital fur, drooping eyelids, piloerection, unkept appearance, and moribund appearance; death of 4 animals (groups not described for deaths or survivors) occurred at 10 - 17 min; most deaths occurred at 1 - 3 d; survivors recovered at 3 - 7 d; necropsy revealed mottled and red to pink lungs, whit to red glandular stomachs with a thickened appearance, red focal areas in a few stomachs, a few gas-filled stomachs, red intestines, tan foci in one liver, tan calculus in one kidney, and small testes in one male (groups not described for necropsy results) | ⁷ |

Table 4. Acute toxicity studies

| Test Article | Vehicle | Animals/Group | Concentration/Dose | Protocol | LD ₅₀ /LC ₅₀ /Results | Reference |
|-------------------|---------|---|--------------------------------|--|--|--------------|
| INHALATION | | | | | | |
| Dibutyl Phthalate | air | groups of 5 male and 5 female Sprague-Dawley rats | 0, 12.45, 15.68, or 16.27 mg/l | Rats exposed to test material as an aerosol for 4 h; type of inhalation exposure not specified. Respirable fraction (diameter < 4.7 µm) was 64.4, 56.9, and 59.9% at 12.45, 15.68, and 16.27 mg/l, respectively; the rats were observed for 14 d | In the 15.68 mg/l group, 2 males and 3 females died; no mortalities were observed in the other 2 dose groups. A reduction in respiratory rate was observed in the 15.68 dose group, and poor coat condition was seen in all surviving animals during the observation period due to excessive grooming behavior. Lung/body weight ratios in rats that died prematurely at 15.68 mg/l were elevated, while these ratios were lower than those of controls in males at 12.45 and 16.27 mg/l. Macroscopy of the lungs revealed red/dark foci in several animals scattered among the treatment groups. One male and one female rat in the 15.68 mg/l group had white foci in all lung lobes. Dark red areas were seen in the lungs of 2 females at 12.45 mg/l and in 1 male and 1 female rat at 16.27 mg/l. Due to the unusual death pattern, no LC ₅₀ value could be determined, but the LC ₅₀ value was estimated to be ≥ 15.68 mg/l in this study. No further details available. | ⁵ |
| Diethyl Phthalate | air | 3 male and 3 female rats; strain not reported | 4.64 mg/lg (511 ppm) | Rats exposed to test material via whole body inhalation chamber for 6 h and then observed for 14 d; no further details available | > 4.64 mg/l; no further details available | ⁶ |
| OTHER | | | | | | |
| Dibutyl Phthalate | water | <i>Danio rerio</i> ; 10 fish per tank | 0.6, 1.3, 2.5, 5.0, or 10 mg/l | Toxicity study in fish (OECD TG 236); zebrafish were exposed to test material at up to 96 h | After 96 h, the LC ₀ is 1.3 mg/l and the LC ₁₀₀ is 5 mg/l | ⁵ |

Table 5. Short-Term Toxicity Studies

| Test Article | Vehicle | Animals/Group | Study Duration | Dose/Concentration | Protocol | Results | Reference |
|-------------------|--------------|---|----------------|--|--|---|---------------|
| ORAL | | | | | | | |
| Dibutyl Phthalate | dietary feed | groups of 5 male and 5 female Wistar rats | 28 d | 0, 150, 1000, or 7000 ppm Dibutyl Phthalate, with or without 3000 ppm diethylhexyl phthalate | Rats received test material in dietary feed for 28 d. Rats were checked daily for clinical signs of toxicity and mortalities. Feed consumption was determined on days 6, 13, 20, and 27 and body weight was measured on days 0, 3, 6, 13, 20, and 27. Rats were killed at study end and liver and testes weights were measured. Liver tissue homogenates tested for cyanide non-sensitive palmitoyl-CoA oxidation. | No clinical signs of toxicity observed in any dose groups. Body weight gains and feed consumption were not affected in any dose groups, with the exception of a non-statistically significant reduction in the body weight in the 7000 ppm Dibutyl Phthalate + diethylhexyl phthalate males. An increase of absolute liver weight in males and females were observed in the 7000 ppm Dibutyl Phthalate group. Relative liver weights were increased in males and females of the 7000 ppm dose group and in males of the 1000 ppm dose group. In the combined groups, an increase of absolute and relative liver weights were observed in all animals in all combined treatment groups. No effects were observed on absolute or relative testes weights in the any of the Dibutyl Phthalate alone groups, and very weak increases of absolute and relative testes weights were observed in the combined groups (no further details). A statistically significant ($p < 0.01$) increase in cyanide-insensitive palmitoyl-CoA oxidation levels were observed only in the 7000 ppm males; statistically significant increases were observed in all combined groups. | ³³ |

Table 5. Short-Term Toxicity Studies

| Test Article | Vehicle | Animals/Group | Study Duration | Dose/Concentration | Protocol | Results | Reference |
|------------------------------|--|---|----------------|--------------------|--|--|---------------|
| Dibutyl Phthalate, >99% pure | not specified; saline used for control | groups of 6 male Balb/c mice; 6 groups total; 108 mice across 3 rounds of experiments | 60 d | 40 mg/kg bw/d | <p>From days 1–53, saline or Dibutyl Phthalate was administered to mice via gavage; melatonin (5 mg/kg, oral) was given 5 h post-Dibutyl Phthalate in the melatonin groups. On days 21, 35, 42, and 49, groups received either saline or OVA (83.3 mg with aluminum hydroxide i.p.), followed by aerosol challenge (saline or 1% OVA, 30 min/d) on days 54–60.</p> <p>Following the treatments, endpoints measured include serum total IgE and OVA-specific IgE, BAL fluid cytokines (IL-4, IL-5, IL-13, IL-17A), lung histopathology, airway hyperresponsiveness (AHR; Ri, Re, Cldyn), oxidative stress markers (GSH, MDA, 8-OHdG), and CGRP.</p> | <p>Dibutyl Phthalate exposure group showed no significant differences in body weight gain or lung function parameters (Ri, Re, Cldyn) compared to saline controls. However, oxidative stress biomarkers in lung tissue were significantly altered, including 8-OH-dG, GSH and MDA ($p < 0.05$). Serum total IgE levels showed a statistically significant increase compared to controls ($p < 0.05$), whereas cytokine levels (IL-4, IL-5, IL-13, IL-17A) and inflammatory cell counts in BAL fluid were not significantly different from controls.</p> <p>Compared to OVA alone, mice treated with OVA and Dibutyl Phthalate exhibited significantly greater inflammatory cell infiltration into the lungs, elevated serum IgE levels ($p < 0.01$), and increased concentrations of Th2 (IL-4, IL-5, IL-13) and Th17 (IL-17A) cytokines in bronchoalveolar lavage fluid ($p < 0.01$), accompanied by pronounced alterations in lung histology and airway hyperresponsiveness ($p < 0.01$), which are characteristic features of allergic asthma. Oxidative stress markers were altered, with reduced GSH and increased MDA and 8-OHdG levels in the OVA+Dibutyl Phthalate group compared to OVA alone ($p < 0.01$). In addition, CGRP concentrations in lung tissue were significantly elevated in the OVA+Dibutyl Phthalate group compared to OVA alone ($p < 0.01$). Blocking oxidative stress by co-treatment with melatonin attenuated oxidative stress and CGRP expression, and alleviated the asthma-like symptoms. Taken together, these findings indicate that Dibutyl Phthalate exacerbates asthma-like pathologies by increasing the expression of CGRP mediated by oxidative stress.</p> | ³⁴ |

Table 5. Short-Term Toxicity Studies

| Test Article | Vehicle | Animals/Group | Study Duration | Dose/Concentration | Protocol | Results | Reference |
|------------------------------|----------|---|----------------|--------------------|---|---|-----------|
| Dibutyl Phthalate, >98% pure | corn oil | male Wistar rats, 6–8 wk old; 4 groups (12/group) | 9 wk | 750 mg/kg bw/d | Rats were randomly divided into four groups: control, Dibutyl Phthalate, Dibutyl Phthalate + resveratrol (Res), and Res. The Dibutyl Phthalate and Dibutyl Phthalate + Res groups received a daily intragastric administration of Dibutyl Phthalate. The Dibutyl Phthalate + Res and Res groups were given Res at 100 mg/kg bw/d. On days 42, 49, and 56, rats in the Dibutyl Phthalate and Dibutyl Phthalate + Res groups were intraperitoneally injected with 1 ml of OVA sensitizer (1 mg OVA and 100 mg aluminum hydroxide in 1 mL saline). From days 57 to 63, rats were nebulized with 1% OVA for 30 min daily, while the control group and the Res group were atomized with 0.9% (w/v) saline. After 24 h of the last treatment, the rats were weighed and anesthetized. | <p>Compared with controls, Dibutyl Phthalate-exposed rats showed lung swelling, congestion, increased lung wet weight and lung/body weight index ($p < 0.05$), along with significant histopathological lesions including inflammatory cell infiltration, epithelial shedding, and mucus secretion.</p> <p>Dibutyl Phthalate-exposed group showed significant increase in inflammatory cell counts in BAL fluid ($p < 0.05$), accompanied by elevated Th2 cytokines (IL-4 and IL-13), as determined by ELISA.</p> <p>OVA-IgE levels and LDH activities were markedly increased in the Dibutyl Phthalate group compared with the control group ($p < 0.05$).</p> <p>In the Dibutyl Phthalate-exposed group, oxidative stress was evident, as indicated by increased MDA and decreased GSH and SOD. TUNEL staining revealed increased apoptosis in lung tissue.</p> <p>Western blot analysis showed elevated the protein levels of thymic stromal lymphopoietin TSLP, JAK1, and phosphorylated STAT6, while the nuclear translocation of Nrf2 and its downstream antioxidant proteins NAD(P)H quinone oxidoreductase 1(NQO1) and heme oxygenase-1 were suppressed.</p> <p>In general, Res co-treatment markedly attenuated the above Dibutyl Phthalate induced effects. In general, Res co-treatment markedly attenuated the above Dibutyl Phthalate induced effects.</p> | 35 |

Table 5. Short-Term Toxicity Studies

| Test Article | Vehicle | Animals/Group | Study Duration | Dose/Concentration | Protocol | Results | Reference |
|-------------------|--------------|---|-----------------|---|---|---|-----------|
| Diethyl Phthalate | dietary feed | Groups of 5 male and 5 female Sprague-Dawley rats at 2 and 6 wk standard study; groups of 15 males and females in 16 wk standard study; groups of 6 males and 6 females in pair feeding study | 2, 6, and 16 wk | 0, 0.2, 1, and 5% in standard study; 0 and 5% in pair feeding study (approximately equivalent to 0, 150, 750, and 3170 mg/kg/d) | <p>Rats were fed continuously for 2, 6, and 16 wk for standard study and 16 wk for pair feeding study.</p> <p>Standard study: body weight, feed and water consumption were measured at intervals. At the end of each designated treatment period, the rats were killed and blood samples were collected for hematological and serum examination (16 wk groups only). Necropsy was performed, organs weighed and underwent histological examination.</p> <p>Pair feeding study: groups of 6 rats of each sex, individually housed and fed untreated diet (controls) or diet containing test material. Each control animal was from the same litter as one of the treated rats of the same sex. The treated rats were fed ad libitum and the weight consumed by each one was recorded daily. Each control rat was given a weight equal to that consumed by its paired litter mate in the previous 24 h period. Body weights were recorded at intervals during the treatment period.</p> | <p>NOAEL = 150 mg/kg diet</p> <p>In the standard study throughout treatment, male and females of the 5% group and females of the 1% group gained significantly less weight than the controls, transient lower body weight gains were observed in the 1% males; overall feed consumption was significantly lower than that of controls in 5% male and female rats, and 1% females. In 6 wk groups, the erythrocyte count in 5% males was significantly higher than in controls; this change was associated with a non-significant increase in hemoglobin levels when compared with controls. Male and female rats that received 5% Diethyl Phthalate for 16 wk had statistically significantly lower absolute weights of the brain, heart, spleen & kidneys than controls; reductions in the absolute weights of the female gonads and male heart, spleen, and kidneys were also observed in the 5% groups after 2 and 6 wk of treatment. The weights of stomach of rats in the 5% group increased at times during the study but attained statistical significance with respect to controls after 2 wk (males) and 2 & 16 wk (females). The weight of the full caecum was also increased in females 5% group that received the test material for 16 wk. Increased kidney weights were seen in males & lower pituitary weights in females in the 1% group after 2 wk of treatment. Increased gonad weights were also seen in females of the 1% group after 6 wk. In the 5% group, the relative weights of the brain, kidney (16 wk only), liver, stomach, small intestine, and full caecum were increased in rats of both sexes & of testes of males. In the rats given 0.2 or 1%, there were similar increases in the relative weights of the liver, stomach & small intestine.</p> <p>In pair feeding study, rats that received 5% Diethyl Phthalate lost more weight than their respective controls over the 1st day of treatment; weight gain in the 5% group was also lower than controls thereafter. Statistical significance was attained in week 16. Total feed consumed in the 5% groups were greater than that consumed by their respective controls. No further details were available.</p> | 6 |

Table 5. Short-Term Toxicity Studies

| Test Article | Vehicle | Animals/Group | Study Duration | Dose/Concentration | Protocol | Results | Reference |
|--|----------|--------------------------------------|----------------|--|---|--|---------------|
| Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate, monobutyl phthalate, monomethyl phthalate, monomethyl phthalate, and phthalic acid | corn oil | groups of 6 male Sprague-Dawley rats | 4 wk | 250 mg/kg bw/d for monoesters and phthalic acid; 500 mg/kg bw/d for diesters | Rats received test materials orally via gavage for 4 wk; control group received corn oil only; animals observed for clinical signs of toxicity and mortality daily; body weights were measured every 3 d; feed consumption measured twice per wk; collected urine underwent analysis; organs were weighted at necropsy; blood was also collected at necropsy for analysis; and sperm was collected for sperm count analysis and motility analysis (results presented in the DART section below) | No mortalities observed. Body weights of rats in the Dibutyl Phthalate, monobutyl phthalate, and phthalic acid groups were significantly decreased ($p < 0.05$) after 2-wk administration periods, but no significant differences were observed in feed consumption between treatment groups and controls. Livers weights were significantly increased in groups treated with Dibutyl Phthalate and monobutyl phthalate compared to control group. Red blood cells and hematocrit were significantly lower and mean corpuscular hemoglobin concentration and platelet count were significantly higher in the Dibutyl Phthalate group (the monoester also had significantly higher platelet counts), and hemoglobin level was reduced only in the Dimethyl Phthalate group. Glutamate oxaloacetate transaminase levels in Dibutyl Phthalate and monobutyl phthalate groups and alkaline phosphatase levels in the Dibutyl Phthalate and Dimethyl Phthalate groups were significantly higher than the controls | ³⁶ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--|------------------------------------|--|--|---|---|--------------|
| GENERAL DEVELOPMENTAL AND REPRODUCTIVE TOXICITY STUDIES | | | | | | |
| IN VITRO | | | | | | |
| <i>monobutyl phthalate</i> | <i>not reported</i> | <i>day 10 rat embryos, no further details</i> | <i>0, 0.5, 1, 1.5, 2, 2.5, or 5 mM</i> | <i>Whole rat embryos were incubated without antibiotics in cell culture flasks containing growth medium with inactivated rat serum. Embryos were incubated with the test material in culture for 46 h. Embryos with beating hearts were considered viable and examined microscopically for growth and morphological development</i> | <i>A decrease was observed in head length at 1 mM and in all growth parameters at 1.5 mM. Morphological abnormalities at 1 mM and higher occurred in the craniofacial region, the prosencephalon and/or optic system, the auditory system, mesencephalon, and in shortening of the trunk. At 5 mM, 20% of the embryos were non-viable and those surviving had severe dysmorphogenesis.</i> | ⁸ |
| DERMAL | | | | | | |
| <i>Diethyl Phthalate</i> | <i>0.5% carboxymethylcellulose</i> | <i>groups of 12 pregnant New Zealand White rabbits</i> | <i>0, 5, 15, or 50%</i> | <i>Rabbits received test material (2 ml/kg) daily on a 10 cm² shave area of the dorso-lumbar skin on gestation days 6 through 18; test sites were not occluded and were washed with tepid water 7 h after application; rabbits were killed on gestation day 29</i> | <i>The dams showed no clinical signs of toxicity, and performance on behavioral tests in dosed animals showed no difference from vehicle controls. Maternal body weight gain and feed consumption were not affected by the treatments. There were no differences between the groups in the numbers of fetuses, the sex ratio, weights of fetuses, or incidence of dead fetuses. One dam in the 50% dose group had a higher number of early resorptions, but this did not provide a statistical significance between treatment conditions. Fetal examinations revealed 1 fetus with acranie and malformations of the ribs and lumbar and coccygeal vertebrae, and 1 fetus showed hernia umbilicalis with prolapse of abdominal organs and incurved ribs. Both malformed fetuses were in the 50 % treatment group. However, the researchers concluded that these were not treatment-related effects, based on the historical occurrence of these malformations.</i> | ⁸ |
| <i>Dimethyl Phthalate</i> | <i>paraffin oil</i> | <i>groups of 15-25 pregnant outbred Mol:Wis rats</i> | <i>0, 0.5, 1, or 2 ml/kg bw</i> | <i>Rats received test material on shaved skin for 2 h/d on gestation days 6 to 15 or 1 to 20; test sites were occluded; dams were killed on gestation day 21</i> | <i>No treatment-related signs of toxicity in the dams; no treatment-related effects observed in fetal survivability or development</i> | ⁸ |
| ORAL | | | | | | |
| <i>Dibutyl Phthalate</i> | <i>in feed</i> | <i>pregnant mice, no further details available</i> | <i>80 - 2100 mg/kg/d</i> | <i>Mice received test material in feed throughout gestation and were killed on day 18 of gestation</i> | <i>Implantation was not affected, but resorptions and fetal deaths increased with dose. Maternal weight gain was depressed at higher doses and was due to increased embryonic or fetal death. 2 of 3 live fetuses from the 2100 mg/kg Dibutyl Phthalate group had neural tube defects. Ossification was depressed, but malformation and resorption rates and fetal weights were not significantly affected by Dibutyl Phthalate up to 350 mg/kg/o</i> | ² |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|---------------------|--|--|--|--|--------------|
| <i>Dibutyl Phthalate</i> | <i>olive oil</i> | <i>groups of 10 female rats</i> | <i>120 and 600 mg/d</i> | <i>Rats received test material via gavage approximately 3 mo prior to being mated; additional groups of female rats received the same doses for 21 d following fertilization; rats were killed on day 21 of gestation</i> | <i>There were 4, 2, and 22 resorptions in the 0, 120, and 600 mg/d dose groups, respectively. Fetuses from treated and control rats did not differ significantly in number of sternum ossification foci, in development of the bones of the base of the skull or in the paws of the front and hind extremities, and in rib fusion. Dosing for Dibutyl Phthalate before gestation did not cause any significant changes in other measured parameters. Lower placental weights and fetal weights were significantly lower in the high dose group.</i> | ² |
| <i>Dibutyl Phthalate</i> | <i>olive oil</i> | <i>groups of 11 or 12 pregnant Wistar rats</i> | <i>0, 0.5, 0.63, 0.75, or 1.0 g/kg/d</i> | <i>Rats received test material via gavage on days 7 through 15 of gestation; rats were killed on day 20 of gestation</i> | <i>Maternal weight gain was reduced ($p < 0.05$) on days 15 - 20 in the 0.63 g/kg group and on day 7 to 20 in the 0.75 and 1.0 g/kg groups. Maternal feed consumption was reduced ($p < 0.05$) on days 7 to 15 in the 0.75 g/kg group and on days 7 to 20 in the 1.0 g/kg group. The number of litters and number of implantations per litter were not affected by treatment. Number resorptions, dead fetuses, and post-implantation loss per litter were increased ($p < 0.05$) in the 0.63, 0.75, and 1.0 g/kg groups. The number of litters totally resorbed was increased ($p < 0.05$) in the 0.75 and 1.0 g/kg groups. The number of live fetuses per litter was reduced ($p < 0.05$) in the 0.63, 0.75, and 1.0 g/kg groups. No live pups were born in the 1.0 g/kg group. Body weights of live fetuses were reduced ($p < 0.05$) in the 0.63 and 0.75 g/kg groups. Significant increases ($p < 0.05$) of number of fetuses with external malformations were observed in the 0.75 g/kg dose group</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 10 or 11 pregnant rats</i> | <i>0, 0.75, 1.0, or 1.5 g/kg/d</i> | <i>Study related to the one described above; the dosing schedule was divided to expose pregnant rats to test material by daily gavage either on gestation days 7 to 9, 10 to 12, or 13 to 15; dams were killed on gestation day 20</i> | <i>There were no live fetuses in the 1.5 g/kg group. The 0.75 and 1.0 g/kg groups produced similar external, skeletal, and internal malformations as those seen in the study above; however there was a higher incidence of skeletal malformations in the vertebral column and ribs. Of the 3 dosing schedules, the fetuses dosed on days 7 to 9 and on days 13 to 15 showed the higher incidence of fetotoxicity and incidence of malformations. The fetuses exposed on gestation days 10 to 12 showed less severe malformations than those exposed before or after this period.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|---------------------|--|-----------------------------------|---|--|--------------|
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 11 or 13 pregnant rats</i> | <i>0, 0.75, 1.0, or 1.25 g/kg</i> | <i>Study related to the two described above; the pregnant rats were dosed by daily gavage on gestation days 7 to 9, 10 to 12, or 13 to 15; dams were killed on gestation day 20</i> | <i>There was a significant increase in post-implantation loss at all doses in animals dosed in all 3 dosing schedules. Skeletal malformations were seen in the ribs and vertebrae at all doses in animals dosed on day 7 to 9. No increased incidence of malformation due to Dibutyl Phthalate were observed in animals dosed on days 10 to 12. Increased incidence of cleft palate and fused sternebrae were seen at all doses tested in animals exposed on days 13 to 15. As seen in the study above, fetuses were more vulnerable to Dibutyl Phthalate-induced malformations on days 7 to 9 and on days 13 to 15 than on days 10 to 12.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>in feed</i> | <i>groups of 10 pregnant rats, strain not reported</i> | <i>2%</i> | <i>Rats received test material in feed ad libitum on gestation days 0 through 20; most dams were killed on gestation day 20, some were killed on gestation days 7, 9, or 11; no further details available</i> | <i>Body weight gain and feed consumption were reduced in dams exposed to test material. Number of implantations and number of corpora lutea were not affected by test material, but post-implantation loss and number of resorptions per litter were increased, and number of live fetuses was reduced in the exposure group. There were no live male fetuses in the treated group. Due to high fetal death, malformations were not observed. Ovarian weight, uterine weight, and serum progesterone were reduced in the treated dams on gestation day 11.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>in feed</i> | <i>groups of 10 or 15 pregnant rats, strain not reported</i> | <i>0, 0.5, 1, of 2%</i> | <i>Rats received test material in feed ad libitum on gestation days 11 through 21; dams were killed on gestation day 21</i> | <i>Body weight gain and feed consumption were reduced in dams in the 1 and 2% dose groups. There was no adverse effects observed on fetal survival, number of resorptions, or sex ratio of live fetuses. Fetal body weights in both sexes were reduced in the 2% dose group. Anogenital distance was reduced in male fetuses in the 1% and 2% dose groups. Cleft palate and fused sternebrae were observed in the 2% dose group. Undescended testes were observed in the 1% and 2% dose groups.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 10 - 12 pregnant rats, strain not reported</i> | <i>0 or 1500 mg/kg</i> | <i>Dams received a single dose of 0 or 1500 mg/kg Dibutyl Phthalate once via gavage between gestation day 6 through 16; dams were killed on gestation day 20</i> | <i>A significant increase in post-implantation loss in litters of all treatment days except in animals dosed on days 7 and 11. Cleft palate and fusion of sternebrae were seen in fetuses exposed on day 15. There were malformations of the cervical and thoracic vertebrae and of the ribs in animals exposed on day 8 and 9. Dilatation of renal pelvis was observed in animals exposed on day 9.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|---|---|---|--|---|--------------|
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 10 - 13 pregnant rats, strain not reported</i> | <i>0, 1000, or 1500 mg/kg</i> | <i>Dams received daily doses of test material via gavage on gestation days 12 to 14, 15 to 17, or 18 to 20; dams were killed on gestation day 21</i> | <i>Body weight gain and feed consumption were reduced in dams exposed to 1000 or 1500 mg/kg in all dosing schedules. Litters exposed to 1500 mg/kg on days 12 to 14 or on days 15 to 17 had increases in the number of litters totally resorbed, in the number of resorptions per litter, in post-implantation death, and in the number of males with undescended testes. Fetal body weights were reduced in both male and female fetuses that had been exposed to 1500 mg/kg at all three dosing schedules. The incidence of undescended testes was increased in fetuses exposed on gestation days 12 to 14 and 15 to 17 but not 18 to 20. The anogenital distance was reduced in males exposed in all treatment schedules.</i> | ⁸ |
| <i>Dibutyl Phthalate or monobutyl phthalate</i> | <i>olive oil for Dibutyl Phthalate; ammonium chloride for monobutyl phthalate</i> | <i>groups of 11 - 15 pregnant Sprague-Dawley rats</i> | <i>0, 1.8, 3.6, 5.4, or 7.2 mmol/kg</i> | <i>Dams received a single dose of the test materials via gavage on gestation day 10; dams were killed on day 12. Further study to determine the rate of non-viability of fetuses was performed with monobutyl phthalate treated the dams with 0, 5.4 or 7.2 mmol/kg on day 10 and killed the rats on day 12, 13, 14, or 18 (6 to 12 rats/exposure)</i> | <i>11 to 15 litters were examined/ dose group. Embryos exposed to Dibutyl Phthalate showed dose-dependent growth retardation and dysmorphogenesis. At the 1.8 mmol/kg dose, few adverse effects were seen. At the 3.6 mmol/kg and higher doses all measures of growth and development were reduced from vehicle and non-treated controls, and the incidence of malformed embryos was increased 4-fold. The observations in the higher groups for Dibutyl Phthalate included defects of the prosencephalon, the optic system, and the mandibular process. The most severely affected embryos also exhibited underdevelopment of the auditory system, maxillary process, and caudal extremity. The total number of viable embryos per litter was not affected by the Dibutyl Phthalate treatment. The embryos exposed to monobutyl phthalate exhibited the same types of growth retardation and malformations, but at higher doses (5.4 and 7.2 mmol/kg). These effects were not observed at the 1.8 or 3.6 mmol/kg monobutyl phthalate doses. In the viability study with monobutyl phthalate, an increase in the rate of non-viable fetuses was observed on days 13 and 14. On day 18, surviving fetuses from the exposed groups had eye anomalies, absent digits, misshapen lower jaws, enlargement of the pericardium and/or cardiac tube, and brain anomalies.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|-------------------|---|--|--|---|--------------|
| <i>Dibutyl Phthalate</i> | <i>in feed</i> | <i>number of pregnant F344/N rats not reported; groups of 10 male and 10 female pups per dose</i> | <i>pregnant rats = 0 or 10,000 ppm 13-wk pups = 0, 2500, 5000, 10,000, 20,000, or 40,000 ppm</i> | <i>Dams were exposed to 0 or 10,000 ppm Dibutyl Phthalate in feed from the beginning of gestation through lactation; when pups were able to eat solid food, they were exposed to the same feed conditions as their mothers until they were 8 wk old; at postnatal week 13, pups exposed to 10,000 Dibutyl Phthalate were divided into dosing groups to receive up to 40,000 ppm Dibutyl Phthalate in feed for 13 wk; at the end of the 13 wk, animals were killed for examination</i> | <i>The number of live pups born was not different between the control and 10,000 ppm perinatal groups. However, by the first day after parturition the survivability of the exposed pups was significantly lower than that of control pups and remained so through weaning. Pup body weights were also lower in the 10,000 ppm perinatal group from the time of birth through weaning. All rats survived the 13-wk dosing portion of the study. The mean body weights of control animals who had not been exposed to Dibutyl Phthalate perinatally were higher than all animals that had been dosed perinatally. Female rats that had been exposed perinatally but not during the 13-wk dosing period gained weight more rapidly than other groups that had been exposed perinatally, and by the end of the study they had body weights similar to those controls that were not dosed perinatally. The mean body weights of all groups of males exposed perinatally and of females that received 20,000 or 40,000 ppm as adults remained less than those unexposed controls at the end of the study. Males in the 40,000 ppm group lost weight during the study. All male and female animals in the 40,000 ppm group were emaciated at the end of the study and had lower food consumption rates than controls. Males in this group had abnormal posture and ruffled fur and appeared hypoactive in weeks 2 through 4 of dosing. Males that had been exposed perinatally but not during the 13-wk dosing period had lower testicular weights than unexposed control males. No other organ weights differed between these two groups. There were signs of toxicity to the liver and kidneys in both sexes and in the reproductive tissues of the males. There were no dose-related effects on the estrous cycles of the females.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>sesame oil</i> | <i>groups of 6 pregnant or groups of 6-8 pseudo-pregnant female Holtzman rats</i> | <i>pregnant rats = 0, 250, 500, 1000, or 2000 mg/kg/d pseudo-pregnant rats = 0, 500, 1000, or 2000 mg/kg/d</i> | <i>Study on the maternal effects of test material on fetotoxicity; some females were paired with males for regular breeding., while other females were subjected to artificial stimulation of the cervix by a glass rod during proestrus or estrus in order to induce a pseudopregnancy. Deciduomata were induced surgically to simulate implantation of a blastocyst in pseudo-pregnant rats. All rats were dosed by oral gavage on day 1 through day 8 of gestation or pseudo-gestation. On day 9 rats were killed, and blood was collected. Uteri and ovaries were dissected and cleaned of adjacent tissues and weighed.</i> | <i>Dibutyl Phthalate at doses up to 2000 mg/kg/d did not significantly affect whole body weights, uterine weights, ovarian weights, or serum progesterone levels in the pregnant or pseudo-pregnant rats. In pregnant rats, there was no effect of Dibutyl Phthalate on the number of implantations. The authors concluded that any fetotoxic effects of short-term Dibutyl Phthalate exposure are due to the test material's interaction with the fetus and not a result of maternal effects in early pregnancy, and that the viability of preimplantation embryos is not affected by Dibutyl Phthalate</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|--------------------------------|---|--|---|---|--------------|
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 10-13 pregnant or pseudo-pregnant female rats</i> | <i>0, 250, 500, 740, 1000, 1250, or 1500 mg/kg</i> | <i>Rats received test material via oral gavage on gestation day 0 through day 8; pregnant rats were killed on day 20 and pseudo-pregnant rats were killed on day 9</i> | <i>Body weight gain and feed consumption were reduced in dams exposed to doses of 500 mg/kg and higher. Increases in preimplantation loss were seen in doses of 1250 and 1500 mg/kg. The numbers of resorptions and dead fetuses per litter were increased in 1000 mg/kg and higher doses. Preimplantation loss was increased, and number of live fetuses per litter was reduced at doses of 750 mg/kg and higher. The fetal sex ratio was not affected by Dibutyl Phthalate, but both male and female fetal weights were reduced at all doses tested. Body weight gain and feed consumption were reduced in pseudo-pregnant rats exposed to doses of 750 mg/kg and higher. In the pseudo-pregnant rats ovarian weight, number of corpora lutea, and serum estradiol were not affected by Dibutyl Phthalate, but uterine weight was reduced at doses of 750 mg/kg and higher. Serum progesterone was reduced in pseudo-pregnant rats only at the 1500 mg/kg dose.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>in feed</i> | <i>2 strains of mice, no further details available</i> | <i>10 and 100 mg/kg/d</i> | <i>3 generation reproduction study; no further details available</i> | <i>An increase in the formation of renal cysts was observed in the F1 and F2 generations; no further details available</i> | ² |
| <i>Dibutyl Phthalate</i> | <i>oil, kind not specified</i> | <i>female rats, no further details available</i> | <i>50% solution; daily dose of 1 ml/kg</i> | <i>3 generation reproduction study; female rats dosed daily for 6 wk prior to mating with untreated males; the offspring were bred to produce 2 additional generations; no information available on whether the second and third generations were dosed with Dibutyl Phthalate</i> | <i>No impairment of reproductive performance was noted. Development, growth, and fertility were normal for all 3 generations. No further details available</i> | ² |
| <i>Dibutyl Phthalate</i> | <i>in feed</i> | <i>groups of 20 male and 20 female COBS Crl:CD-1, (IC)BR mice</i> | <i>0, 0.03, 0.3, or 1%</i> | <i>Continuous breeding protocol; mice exposed to test material for a 7-d remating period and then assigned to mating pairs within the dose groups; cohabitation period was 98 d; newborn pups were collected within 12 h of birth and discarded after physical data (body weight, proportion of males, number of litters per pair, number of live pups) were recorded. At the end of the cohabitation period, mating pairs were separated for 21 d while Dibutyl Phthalate in the diet continued. During this 21-d period, any remaining litters were delivered and kept for at least 21 d. At the end of the continuous breeding phase, mice in the high dose group were re-paired for crossover mating, to create 3 groups of pairs: control male x control female; control male x 1% female; or 1% male x control female. The offspring of the crossover pairs were analyzed, and the parents were necropsied.</i> | <i>The fertility of the breeding pairs, number of litters per pair, number of live pups per litter, and proportion of pups born alive were significantly ($p < 0.01$) reduced in the 1% dose group, compared to control. The lower doses did not affect these parameters. Live pup weight was not different from control in any treatment groups. Pups that were born to control female x 1% male pairs did not differ from pups born to two control parents. However, significantly reduced fertility ($p < 0.01$), number of live pups per litter ($p < 0.01$), proportion of live pups per litter ($p < 0.05$), and live pup weight ($p < 0.05$) were observed when 1% females were mated with control males. In male mice of the 1% dose group, total body weight was reduced ($p < 0.01$), but organ weights were not different from control. Liver weights were increased and uterus weights were reduced in 1% treated female mice, compared to controls ($p < 0.05$). Additionally, there were no differences in sperm motility, sperm concentration, or percentage of abnormal sperm in male mice exposed to 1% dose group, compared to controls</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|----------------|--|--------------------|---|--|-----------|
| <i>Dibutyl Phthalate</i> | <i>in feed</i> | groups of 20 male and 20 female VAF Crl:CD BR outbred Sprague-Dawley albino rats | 0, 0.1, 0.5, or 1% | Similar study to the continuous breeding protocol described above. Crossover pairs were as follows: control male x control female; 1 % male x control female; and control male x 1 % female (n = 20 pairs/crossover group). | <p>The number of litters per pair was not different from control in any of the treatment groups. The number of live pups per litter and pup weights was significantly reduced ($p < 0.05$) in all dose treatments, compared to control. Live pup weights were significantly reduced ($p < 0.05$) in the 0.5 and 1.0 % groups, compared to control. The pregnancy rate, fertility, and live birth rate of the three crossover groups were not significantly different. Live pup weights were significantly reduced ($p < 0.05$) in the control male x 1% female group, compared to those in the control male x control female group.</p> <p>Body weight, liver weight, and kidney weight were reduced ($p < 0.05$) in 1% female F_0 rats, compared to controls. In 1% male F_0 rats, body weight was not different, but kidney and liver weights were reduced ($p < 0.05$), compared to controls. Sperm endpoints (count, morphology, and motility) were not affected by treatment in F_0 males at the doses tested.</p> <p>The mating index (number of females with copulatory plugs per number of cohabited pairs), pregnancy index (number of fertile pairs per number of cohabited pairs), and fertility index (number of fertile pairs per number of females with plugs) were all significantly reduced ($p < 0.05$) in the 1 % F_1 mating pairs. Live pup weights were reduced ($p < 0.05$) in all concentrations tested compared to controls. Necropsy of F_1 rats revealed reduced ($p < 0.05$) body weights for both males and females in the 1 % treatment. In the 1% males, the prostate, seminal vesicles, and right testes had reduced ($p < 0.05$) weights, compared to controls, and the liver and kidneys weights were increased ($p < 0.05$). While epididymal sperm motility, number, and morphology were not affected by treatments, the total spermatid heads per testis and per g testis were reduced ($p < 0.05$) in the 1% dose group, compared to control.</p> | 8 |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|----------------------------|--------------------------|---|---|--|--|--------------|
| <i>Diethyl Phthalate</i> | <i>in feed</i> | <i>groups of male and female Swiss CD-1 mice</i> | <i>0, 0.25, 1.25, or 2.5%</i> | <i>Continuous breeding protocol; F₁ mice from the control and high dose group were mated within their treatment groups (not cross-mated between groups); no further details available</i> | <i>Feed consumption was not affected by test material. There were no effects by Diethyl Phthalate on fertility or reproductive outcome of the F₀ generation. There were no treatment-related differences between the treated and control groups of the F₂ generation. Male and Female F₁ mice of the high dose group had a reduced adult body weight and increased liver weight. Male F₁ mice of the high dose group showed an increased prostate size and a reduced sperm count. However, motility and morphology of the sperm were not affected by Diethyl Phthalate and neither were testicular, epididymal, and seminal vesicle weights.</i> | ⁸ |
| <i>Diethyl Phthalate</i> | <i>in feed</i> | <i>groups of 26 - 32 pregnant Sprague-Dawley rats</i> | <i>0, 0.25, 2.5, or 5% (0, 0.2, 1.91, or 3.21 g/kg/d)</i> | <i>Dams received diet containing test material on gestation days 6 to 15; dams were killed on gestation day 20</i> | <i>5% dose group had reduced body weight gain during the dosing period; however, feed and water consumption were reduced in this group during dosing, so the effect on body weight gain may have been due to palatability of the test material in the feed. There was no effect of the test material on any parameter of fetal viability or development, except for an increased incidence of supernumerary ribs in the 5% group.</i> | ⁸ |
| <i>Dimethyl Phthalate</i> | <i>corn oil</i> | <i>groups of 36 - 40 pregnant CD-1 mice</i> | <i>0 or 3500 mg/kg</i> | <i>Mice received test material daily via gavage on gestation days 7 through 15; offspring were examined externally and weighed 12 to 24 h after birth and again on postnatal day 3</i> | <i>There were no signs of maternal toxicity or changes in weight gain due to test material. There were no treatment-related effects on fetal survivability, litter size, litter weight, or physical development of the fetus.</i> | ⁸ |
| <i>Dimethyl Phthalate</i> | <i>in feed</i> | <i>groups of 25-30 pregnant Sprague-Dawley rats</i> | <i>0, 0.25, 1, or 5% (0, 0.2, 0.84, or 3.57 g/kg/d)</i> | <i>Rats received test material in feed on gestation days 6 to 15; dams were killed on gestation day 20</i> | <i>The high dose group had increased maternal liver weights and reduced body weight gain. However, feed and water consumption were reduced in the 5% group during dosing, so the effect on body weight gain may have been due to palatability of the test material in the feed. There was no effect of Dimethyl Phthalate on any parameter of fetal viability or development.</i> | ⁸ |
| <i>monobutyl phthalate</i> | <i>ammonium chloride</i> | <i>groups of 11-15 pregnant Wistar rats</i> | <i>0, 250, 500, or 650 mg/kg</i> | <i>Rats received test material via gavage on gestation days 7 to 15. Maternal body weights and feed consumption recorded daily. Dams were killed on gestation day 20</i> | <i>Maternal body weight gain and feed consumption rates were reduced in the 500 and 650 mg/kg doses ($P < 0.01$). Fetal effects observed at 500 and 650 mg/kg included increased post-implantation loss (resorptions and dead fetuses), reduced number of live fetuses, and lower body weights ($p < 0.01$). An increased incidence of external malformations (cleft palate), skeletal malformations of the cervical and thoracic vertebral arches and sternbrae, and internal malformations (dilation of the renal pelvis) was observed in the 500 and 650 mg/kg fetuses ($p < 0.05$). No developmental effects were observed at 250 mg/kg.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|--------------------------|---|--|---|--|--------------|
| <i>monobutyl phthalate</i> | <i>ammonium chloride</i> | <i>groups of pregnant Wistar rats, number per group not specified</i> | <i>0, 500, 625, or 750 mg/kg</i> | <i>Rats received test material via gavage on gestation days 7 to 9, 10 to 12, or 13 to 15. Maternal body weights and feed consumption recorded daily. All dams were killed on gestation day 20</i> | <i>Maternal body weight gain and feed consumption were reduced during the dosing periods. In the 7750 mg/kg group, dams dosed on gestation days 10 - 12 and 13 - 15 had continued reductions in feed consumption and weight gain through gestation day 20. The percentage of post-implantation loss was increased in the 625 and 750 mg/kg dose groups at all dosing schedules ($p < 0.01$) and was as high as 95% at 750 mg/kg in rats exposed on gestation days 13 - 15. Body weights of live fetuses were most reduced by monobutyl phthalate at all doses in those exposed on gestation days 7 - 9 ($p < 0.05$). Fetuses of the 625 and 750 mg/kg groups exposed during gestation days 13 - 15 had cleft palate ($p < 0.01$). Skeletal malformations of the cervical vertebral arches were observed at all doses exposed during gestation days 7 - 9, and fusion of the sternbrae was observed at the two higher doses exposed on gestation days 13 - 15 ($p < 0.05$).</i> | ⁸ |
| <i>monobutyl phthalate</i> | <i>not reported</i> | <i>groups of 16 pregnant Wistar rats</i> | <i>0, 250, 500, or 750 mg/kg</i> | <i>Rats received test material on gestation days 15 through 17 and were killed on gestation day 21. Pregnancy outcome and fetal development were recorded; no further details available</i> | <i>Maternal feed consumption was reduced in the 750 mg/kg dose group, and maternal body weight gain was reduced in the 500 and 750 mg/kg dose groups. LOELs are listed with the following fetal parameters: increased number of resorptions and dead fetuses per litter (500 mg/kg), percent post-implantation loss per litter (500 mg/kg), reduced body weights for male and female fetuses (750 mg/kg each), increased number of male fetuses with undescended testes (250 mg/kg), reduced anogenital distance of male fetuses (250 mg/kg), and reduced ratio of anogenital distance/body weight (250 mg/kg).</i> | ⁸ |
| <i>monobutyl phthalate</i> | <i>not reported</i> | <i>groups of 16 pregnant or pseudopregnant Wistar rats</i> | <i>0, 250, 500, 750, or 1000 mg/kg</i> | <i>Rats received test material on gestation days 0 through 8. Pregnant rats were killed on gestation day 20 and pseudopregnant rats were killed on day 9. Pregnancy outcome of the pregnant rats and uterine weights of the pseudopregnant rats were recorded; no further details available</i> | <i>The incidence of preimplantation loss was increased at 1000 mg/kg. Uterine decidualization in the pseudopregnant rats was decreased in the 1000 mg/kg dose group.</i> | ⁸ |
| PARENTERAL | | | | | | |
| <i>Dibutyl Phthalate and Dimethyl Phthalate</i> | <i>saline</i> | <i>groups of 5 pregnant female rats, strain not specified</i> | <i>2 or 4 ml/kg for Dibutyl; 0.5, 1, or 2 ml/kg for Dimethyl</i> | <i>Rats received test materials intraperitoneally on days 3, 6, and 9 of gestation; controls received just 4 ml/kg saline; no further details available</i> | <i>1 rat died in the 4 ml/kg Dibutyl Phthalate dose group, implantations occurred in 4 and 3 rats, respectively, in the 2 and 4 ml/kg Dibutyl groups. Dibutyl Phthalate resulted in a 50% reduction in the number of pups weaned/litter. 2 male pups, 1 from each of 2 litters in the 2 ml/kg Dibutyl group had no eyes. In the 0.5, 1, and 2 ml/kg Dimethyl Phthalate groups, 2 rats died in the 1 ml/kg dose group while all survived in the 0.5 and 2 ml/kg dose groups, and only 1 implantation occurred in the 1 ml/kg dose group. The numbers of pups weaned were not significantly different from the controls</i> | ² |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|---------------------|--|---|---|---|--------------|
| <i>Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate</i> | <i>not reported</i> | <i>Groups of 5 pregnant female rats, strain not specified</i> | <i>3.05 ml/kg for Dibutyl; 5.06 ml/kg for Diethyl; 3.4 ml/kg for Dimethyl</i> | <i>Rats received test materials intraperitoneally on days 5, 10, and 15 of gestation; control rats were untreated or received distilled water, normal saline, or cottonseed oil; maternal rats were killed on day 20; no further details available</i> | <i>Phthalate administration did not interfere with fertility, as reflected by corpora lutea:implantation site ratios. Significant effects on embryonic and/or fetal development were observed; average weights of the fetuses from the treated groups and those administered saline were significantly lower than the average weight of the fetuses from the untreated controls. A significantly higher number of skeletal abnormalities in the fetuses from the test group as compared to the controls was observed. The failure to include historical control data, as well as a positive control in the test program, made it difficult to evaluate the significance of the results.</i> | ² |
| UTEROTROPIC AND VAGINAL EFFECTS | | | | | | |
| ORAL | | | | | | |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 10 immature ovariectomized Sprague-Dawley rats</i> | <i>0, 20, 200, or 2000 mg/kg/d</i> | <i>Rats received test material via gavage for 4 d; on the day after the fourth dose, all animals were killed and the uteri were weighed; no further details available</i> | <i>The uterine weights of the treated animals did not differ from those of control animals; no further details available</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>mature ovariectomized female Sprague-Dawley rats; number not reported</i> | <i>0, 20, 200, or 2000 mg/kg/d</i> | <i>Rats received test material via gavage for 4 d; each day of the study, physiological saline was used for vaginal lavage of the rats to determine cornification of vaginal cells., the percentage of cornified cells was used as an indicator of estrogenic activity</i> | <i>The percentages of cornification of vaginal cells from the treated rats were similar to those of control rats.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>ovariectomized rats; no further details available</i> | <i>1000 mg/kg/d</i> | <i>Rats were exposed to estradiol benzoate (10 µg) or test material for 2 d; on the third day, the rats were given 0.5 mg progesterone; 6-8 h after the progesterone injection, the females were paired with a sexually active male for observation of lordosis behavior. After the lordosis behavioral observations, the females were killed and their uteri were cleaned of mesenteric fat and weighed</i> | <i>Dibutyl Phthalate treatments did not affect uterine weights or lordosis behavior, compared to control (oil) treatment. The estradiol group showed significant increases in these parameters.</i> | ⁸ |
| <i>Diethyl Phthalate</i> | <i>peanut oil</i> | <i>groups of 10 immature female Wistar rats</i> | <i>0, 50, 150, 500 mg/kg/d</i> | <i>Rats received a daily dose of test material of 0.4 mg/kg β-estradiol (positive control) for 3 d; 24 h after last dose, rats were killed and uteri were removed and weighed</i> | <i>The uteri of rats in the β-estradiol positive control group were significantly heavier than those of the vehicle control group, as expected. The weights of the uteri of rats exposed to Diethyl Phthalate were similar to those in the control group.</i> | ⁸ |
| PARENTERAL | | | | | | |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>ovariectomized rats; no further details available</i> | <i>200 or 400 mg/kg/d</i> | <i>Rats were exposed to estradiol benzoate (10 µg) or test material subcutaneously for 2 d; on the third day, the rats were given 0.5 mg progesterone; 6-8 h after the progesterone injection, the females were paired with a sexually active male for observation of lordosis behavior. After the lordosis behavioral observations, the females were killed and their uteri were cleaned of mesenteric fat and weighed</i> | <i>Dibutyl Phthalate treatments did not affect uterine weights or lordosis behavior, compared to control (oil) treatment. The estradiol group showed significant increases in these parameters.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|--|--|--|--|--|-----------|
| TESTICULAR EFFECTS | | | | | | |
| IN VITRO | | | | | | |
| monoethyl phthalate | DMSO | Wistar rat Leydig cells | 0 or 1000 μ M | Cell cultures were exposed test material or vehicle for 3 h. Some cells were exposed to luteinizing hormone to test for Leydig cell function integrity. Cells were fixed and processed for histopathological evaluation | Monoethyl phthalate did not affect testosterone output or induce any structural changes in Leydig cell morphology or ultrastructure. | 8 |
| ORAL | | | | | | |
| Dibutyl Phthalate or sodium salt of monobutyl phthalate | corn oil for Dibutyl Phthalate; aqueous solution of sodium salt of monobutyl phthalate | Groups of 6 male Sprague-Dawley rats, groups of 10 male TO mice, groups of 5 male Dunkin-Hartley guinea pigs, and groups of 8 male DSN Syrian hamsters | 0, 2000, or 3000 mg/kg for Dibutyl Phthalate; at least up to 1600 mg/kg/d for monobutyl phthalate solution | Animals received test material for 7 to 9 d; histological examinations of the testes were performed; in additional experiments, male rats were exposed to 3000 mg/kg/d Dibutyl Phthalate for 9 d or to 800 mg/kg/d monobutyl phthalate for 5 d | Dibutyl Phthalate produced severe atrophy in the seminiferous tubules of rats and guinea pigs, with milder effects in mice. Hamster testes were not affected by Dibutyl Phthalate. Results for monobutyl phthalate not reported. In additional experiments, each treatment produced a significant reduction ($p < 0.001$) in testicular weight, compared to control treatments. Histologically, the treated rats showed $> 90\%$ tubular atrophy in the testes. However, neither 3000 mg/kg Dibutyl Phthalate nor 1600 mg/kg/d monobutyl phthalate produced any changes in testicular weight in hamsters. No further details available | 8 |
| Dibutyl Phthalate | not reported | male mice, rats, hamsters, guinea pigs, and ferrets; number and strains not reported | 2 g/kg/d | Animals received test material orally for 10 d; no further details available | Testicular atrophy was observed in mice, rats, guinea pigs, and ferrets, but not in hamsters | 2 |
| Dibutyl Phthalate | not reported | 28 male Wistar rats | 8.6 mmol/kg/d | Rats received single dose of test material; no further details available | A decrease in testicular fructose and glucose and caused sloughing of germ cells in the first days of dosing. Testicular iron and zinc levels decreased, and inositol and cholesterol levels increased. After repeated doses, sloughing led to atrophy and dissociation of germ cells from Sertoli cells, and testicular levels of triglycerides, cholesterol, phosphatidyl choline, and phosphatidyl ethanolamine were reduced. Additionally, increases were observed in γ -glutamyl transferase, lactate dehydrogenase, lactate, alkaline phosphatase, and flavin adenine dinucleotide in the testes. There were also decreases in sorbitol, sorbitol dehydrogenase, succinate dehydrogenase, and aldose reductase in the testes. Transferrin was increased in Sertoli cells, Sertoli-germ connection, epididymis and liver and decreased in the seminal vesicle. Ferritin was reduced in the seminiferous lumen. | 8 |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|---------------------|--|-------------------------------------|---|---|--------------|
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>28 male Wistar rats</i> | <i>8.6 mmol/kg/d</i> | <i>Rats received oral doses of test material for 7 d; no further details available</i> | <i>A decrease in testicular fructose and glucose and caused sloughing of germ cells in the first days of dosing. Testicular iron and zinc levels decreased, and inositol and cholesterol levels increased. After repeated doses, sloughing led to atrophy and dissociation of germ cells from Sertoli cells, and testicular levels of triglycerides, cholesterol, phosphatidyl choline, and phosphatidyl ethanolamine were reduced.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 6 male Wistar rats</i> | <i>0, 250, 500, or 1000 mg/kg/d</i> | <i>Rats received test material via gavage for 15 d; animals were weighed daily; on day 16 rats were killed and testes were promptly weighed and prepared for histopathological examination and enzyme activity assessments</i> | <i>No unscheduled deaths observed, but body weights and testicular weights were reduced in the 500 and 1000 mg/kg doses. Testicular sorbitol dehydrogenase activity was reduced in 500 and 1000 mg/kg dose groups. The activity of testicular acid phosphatase was reduced at all doses. Glucose-6-phosphate dehydrogenase, β-glucuronidase, and γ-glutamyl transpeptidase in the testes showed increased activities at all doses tested. Histopathological examination of the testes revealed dose-dependent damage to the seminiferous tubules. Spermatocytes in the 1000 mg/kg dose group showed pycnotic nuclei.</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>corn oil</i> | <i>pregnant rats; no further details available</i> | <i>0 or 500 mg/kg/d</i> | <i>Rats received test material via gavage on gestation days 12 to 21; no further details available</i> | <i>On gestation days 16-21, Dibutyl Phthalate caused focal hyperplasia of Leydig cells in male fetuses. At gestation day 21, Dibutyl Phthalate exposed fetuses showed testicular atrophy and seminiferous cords were enlarged. Dibutyl Phthalate-exposed male fetuses had reduced testosterone levels, fewer epididymal ducts, and reduced "AR" staining. The Leydig cell hyperplasia was proposed to be a compensatory response to increase steroidogenesis triggered by reduced testosterone. However, the overall androgen concentration was not corrected to the levels necessary to prevent malformations of the reproductive tract.</i> | ⁸ |
| <i>Diethyl Phthalate</i> | <i>corn oil</i> | <i>3 male Wistar rats</i> | <i>0 or 2 g/kg/d</i> | <i>Rats received test material via gavage for 2 d; all rats killed 24 h after the second dose; bodies were perfused with a fixative solution, and the testes were removed, sliced, and stained for histological examination</i> | <i>Diethyl Phthalate produced no change in seminiferous tubular structure or Leydig cell morphology. However, the Leydig cells of treated rats showed ultrastructural changes, specifically swelling of the mitochondria and vesiculation and focal dilatation of the smooth endoplasmic reticulum.</i> | ⁸ |
| <i>Diethyl Phthalate</i> | <i>corn oil</i> | <i>5 pregnant Sprague-Dawley rats</i> | <i>0 or 750 mg/kg/d</i> | <i>Dams received test material on gestation day 14 through postnatal day 3; developing male pup were evaluated for affects; no further details available</i> | <i>Maternal weight gain, litter size, pup weight, and sex ratio of pup were not affected by Diethyl Phthalate. No effects from Diethyl Phthalate were observed with weights of the testes and other reproductive tissues, anogenital distance, and incidence of males with areolas.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|---------------------|--|---|---|--|--------------|
| <i>Dimethyl Phthalate</i> | <i>corn oil</i> | <i>5 pregnant Sprague-Dawley rats</i> | <i>0 or 750 mg/kg/d</i> | <i>Dams received test material on gestation day 14 through postnatal day 3; developing male pup were evaluated for affects; no further details available</i> | <i>Maternal weight gain, litter size, pup weight, and sex ratio of pup were not affected by Dimethyl Phthalate. No effects from Diethyl Phthalate were observed with weights of the testes and other reproductive tissues, anogenital distance, and incidence of males with areolas.</i> | ⁸ |
| <i>Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate</i> | <i>corn oil</i> | <i>groups of 12 male rats, strain not reported</i> | <i>2 g/kg/d for Dibutyl, 1.6 g/kg/d for Diethyl, or 1.4 g/kg/d for Dimethyl</i> | <i>Test material was administered via gavage for 4 d; no further details available</i> | <i>No significant changes in feed intake or body weight observed in any dose group. Diethyl Phthalate and Dimethyl Phthalate did not result in significant changes in testes weight, no testicular atrophy was observed, and urinary zinc excretion was unaffected. Decreased testes weight and severe atrophy of the seminiferous tubules were observed following administration with Dibutyl Phthalate. Most of the tubules had complete loss of spermatocytes and spermatids. An increase in the urinary excretion of zinc was also observed, and there was a decrease in the zinc content of testes on an absolute and relative weight basis. The administration of zinc, concurrently with Dibutyl Phthalate, provided substantial protection against Dibutyl Phthalate-produced testicular damage.</i> | ² |
| <i>monobutyl phthalate</i> | <i>sesame oil</i> | <i>groups of 7 pregnant Wistar-King A rats</i> | <i>0 or 3 ml dose of 0.3 g/d</i> | <i>Dams received test material on gestation days 15 to 18. On gestation day 20, 3 treated rats and 3 control 3 were killed and the male fetuses were examined for testicular effects. The remaining dams were allowed to deliver and the male offspring were killed on postnatal day 30 or 40 for analysis of the testes position and frequency of cryptorchidism</i> | <i>Males that were examined from gestation day 20 had relatively high positioned testes in the abdominal cavity, with some near the kidneys. In the controls of the males killed on postnatal day 30 or 40, the testes were allocated in the scrotum and there was a 0% incidence of cryptorchidism. In the treated males of the postnatal day 30 o4 40 group, 22 out of 26 had cryptorchidism (14 unilateral and 8 bilateral undescended testes). The majority of the undescended testes were in the abdominal cavity, while the remaining were at the external inguinal ring.</i> | ⁸ |
| ANOGENITAL EFFECTS | | | | | | |
| ORAL | | | | | | |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 10 pregnant CD rats</i> | <i>0, 250, 500, or 750 mg/kg/d</i> | <i>Dams were exposed to test material throughout pregnancy and lactation until postnatal day 20; no further details available</i> | <i>The number of live pups per litter was decreased at the 750 mg/kg/d dose, probably due to maternal effects. Anogenital distance was reduced in male (but not female) pups exposed to 500 and 750 mg/kg/d. Other effects observed in male Dibutyl Phthalate-exposed offspring include absent or underdeveloped epididymis, hypospadias, ectopic or absent testes, germ cell loss, testicular atrophy, small testes and seminal vesicles, and absent prostate and seminal vesicles. These are all androgen-mediated effects. Estrogen-mediated development in female offspring such as vaginal opening and estrous cycle were not affected by Dibutyl Phthalate exposure.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|---------------------|---|--------------------------------------|--|--|--------------|
| <i>Dibutyl Phthalate</i> | <i>corn oil</i> | <i>groups of 10 pregnant CD rats</i> | <i>0, 250, or 500 mg/kg/d</i> | <i>Dams were exposed to test material or flutamide (100 mg/kg/d in 5 rats) on gestation days 12 – 21; male pups were killed and examined either on postnatal day 60 or after postnatal day 100 to 105; female pups were killed and examined on postnatal day 20 to 30; pup examinations were focused on reproductive tissues; an immunohistological evaluation for androgen receptor expression was performed</i> | <i>Dibutyl Phthalate exposure induced the following morphological effects in a dose-dependent fashion that was similar to flutamide: reduced anogenital distance in males, increased incidence of thoracic nipple development, hypospadias, absent prostate, absent or partially developed epididymis, absent or underdeveloped seminal vesicles, and degradation of seminiferous epithelium. However, there were some differences between Dibutyl Phthalate and flutamide effects. For example, there was some variation in the sensitivities of androgen-dependent tissues between Dibutyl Phthalate and flutamide. Inguinoscrotal descent, an androgen-dependent event was impaired by flutamide and not Dibutyl Phthalate, whereas transabdominal descent (an androgen-independent event) was affected by Dibutyl Phthalate and not by flutamide. Other androgen-independent events were affected by Dibutyl Phthalate, as well (no further details).</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>not reported</i> | <i>groups of 19 - 20 pregnant CD rats</i> | <i>0, 0.5, 5, 50, or 100 mg/kg/d</i> | <i>Dams were exposed to test material on gestation days 12 to 21; beginning on postnatal day 1, offspring pups were examined for signs of developmental toxicity, including pup survival, clinical signs of toxicity, pup weight, and anogenital distance. On postnatal day 14, the numbers and locations of nipples and areolas on male pups were recorded. The days of acquisition of preputial separation or vaginal opening were recorded for male or female pups, respectively. Male pups were killed for necropsy on postnatal day 110 ± 10 days, and females were killed for necropsy on postnatal day 80 ± 5 days.</i> | <i>There were no effects of Dibutyl Phthalate treatment on rate of pup survival at birth. On postnatal day 14, there was an increase in female pup weights at doses 5 mg/kg/d and higher, but this increase was transient and not observed at later time points. The following observations were made of male pups of the 500 mg/kg/d group: reduced anogenital distance; hypospadias; absent or partially developed epididymis, vas deferens, seminal vesicles, and ventral prostate; decreased weights of testes, epididymis, dorsolateral and ventral prostates, seminal vesicles, and levator ani-bulbocavernosus muscle; seminiferous tubule degradation; focal interstitial cell hyperplasia; and interstitial cell adenoma. Male pups of the 100 and 500 mg/kg/d groups showed a dose-dependent increase in the number of retained areolas or nipples, and androgen-dependent event. The nipple retention at 100 mg/kg/d represented the LOAEL for Dibutyl Phthalate in this study. The NOAEL in this study is 50 mg/kg/d</i> | ⁸ |
| <i>Dibutyl Phthalate</i> | <i>corn oil</i> | <i>8 pregnant Sprague-Dawley rats</i> | <i>0 or 500 mg/kg/d</i> | <i>Dams were exposed to test material from gestation day 14 to postnatal day 3; at 5 mo of age, male offspring (50 control, 48 Dibutyl Phthalate-exposed) were killed and examined externally and internally for indications of reproductive development; in another experiment of the same report, five pregnant LE rats were given daily oral doses of 500mg/kg on gestation days 16 to 19. Six control rats received corn oil on the same schedule. At 5 months of age, male offspring were examined.</i> | <i>In treated males, there was a significant increase in number of retained nipples ($p < 0.0003$), a decrease in anogenital distance ($p < 0.0001$), and decreases in the weights of ventral prostate, epididymis, cauda epididymis, testes, glans penis, and levator ani-bulbocavernosus ($p < 0.0001$ to 0.03). There was also a non-significant increase in the incidence of hypospadias in treated males.</i> | ⁸ |

Table 6. Developmental and reproductive toxicity studies from previous reviews

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--------------------------|-----------------|---------------------------|-------------------------|--|---|--------------|
| <i>Dibutyl Phthalate</i> | <i>corn oil</i> | <i>5 pregnant LE rats</i> | <i>0 or 500 mg/kg/d</i> | <i>Dams were exposed to test material from gestation day 16 to 19; at 5 mo of age, male offspring were examined.</i> | <i>In the treated males, anogenital distance was reduced ($p < 0.0008$) and number of retained nipples was increased ($p < 0.032$). There were reductions in the weights of seminal vesicle, ventral prostate, and levator ani-bulbocavernosus ($p < 0.03$) of treated males. Hypospadias was not seen in treated rat offspring</i> | ⁸ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--|--|--|--|---|---|---------------|
| EMBRYOTOXIC AND TERATOGENIC EFFECTS | | | | | | |
| IN VITRO | | | | | | |
| Dibutyl Phthalate and Diethyl Phthalate | DMSO for Dibutyl Phthalate; none for Diethyl Phthalate | <i>Xenopus laevis</i> embryos, 20 embryos/dose | 0, 1, 3, 5, 7.5, 10, 12.5, or 15 ppm for Dibutyl Phthalate 0, 25, 50, 100, or 200 ppm for Diethyl Phthalate | Effects of Dibutyl Phthalate or Diethyl Phthalate on embryos were analyzed using the 96-h frog embryo teratogenesis assay. Treatments were performed in triplicate. | <p>The mean 96-h LC₅₀ obtained from embryos exposed to Dibutyl Phthalate was 12.88 ppm. The teratogenic index for Dibutyl Phthalate was 1.56. The NOEC for malformations was lowest at 5.8 ppm, indicating malformations as the most sensitive endpoint. The mean minimum concentration to inhibit growth was 8.75 (based on 2 of 3 experiments). Embryos from 5 ppm Dibutyl Phthalate exposures had notable edema, loose gut coiling, craniofacial abnormalities, and few reduced eye abnormalities. Embryos exposed to 7.5 or 12.5 ppm concentrations had similar effects, with hemorrhaging as an additional malformation. Severe abnormalities at 15 ppm were observed. A tumor-like growth was observed in one embryo.</p> <p>The mean 96-h LC₅₀ obtained from embryos exposed to Diethyl Phthalate was 64.5 ppm. The teratogenic index for Diethyl Phthalate was 1.25. The NOEC for malformation was lowest at 17 ppm, indicating malformations as the most sensitive endpoint for Diethyl Phthalate. The mean minimum concentration to inhibit growth was 41.7 ppm. Embryos exposed to Diethyl Phthalate for 96 h had edema and loose gut coiling at concentrations at 50 ppm and higher, with notable muscular tail kinking, kidney abnormalities, hemorrhaging, craniofacial abnormalities, and a reduced eye abnormality observed at 100 ppm.</p> | ³⁷ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------------------|------------------------------------|--|--|--|---|---------------|
| Dibutyl Phthalate, > 99% pure | ethanol or DMSO, details not clear | <i>Xenopus laevis</i> embryos, 25 - 50 embryos/dose | 0, 1, 25, or 200 µM | Embryos were treated 1.5 h after fertilization with control or test material, and were analyzed for abnormalities and mortalities for 2 h intervals for the first 12 h and every 24 h after for up to 96 h. Treatments were performed in duplicate or triplicate | Normal development was observed in the control groups. Many embryos treated with the 3 concentrations of Dibutyl Phthalate had curved body axes in the lateral direction; head, eye, and mouth irregularities; hypertrophied gut and blisters/edema; an overall delayed development, with all embryos in the 200 µM group exhibiting these defects. The mean percent normal development over the span of 96 h decreased as Dibutyl Phthalate concentration and exposure time increased, with the greatest numbers of defects occurring at stage 25 ($p < 0.005$). Many embryos exposed to lower concentrations seemed to self-correct by late tadpole stage 46, as indicated by increased numbers of normal embryos. Dibutyl Phthalate did not result in increased mortality, although the number of normal embryos was still lower than controls at all time points. | ³⁸ |
| Dibutyl Phthalate, > 99% pure | DMSO | <i>Xenopus laevis</i> embryos, 40 embryos/dose | 0, 5, 10, 12.5, 15, 17.5, 20, or 30 mg/l | Embryos were treated starting at blastula stage with control or test material. At 24 h intervals, mortalities were determined and after 96 h, surviving embryos (tadpole stage) were counted and fixed for examination. RNA of embryos was also analyzed. | At 20 mg/l, embryonic survival was significantly decreased from 72 h onward, and at 30 mg/l, most of the embryos (67.5%) died before 48 h and all embryos died by 72 h. The 96 h LC ₅₀ value of Dibutyl Phthalate was calculated to be 18.3 mg/l. Tadpoles that received Dibutyl Phthalate had significantly decreased body length, tail length, and total length at 10, 5, and 10 mg/l, respectively. Treated tadpoles had significantly increased occurrences of abnormal gut coiling, head malformation, and ventral blister at 10 mg/l. The EC ₅₀ value of Dibutyl Phthalate for malformations by 96 h was calculated to be 7.5 mg/l, although Dibutyl Phthalate did not change the body length parameters of tadpoles. <i>BiP</i> mRNA levels were significantly increased at 1 mg/l in Dibutyl Phthalate-treated tadpoles, but <i>atf4</i> mRNA levels were not significantly changed. Spliced <i>xbp1</i> mRNA was significantly increased at 1 mg/l Dibutyl Phthalate. <i>CHOP</i> mRNA was not significantly changed up to 20 mg/l Dibutyl Phthalate. <i>bcl2</i> , <i>bax</i> , <i>bad</i> , and <i>bak</i> mRNA levels were not significantly changed but the ratio of <i>bax/bcl2</i> mRNA was increased at 5 mg/l. | ³⁹ |
| Dibutyl Phthalate, 99% pure | not reported | <i>Danio rerio</i> embryos, wild type (AB strain); number not reported | 0, 50, or 250 µg/l | Dibutyl Phthalate was dissolved in acetone. Fish embryos were exposed to different concentrations of Dibutyl Phthalate for up to 96 h. Embryos were observed for deaths or developments daily. Teratogenic effects were recorded. Body length along with yolk sac width and height were measured. Lipidomic analysis, immune response, gene expression analysis, and an estrogenic activity test were performed. | No significant effects on survival or hatching of embryos following exposure to Dibutyl Phthalate were observed. Dibutyl Phthalate induced body length decrease, yolk sac abnormalities, and immune responses (up-regulation of immune proteins and genes). At 50 µg/l, Dibutyl Phthalate significantly reduced fatty acid, triglycerides, diacylglycerol, and cholesterol levels. Co-exposure to Dibutyl Phthalate and estrogenic receptor antagonist did not significantly relieve the toxic symptoms compared with exposure to Dibutyl Phthalate alone. | ⁴⁰ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--|--------------|--|---|--|--|-----------|
| Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate | DMSO | <i>Danio rerio</i> embryos, 30 per treatment | 0.0005 - 0.2 mg/l for Dibutyl; 1.5 - 10 mg/l for Diethyl; 1.5 - 8 mg/l for Dimethyl | The acute developmental toxicity of Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate in addition to 3 other phthalate esters on zebrafish embryos; control was DMSO. Fertilized eggs were exposed with 5 ml of the test solution; embryos observed at 12, 24, 48, 72, 96, 120 h post-fertilization. Alcian blue and alizarin red double staining were performed to detect skeletal development of the fish larvae | All phthalates tested induced different developmental abnormalities in the fish larvae, including abnormal movement, decreased heart rate, spinal curvature, and pericardial edema. Bone development was also affected. Malformation rate caused by Dibutyl Phthalate at 0.05 mg/l was 30%, and exposure to higher than that exposure concentration led to death. Dibutyl Phthalate caused the greatest mortality in the fish of the phthalates tested, even at low concentrations; mortality rates varied with the exposure concentrations of the phthalates. Dimethyl Phthalate was not correlated with larvae mortality. Gene expression levels of skeleton-related genes showed the upregulation of <i>runx2b</i> and <i>shha</i> genes after Dibutyl Phthalate exposure | 41 |
| DERMAL | | | | | | |
| Diethyl Phthalate | not reported | groups of 17-20 female Jcl:ICR mice | 0, 500, 1600, or 5600 mg/kg/d | Test material was applied to the skin of pregnant mice on gestation days 0-17; no further details available | NOAEL < 500 mg/kg/d in dams; dose-related abnormal behavior (possibly caused by pain) was observed; reduced thymus and spleen weights were observed at all doses and increased adrenal gland weights at 5600 mg/kg/d NOAEL = 1600 mg/kg/d in offspring; when compared to controls, fetal body weight was significantly reduced in the 5600 mg/kg/d group; an increase in the incidence of cervical and lumbar rib variations was also observed; no external, visceral, or skeletal anomalies were observed in the fetuses that were attributed to treatment with Diethyl Phthalate | 6 |
| ORAL | | | | | | |
| Dibutyl Phthalate; purity not reported | olive oil | Pzh:Sfis outbred mice F ₀ and F ₁ : 18-20 males/dose group mated with 2 females each | 0, 500, or 2000 mg/kg bw | A 3-generation reproductive and developmental toxicity study. Male mice were exposed to test material via gavage 3 d/wk for 8 wk; 6 – 7 males from each dose group were killed at 4, 8 and 12 wk after the start of exposure to examine sperm count and quality. At the end of exposure, remaining males mated with 2 unexposed females each; ¼ of females from each dose group were killed 1 d before expected parturition while other females were allowed to deliver and rear litters. Males from the F ₁ generation were mated with females from the same group but different litter in order to examine prenatal development of the F ₂ generation; remaining F ₁ males were killed at the same age to check sperm count and quality. A comet assay was utilized to determine induction of DNA breaks. | Exposure of F ₀ males to Dibutyl Phthalate induced skeletal malformations in surviving fetuses, causing significant mortality in postnatal life and a skewing of the sex ratio (F ₁ females had higher survival rates); increased frequency of DNA damage in the germ cells was observed in F ₁ males; exposure of F ₀ males to Dibutyl Phthalate did not affect F ₁ male fertility and pregnancy frequency; mean numbers of total live and dead fetuses were not significantly different compared to the control | 42 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------------------|--------------|--|--|--|--|---------------|
| Dibutyl Phthalate; > 99% pure | dietary feed | Sprague Dawley rats; 45 - 47 time mated female rats in F ₀ generation/ dose group; 50 male and 50 female rats in F ₁ generation/dose group | 0, 300, 1000, 3000, or 10,000 ppm, equivalent to 16 - 17, 54 - 57, 152 - 169, and 510 - 600 mg/kg bw/d, respectively | 2-yr carcinogenicity study (see the Carcinogenicity Studies section for full details); body weights and feed consumption measured (both F ₀ and F ₁), clinical observations made through study period, and complete necropsies and microscopic examinations performed on all F ₁ rats at study end | <p>Maternal mean body weights and body weight gains of the exposed dams during gestation were comparable to the control group; during lactation, minimal differences in mean body weights noted between the high dose group and the control group; feed consumption during gestation significantly ($p < 0.05$; $p < 0.01$) higher only in the 10,000 ppm compared to control group on gestation days 12 - 18, feed consumption during lactation was significantly ($p < 0.05$; $p < 0.01$) increased compared to the control group on days 1 - 4 in all groups except the high dose group</p> <p>In the 10,000 ppm dose group male offspring, high incidence of small or absent organs of the male reproductive tract and undescended testes observed; some gross lesions correlated with microscopic lesions in the testes (germinal epithelium atrophy), epididymis (hypospermia), and prostate and seminal vesicles (decreased secretory fluid) also observed in this dose group. Additional microscopic lesions in the reproductive tract in rats included seminiferous tubule dysgenesis, testicular interstitial cell hyperplasia, testicular edema, and fibrosis and granuloma of the rete testis.</p> | ⁴⁴ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------------------|--------------|--|--|--|--|-----------|
| Dibutyl Phthalate | dietary feed | groups of 17 male and 17 female Sprague-Dawley rats | 1, 4, 10, 30, 100, 1000, or 10,000 ppm | 2-generation reproduction toxicity study; 1 ppm dose group served as control. The F ₀ generation received test material in feed as adults starting at study day 1 until day before necropsy (weeks 29-32), and were bred to produce F _{1a} , F _{1b} , and F _{1c} offspring. F _{1b} adults were reared and bred to produce F _{2a} , F _{2b} , and F _{2c} offspring. An "Outbreeding Cohabitation" was conducted by mating 10,000 ppm F ₁ males with naive females and 10,000 ppm F ₁ females with naive males. Body weights, feed consumption, clinical observations, vaginal cytology, reproductive performance, neonatal anogenital distance, pup survival, sexual development, computer-assisted sperm analyses, gross pathology, organ weights, and selected histopathology were performed over the course of the study. | <p>No adverse effects in body weight gains, feed consumption, clinical observations, incidental gross findings, and mortality were observed in any dose levels of Dibutyl Phthalate during any generation of the Outbreeding Cohabitations. Deaths observed during the F₀ and F₁ cohabitations were not treatment-related.</p> <p>No consistent treatment-related effects were observed in litter data from the F₀ and F₁ cohabitations. In the F₀ cohabitation, the mean anogenital distance in the 10,000 ppm F_{1a} male pups was decreased in a statistically significant manner by 15% on postnatal day 1, while the anogenital distance of the 10,000 ppm F_{1b} male pups was not decreased significantly. During F₁ cohabitations, the mean anogenital distance of 10,000 ppm F_{2a}, F_{2b}, and F_{2c} male pups decreased in a statistically significant manner by 13 - 15%. The mean anogenital distance in the Outbreeding Cohabitation of male pups born to 10,000 ppm F₁ dams decreased by a statistically significant 14%. A statistically significant mean delay in preputial separation and testicular descent was observed in 10,000 ppm F_{1b} males when compared to controls.</p> <p>Seminiferous tubular atrophy was noted in 10,000 ppm F₁ males. No other adverse effects were noted in gross pathology, organ weight changes, or microscopic pathology in adult F₀ or F₁ rats. No changes were noted in the sperm endpoints in the F₀ or F₁ males.</p> | 43 |
| Diethyl Phthalate, 99.8% pure | dietary feed | groups of 24 male and 24 female Crj:CD (SD) IGS rats | 0, 600, 3000, or 15,000 ppm | Two-generation reproduction toxicity study performed in accordance with OECD TG 416; males received treated feed for 15 wk and females received treated feed for 17 wk in the F ₀ and F ₁ generations; F ₁ animals were mated at 10 wk of age | <p>NOAEL for general toxicity and reproductive performance in the parental animals is 15,000 ppm (equivalent mean intake = 1016 - 1297 mg/kg/d). No adverse effects observed in clinical findings, body weights, feed consumption, reproductive parameters, or gross or histopathological findings in any treated group of the F₀ and F₁ parental animals. F₀ males in the 15,000 ppm group had increased liver weights and enhanced activities of metabolic enzymes. This group of males also had an increase in the content of CYP3A2, and a decrease in serum testosterone levels was observed in F₀ males of the 3000 and 15,000 ppm groups; however, these effects were not considered adverse since it did not affect the ability to produce offspring.</p> <p>NOAEL for development and growth of pups is 3000 ppm (equivalent mean intake = 222 - 267 mg/kg/d) due to decreased body weight gain at 15,000 ppm (equivalent mean intake = 1150 - 1375 mg/kg/d). Body weight gains before weaning were inhibited in F₁ and F₂ pups and vaginal openings were slightly delayed in F₁ females at 15,000 ppm.</p> | 6,45 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--|---------|---|--------------------|---|--|-----------|
| EFFECTS IN FEMALE REPRODUCTIVE ORGANS | | | | | | |
| <i>IN VITRO</i> | | | | | | |
| Dibutyl Phthalate and monobutyl phthalate | DMSO | ovarian antral follicles from CD-1 mice, 6 - 12 follicles per treatment group | 0.001 - 1000 µg/ml | Mouse ovarian antral follicles were treated with vehicle, Dibutyl Phthalate, or monobutyl phthalate for 24 - 72 h. Follicle diameter, ATP production, q-PCR, and TUNEL were used to measure follicle growth, viability, cell cycle and apoptosis gene expression and cell death-associated DNA fragmentation, respectively. | <p>After 24 h, follicles treated with Dibutyl Phthalate at 1000 µg/ml were significantly smaller than controls, growth inhibition was evident in follicles treated with 500 and 1000 µg/ml after 48 h, and follicles treated with all concentrations of Dibutyl Phthalate above 100 µg/ml after 72 h were significantly reduced compared to the controls ($p \leq 0.05$). Dibutyl Phthalate exposure at 10 µg/ml and greater resulted in a significant decrease in the percentage of follicles growing. Cytotoxicity was observed to Dibutyl Phthalate at 500 µg/ml and greater. Expression of <i>Ccnd2</i> was decreased in follicles treated with Dibutyl Phthalate at 1000 µg/ml at both 48 and 72 h. The expression of the cell cycle arrest gene <i>Cdkn1a</i> was significantly increased in follicles treated with the highest concentration of Dibutyl Phthalate at 48 h. <i>Cdkn2a</i> expression was increased at 48 h in follicles treated with Dibutyl Phthalate at 0.001 and 10 - 1000 µg/ml, but returned to levels similar to the controls by 72 h. <i>Bax</i> and <i>Bid</i> expression were increased at 48 h in follicles treated with 1000 µg/ml Dibutyl Phthalate in the absence of <i>Bcl2</i> expression changes. <i>Bax</i> expression returned to control levels by 72 h, but <i>Bid</i> mRNA levels remained increased and included an increase in follicles treated with 100 µg/ml Dibutyl Phthalate. A statistically significant decrease in the expression of <i>Bcl2</i> mRNA in follicles treated with 1 µg/ml Dibutyl Phthalate for 72 h. In assessing antral follicle death, a statistically significant difference in DNA fragmentation levels relative to the was observed in follicles treated with 10 and 1000 µg/ml Dibutyl Phthalate.</p> <p>No toxicity to monobutyl phthalate was observed.</p> | 46 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------------------|------------------------------|---|---|--|--|-----------|
| ORAL | | | | | | |
| Dibutyl Phthalate | tocopherol-stripped corn oil | groups of 16 female CD-1 mice | 10 or 100 µg/kg/d or 1000 mg/kg/d | Study to investigate Dibutyl Phthalate exposure on IGF signaling in the ovary and effects on ovarian folliculogenesis; rats received test material daily for 20 consecutive days; animals were subjected to vaginal smears, weighed, and then killed as they reached the proestrus stage; ovaries were dissected from each animal and rid of fat and oviductal tissue; one ovary from each pair was snap-frozen for RNA extraction, while the other was fixed for subsequent histological processing; levels of mRNAs encoding IGF1 and 2 (<i>Igf1</i> and <i>Igf2</i>), IGF1 receptor (<i>Igf1r</i>), and IGF binding proteins 1-6 (<i>Igfbp1-6</i>) were measured in whole ovary homogenates; ovarian follicle counts and immunostaining for phosphorylated IGF1R protein (pIGF1R) were used to evaluate folliculogenesis and IGF1R activation, respectively | Dibutyl Phthalate at 100 µg/kg/d significantly decreased ovarian <i>Igf1</i> (p = 0.028) and <i>Igf1r</i> (p = 0.048) mRNA expression without affecting <i>Igf2</i> (p = 0.997); however, the mRNA levels for all 3 genes did not differ from controls at 10 µg/kg/d or 1000 mg/kg/d. Dibutyl Phthalate did not cause significant deviations in the expression of transcripts encoding IGF binding proteins when compared to the vehicle control; total number of ovarian follicles counted per ovary was reduced in the 100 µg/kg/d group (p = 0.02) when compared to controls; however no significant differences were observed in the 10 µg/kg/d or 1000 mg/kg/d dose groups. Significantly low primordial follicle counts observed in the 100 µg/kg/d group (p=0.006) and in the 1000 mg/kg/d group (p=0.02) compared to controls, also fewer primary follicles were observed in 100 µg/kg/d mice (p=0.04). Immunostaining for pIGF1R showed a significantly reduced <i>Igf1r</i> mRNA and primordial and primary follicle numbers in 100 µg/kg/d mice, but not in the other doses | 47 |
| Dibutyl Phthalate, 99.6% pure | tocopherol-stripped corn oil | groups of 7 or 8 female CD-1 mice | 0, 10, 100, or 1000 µg/kg/d | Females were treated with vehicle or test material for 30 d via oral pipette. At the end of dosing, the mice were killed and ovaries were removed and analyzed for classification and enumeration of ovarian follicles and corpora lutea, RNA extraction, and cDNA synthesis. Gene expression was determined using PCR, and immunostaining was performed for phosphorylated H2AX. Protein was extracted and analyzed with SDS-PAGE and western blotting. | No overt clinical toxicity was observed in the mice. A non-statistically significant increase in body weight gain was observed in the 100 and 1000 µg/kg groups during dosing. No differences in the percentage of time spent in proestrus, estrus, diestrus phases in the treatment groups before or during dosing were noted. Dose-dependent effects on folliculogenesis and gene expression were observed. At 1000 µg/kg, more atretic follicles were observed in the ovaries. Follicle numbers in the 10 and 100 µg/kg doses were comparable to controls. In a dose-dependent manner, Dibutyl Phthalate significantly reduced the expression of genes responsible for homologous recombination (<i>Atm</i> , <i>Brca1</i> , <i>Mre11a</i> , and <i>Rad50</i>), mismatch repair (<i>Msh3</i> , <i>Msh6</i>), and nucleotide excision repair (<i>Xpc</i> , <i>Pcna</i>); however, staining for γH2AX was similar between doses. No differential DNA methylation in the <i>Brca1</i> promoter was observed, but significantly reduced transcript levels for <i>Dnmt1</i> was observed in the ovary. | 48 |
| Dibutyl Phthalate | corn oil | groups of 8 or 9 pregnant Sprague-Dawley rats | 0 or 500 mg/kg bw every second day starting at gestation day 14.5 | Dams received test material on gestation days 14.5 to postnatal day 6 via oral gavage. Corn oil was also subcutaneously injected on gestation days 14.5 and 16.5. Pups were culled on gestation days 10, 24, and 90, and surviving pups were weaned on gestation day 21. Anogenital distance was measured on all pups on postnatal days 2 and 10. Vaginal opening was checked daily from postnatal day 29 onwards in female offspring. | Dams and offspring exhibited no adverse effects. Dams in the control and treatment group delivered exclusively live pups, although there was some subsequent perinatal mortality. Female pups showed no effects to anogenital distance. No adverse effects were observed in the vaginal opening when compared to controls. | 49 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|------------------------------|--|---|--|--|-----------|
| Dibutyl Phthalate | corn oil | immature female rats, at least 6 per group, strain not specified | 0, 10, or 100 mg/kg | In this uterotrophic assay, rats received Dibutyl Phthalate orally from postnatal day 21 for 3 d. Body weights recorded before treatment, during treatment, and prior to necropsy. Clinical signs and symptoms observed daily. On the 4 th day, rats were killed and the uterus was dissected out. Uteri and ovaries were weighed with luminal contents. | No abnormal clinical signs or symptoms were observed in any treated or control animal during study period. Body weight increased significantly in the 10 mg/kg dose group, but was not significantly elevated in the 100 mg/kg dose group. In the 100 mg/kg Dibutyl Phthalate group, the uterine wet weight was significantly decreased, and there were minor variations in the ovary wet weight. | 50 |
| Dibutyl Phthalate | corn oil | immature female rats, at least 6 per group, strain not specified | 0, 10, or 100 mg/kg | In this pubertal onset assay, rats received Dibutyl Phthalate orally from postnatal day 21 for 20 d. Rats examine daily and body weight was recorded before treatment, during treatment, and prior to necropsy. The rats were examined daily for vaginal opening from beginning of postnatal day 21. On the day of vaginal opening, body weight and age was recorded. Vaginal lavage was collected on the day of vaginal opening and estrous cycle was evaluated and stages determined. Rats were killed on postnatal day 42. The uterus, vagina, and ovaries were weighed. | All rats from the treated group and the control group gained body weight during the study period; however, a statistically significant decline in body weight gain was observed at postnatal days 27, 33, and 42, in both treated groups. The weight of uterus and ovary declined significantly and changes in vaginal weight were nonsignificant in both treated groups. However, vaginal opening was not observed in any of the animals in the controls and treated groups until postnatal day 42, except in one animal each in vehicle control and the 100 mg/kg dose group. | 50 |
| mixture of phthalates including 15% Dibutyl Phthalate and 35% Diethyl Phthalate | tocopherol-stripped corn oil | groups of 12 pregnant CD-1 mice in F ₀ generation | 0, 20 µg/kg/d, 200 µg/kg/d, 200 mg/kg/d, or 500 mg/kg/d | Dams orally dosed with control or phthalate mixture daily from gestation day 10 to birth. Adult F ₁ females born to these dams were used to generate the F ₂ generation and adult F ₂ females born to F ₁ females were used to generate the F ₃ generation. Sera, ovaries, uteri, and livers were collected and organ weights, body weights, and anogenital distances were measured from F ₂ and F ₃ females on postnatal days 1, 4, 8, 21, 60 and at 13 mo of age. Fertility tests were also performed in the F ₂ and F ₃ generations. | In the F ₂ generations, prenatal exposure to the phthalate mixture containing Dibutyl Phthalate and Diethyl Phthalate increased uterine weight, anogenital distance, and body weight. Cystic ovaries and breeding and pregnancy complications were observed. In the F ₃ generation, uterine weight was increased, anogenital distance was decreased, and fertility complications were observed. | 51 |
| EFFECTS IN MALE REPRODUCTIVE ORGANS | | | | | | |
| IN VITRO | | | | | | |
| Dibutyl Phthalate | not reported | TM3 and TM4 mouse cells | 0, 5, 10, 50 or 100 mg/l | An in vitro mechanistic study exposing cells to test material for 24-h | Dibutyl Phthalate exposure produced dose-dependent evidence of ferroptosis. Dibutyl Phthalate treatment decreased cellular GSH and mitochondrial membrane potential, and increased MDA, ROS, GSSG, and Fe ²⁺ levels. Morphological evaluation by transmission electron microscopy revealed mitochondrial shrinkage and cristae loss. Real-time PCR and Western blot analyses showed that Dibutyl Phthalate selectively upregulated PRDX6, a negative regulator of ferroptosis, while GPX4 remained unchanged. Further, Dibutyl Phthalate upregulated SP1 expression, which can directly bound to the PRDX6 promoter and transcriptionally activated its expression. | 52 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|---|---|---|--|---|-----------|
| Dibutyl Phthalate | not reported | rat testicular explants | 10 ⁻⁶ , 10 ⁻⁵ , or 10 ⁻⁴ M | Explants were exposed to test material for 24 h to evaluate the effects on juxtacrine communication through Notch signaling pathway | Real-time PCR and Western blotting showed significant upregulation of Notch1, Dll4, and Hey1 at both mRNA and protein levels following Dibutyl Phthalate exposure (p < 0.05; p < 0.01; p < 0.001), while Hes1 expression remained unchanged. Dibutyl Phthalate altered immunoexpression of activated NOTCH1, DLL4, HEY1 and HES5 both in seminiferous epithelium and interstitial tissue, with differential effects across cell types. | 53 |
| Dibutyl Phthalate | DMSO | Leydig cells isolated from male Sprague-Dawley rats | 50 mg/l | Effects of Dibutyl Phthalate on Leydig cells and its mechanism related to gap junction investigated; cells were cultured for 24 h, Dibutyl Phthalate, Dibutyl Phthalate + 10 µM prostaglandin E2, or 40 µM flutamide; radioimmunoassay, semi-quantitative RT-PCR, immunofluorescence, and Western blot were utilized to determine the expression of testosterone and Connexin 43 (Cx43) in the Leydig cells | Expression of testosterone was significantly decreased (p < 0.05) in the Dibutyl Phthalate, Dibutyl Phthalate + prostaglandin E2, and flutamide groups. A significant decrease (p < 0.05) of Cx43 was observed in the Dibutyl Phthalate and flutamide group. Cx43 was up-regulated with the administration of prostaglandin E2, but there was no significant change in testosterone. Testosterone was down-regulated with a significant decrease of Cx43 in the flutamide group. | 54 |
| Dibutyl Phthalate, 99.9% pure | not reported | human sperm from 12 healthy volunteers | 13.47, 67.35, or 134.7 µg/ml | Healthy mature spermatozoa were seeded in 96-well plate in complete Minimum Essential Medium (MEM) culture medium supplemented with Dibutyl Phthalate. The cells were incubated and analyzed for motility, and percent cell viability was assessed between 30 min and 96 h. | A concentration- and duration-dependent decrease in motility was noted. The decline varied from 20%, 33%, 66%, with respect to control, after 18 h exposure to Dibutyl Phthalate at the low, medium, and high concentrations, respectively. | 55 |
| Dibutyl Phthalate, > 99% pure and monobutyl phthalate | human tubal fluid medium containing 1% DMSO | human sperm | 2 nM to 6 µM for Dibutyl; 1 nM to 3 µM for monobutyl | Effects on human sperm function tested in vitro with Dibutyl Phthalate and monobutyl phthalate; sperm treated with Dibutyl Phthalate, monobutyl phthalate, or a mixture of both (2 nM Dibutyl Phthalate + 1 nM monobutyl phthalate to 6 µM Dibutyl Phthalate + 3 µM monobutyl phthalate) for 1 to 4 h; controls were the medium without the phthalates | Results showed that only Dibutyl Phthalate at 6 µM, 3 µM monobutyl phthalate, and the mixture had adverse effects on sperm motility, penetration ability, and capacitation. Additionally, human sperm tyrosine phosphorylation was suppressed at these doses. | 56 |
| ORAL | | | | | | |
| Dibutyl Phthalate; 99.5% pure | corn oil and Tween-80 | groups of 20 female Sprague-Dawley rats | 0, 50, 250, or 500 mg/kg bw/d | Pregnant rats received test material via gavage on gestation day 1 through postnatal day 21. Pregnant rats were weighed and observed daily for clinical signs of toxicity. The developmental condition of F ₁ rats and reproductive system of mature F ₁ male rats were assessed (groups of 20 males per dose group up to postnatal day 70). Dams and female pups were killed on postnatal day 21. | NOAEL = 50 mg/kg bw/d for developmental toxicity. No adverse effects were observed in the dams. At doses of 250 mg/kg and higher, reduced birth weight, body weight gain, number of live pups per litter, anogenital distance (males), epididymis weight, and sperm count and motility were observed. Significantly reduced prostate weight, absent or underdeveloped epididymis, undescended testes, and testicular atrophy were also observed in mature F ₁ males at 250 mg/kg and higher. | 57 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------|----------|---------------------------------------|---------------------|--|--|---------------|
| Dibutyl Phthalate | corn oil | groups of 5 - 10 pregnant Wistar rats | 0 or 500 mg/kg bw/d | Dams rats received Dibutyl Phthalate via gavage on gestation days 13 - 21. Male offspring of both the control and test material groups were tested for fertility in adulthood by mating with a female of proven fertility. Cell-specific immunohistochemistry and confocal microscopy were used to track development of Sertoli, Leydig, germ, and peritubular myoid cells from male fetal life to adulthood. Adult testes of the male offspring of the control and test material groups were examined histologically. Plasma and testicular testosterone were analyzed. | No treatment-related effects were observed in the maternal animals. A high rate (> 60%) of cryptorchidism (mainly unilateral), infertility, hypospadias, and testes abnormalities observed. In scrotal and cryptorchid testes of Dibutyl Phthalate-exposed males, areas of focal dysgenesis were found that contained Sertoli and Leydig cells, and gonocytes and partially formed testicular cords; these dysgenetic areas were associated with Leydig cell hyperplasia at all ages. Suppression (~90%) of testicular testosterone levels on gestation day 19 in Dibutyl Phthalate-exposed males, coincident with delayed peritubular myoid cell differentiation, may have contributed to the dysgenesis. Double immunohistochemistry revealed immature Sertoli cells in dysgenetic areas. The Dibutyl Phthalate-exposed males also exhibited Sertoli cell-only (SCO) tubules, sporadically in scrotal and predominantly in cryptorchid, testes, or foci of SCO within normal tubules in scrotal testes. In all SCO areas the Sertoli cells were immature. Intratubular Leydig cells were evident in the exposed animals and, where these occurred, Sertoli cells were immature and spermatogenesis was absent. Abnormal Sertoli cell \pm gonocyte interaction was evident at gestation day 19 in the test material-exposed rats coincident with appearance of multinucleated gonocytes, although these disappeared by postnatal day 10 during widespread loss of germ cells. | ⁵⁹ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------|----------|--------------------------------------|--------------------|--|--|---------------|
| Dibutyl Phthalate | corn oil | groups of pregnant Crl:CD(SD)Br rats | 0 or 500 mg/kg/d | Dams received test material via gavage on gestation days 12 -21. The dams were examined at dosing and several h post-dosing for clinical signs of toxicity. Body weights were recorded daily. Male fetuses or pups were necropsied on gestation days 16, 17, 18, 19, 20, and 21 and on postnatal days 3, 7, 16, 21, 45, and 70. Fetal and pup weights were recorded at necropsy. Anogenital distance and number of areolae were recorded on postnatal days 1 and 13. The reproductive tracts of all male fetuses were removed using a dissecting microscope, and the testes and epididymides were examined. In male pups killed on postnatal days 3, 7, 16, and 21, the testes, epididymides, and vasa deferentia were removed and analyzed. Kidneys and livers were also analyzed in the pups from postnatal days 16 and 21. Pups from postnatal days 45 and 70 had the testes, epididymides, vasa deferentia, prostates, seminal vesicles, levator ani and bulbocavernosus muscles, liver, and kidneys dissected for analysis. | No maternal toxicity was observed, and no decreases in litter size, pup survival, or body weights of offspring were observed. Examination of the phthalate-exposed males revealed gross or histologic lesions at all the time points chosen. A statistically significant decrease (14%) in anogenital distance was observed on postnatal day 1 and a significant increase in the number of retained areolae in the exposed pups was observed on postnatal day 13. In the fetal testes, large aggregates of Leydig cells, multinucleated gonocytes, and increased numbers of gonocytes were first detected on gestation day 17 and increased in incidence to 100% by gestation day 20 and 21. These lesions resolved during the early postnatal period, while decreased numbers of spermatocytes were noted on postnatal day 16 and 21. On postnatal day 45, there was mild degeneration of the seminiferous epithelium, which progressed to severe seminiferous epithelial degeneration on postnatal day 70. This degeneration was concurrent with ipsilateral malformed epididymides, which caused obstruction of testicular fluid flow and secondary pressure atrophy in the seminiferous tubules. In the fetus, the epididymal lesion was observed as decreased coiling of the epididymal duct. The decreased coiling progressed into the early postnatal period and adulthood, at which time malformed epididymides were apparent. | ⁶¹ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-----------------------------|----------|--|-----------------------------|--|--|---------------|
| Dibutyl Phthalate, 99% pure | corn oil | groups of female Wistar rats; numbers per group not clearly reported | 0, 4, 20, 100, or 500 mg/kg | Pregnant rats received Dibutyl Phthalate via gavage on gestation days 13.5 to either 20.5 or 21.5. Male fetuses were analyzed for testis dysgenesis. Fetal testis weight and testicular testosterone levels were evaluated. A portion of the male offspring were allowed to live until 90 d of age for mating, following which the males were killed and testes were also examined | A significant decrease in testis weight was observed in offspring in the 500 mg/kg dose group at gestation day 21.5 and in adulthood. Animals exposed to the other dose levels did not have any significant change in testis weight at either gestation day 21.5 or adulthood. Testicular testosterone levels in fetal animals were significantly decreased in the 100 and 500 mg/kg dose groups when compared to controls. No significant effects of Dibutyl Phthalate treatment were observed on fetal body weight or on litter size. Fetal males of the 100 and 500 mg/kg dose groups had a significant increase in the occurrence of multinucleated gonocytes and changes in Leydig cell distribution compared to controls. A significant decrease in total Leydig cell cluster number per testis section was also observed in the 100 and 500 mg/kg dose groups. In the adult male offspring exposed to 500 mg/kg Dibutyl Phthalate, a statistically significant 75% infertility rate was observed. The infertility rates for the 100 and 20 mg/kg groups were 33 and 14%, respectively, and were not statistically significant. The 500 mg/kg dose group also had a significantly elevated incidence of cryptorchidism compared with controls. Sertoli cell-only tubules were present in all cryptorchid testes in the 100 and 500 mg/kg dose groups, and 7 out of 11 cryptorchid testes in the 500 mg/kg group had one or more focal dysgenic areas. These effects were not observed in controls. | ⁶² |
| Dibutyl Phthalate | corn oil | groups of 10 female Wistar rats | 0 or 100 mg/kg bw/d | Pregnant rats received Dibutyl Phthalate via gavage on gestation days 12 to postnatal day 21. Five dams of the treatment and control groups were killed on gestation day 20 and testes from male fetuses from these dams were collected for histopathology. The male offspring from the remaining 5 dams of each group were weighed on postnatal day 1, and anogenital distance was measured on postnatal day 4. Remaining rats were killed at postnatal day 90. Blood was collected for hormone assays. The weights of the right testis, epididymis, vas deferens, ventral prostate, and seminal vesicle were recorded. The epididymides and testes were utilized for testicular and serum testosterone, assays, histopathological analysis, immunohistochemistry, proliferation index determination, androgen report analysis, and sperm analysis. | In the male offspring of the treated dams, Leydig cell clusters, the presence of multinucleated germinative cells, and an increased interstitial component were also observed. Anogenital distance was decreased, but this was statistically insignificant. Testosterone levels and reproductive organ weights were similar between the treated and control male offspring. Treatment with Dibutyl Phthalate did not markedly affect relative proportions of epithelial, stromal, or luminal compartments in the epididymis; sperm counts in the testis and epididymis; sperm transit time; or sperm morphology and motility in the adult male offspring. The androgen receptor and aquaporin 9 immunoreactivities and proliferation index were similar between the treatment group and control. | ⁶³ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-----------------------------|-------------------|--|--|--|--|---------------|
| Dibutyl Phthalate, 99% pure | corn oil and feed | groups of 3 pregnant Sprague-Dawley rats | 500 mg/kg in corn oil and 6000 ppm in feed | Dams received test material via gavage on gestation days 12 to 21; after weaning, 6 male offspring (F ₁ ; no more than 2 males/litter) were fed test material in diet until postnatal day 112; control group received corn oil and basal diet. F ₁ body weights and feed consumption assessed every other week until study end; on postnatal day 4, anogenital distance was measured and normalized (each animal's measurement was divided by the cube root of body weight), and all F ₁ animals were killed on postnatal day 112. At necropsy, testes were removed, weighted and underwent histological evaluation | On postnatal day 4, there were no statistically significant differences in the median body weights between groups of the male pups. The normalized anogenital distance of the Dibutyl Phthalate group pups was significantly decreased ($p < 0.05$) by 10% compared to the control animals. At postnatal day 112 Dibutyl Phthalate exposure had no effect on body weights when compared to the control. The absolute and relative testicular weights of the Dibutyl Phthalate group were lower than those of the control, but the decrease was not statistically significant. On postnatal day 112, testes from control animals showed intact tubules with normal spermatogenesis and mature sperm in the lumen. Dibutyl Phthalate-treated rats had mostly normal-appearing tubules although some Sertoli cell-only-type seminiferous tubules also appeared. Testicular impairment in Dibutyl Phthalate-treated animals resulted in a score of 1.5, significantly ($p < 0.05$) higher than the control animals. | ⁶⁴ |
| Dibutyl Phthalate | corn oil | groups of 3 pregnant Sprague-Dawley rats | 500 mg/kg/d | Dams received test material by oral gavage on gestation days 12 - 16 through 20. Dams were sacrificed on gestation days 17 - 21, and fetuses removed by Cesarean section. Additional pups from dams treated gestation days 12 - 21 were killed on postnatal days 1, 2, or 5. Male fetuses were identified by internal sex organ inspection. At gestation days 17 - 19, both testes were left in situ; at gestation days 20 - 21 and postnatal days 1 - 5, the left testis with epididymis was removed for analysis and the right left in situ. | Dibutyl Phthalate exposure produced abnormal morphology of fetal Sertoli cells, including retraction of cytoplasmic processes, convoluted apical membranes, and collapse of the vimentin cytoskeleton. Histopathological examination revealed a decreased number of seminiferous tubule cross-sections per testis, increased interstitial cell aggregates, and a significantly greater incidence of MNG compared to controls, with MNG containing higher numbers of nuclei. The MNG were not apoptotic (TUNEL-negative) but exhibited abnormal mitoses by postnatal day 5. Sertoli-gonocyte contacts were impaired, as shown by loss of cadherin-marked processes, and gonocytes clustered abnormally. While Sertoli cell morphology appeared restored after birth, MNG persisted postnatally and migrated to the basal lamina. Neural cell adhesion molecule expression remained intact in gonocytes, including MNG. Dibutyl Phthalate exposure also reduced fetal testicular testosterone levels, consistent with impaired Leydig cell function. | ⁶⁵ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------|----------|---|---|--|--|-----------|
| Dibutyl Phthalate | corn oil | groups of 8 or 9 pregnant Sprague-Dawley rats | 0 or 500 mg/kg bw every second day starting at gestation day 14.5 | Dams received test material on gestation days 14.5 to postnatal day 6 via oral gavage. Corn oil was also subcutaneously injected on gestation days 14.5 and 16.5. Pups were culled on gestation days 10, 24, and 90, and surviving pups were weaned on gestation day 21. Anogenital distance was measured on all pups on postnatal days 2 and 10. Trunk blood samples were collected on postnatal days 10, 24, and 90 and tail vein blood was collected on postnatal days 30, 35, 40, 45, and 55. Testosterone, insulin-like 3 (INSL3), and luteinizing hormone levels were measured. Testes were collected from the male offspring and underwent RNA analysis and Leydig cell counting. | Dams and offspring exhibited no adverse effects. Dams in the control and treatment group delivered exclusively live pups, although there was some subsequent perinatal mortality and testicular dysgenesis. At postnatal day 10, Dibutyl Phthalate-treated male pups had a reduced anogenital distance when compared to the controls. Male offspring from the Dibutyl Phthalate treatment group had relatively small testes at postnatal days 10 and 24, though larger testes at postnatal day 90. When compared to controls, there were no differences in Leydig cell counts or in testosterone or luteinizing hormone levels. Analysis of circulating INSL3 showed that up to postnatal day 40, values tended to be higher in the Dibutyl Phthalate-treated groups compared to controls. For the <i>Star</i> , <i>Hsd11b1</i> , and <i>Hsd17b3</i> gene products, there was an effect of Dibutyl Phthalate treatment at postnatal day 10, and a trend also at this time for the side-chain cleavage enzyme gene, <i>Cyp11a1</i> . A trend was also evident for Dibutyl Phthalate treatment for <i>Insl3</i> gene transcripts at postnatal day 24, but no effects were observed for any transcript at postnatal day 90. | 49 |
| Dibutyl Phthalate | corn oil | groups of 20 male Sprague-Dawley rats | 0, 100, 250, or 500 mg/kg/d | Rats received test material via gavage for 21 d. The rats were killed on day 24 and the testis and epididymis were removed for analysis. HPLC was used to detect Dibutyl Phthalate in the testis and expression of expression of H3 histone was measured with immunohistochemistry and immunofluorescence. Morphological changes were performed with optical microscope or transmission electron microscope (TEM). Semen quality was also assessed. | Residuary Dibutyl Phthalate in the testes was increased in a dose- dependent manner, and was significantly increased in the dose groups when compared to the control (p < 0.05). When compared to the control group, the microscopic structure of seminiferous tubule, spermatogonial cell, or spermatid cell in Dibutyl Phthalate groups was dramatically changed. In HE staining analysis, the separation between basement membrane of seminiferous tubule and Leydig was observed. In TEM analysis, karyopyknosis, mitochondrial pyknosis, enlarged vacuoles, karyorrhexis, mitochondrial decay and chromatin dissolution were found in spermatogonial cell or spermatid cell. The average optical density, positive negative area, and fluorescent intensity of H3 histone expression were significantly different between control group and each Dibutyl Phthalate group (p < 0.05). The quality of sperm in epididymis in the Dibutyl Phthalate groups was significantly reduced when comparing to the control (p < 0.05). | 66 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--|----------|---|---|--|---|-----------|
| Dibutyl Phthalate, 99.8% pure | corn oil | groups of 4 pregnant Sprague-Dawley rats | 0, 10, 30, 50, or 100 mg/kg/d | Pregnant rats received test material via gavage on gestation days 12 - 21. Offspring were weighed and sexed at birth and litters reduced to 4 males and 4 females per dam, with a mother not given further treatments after gestation day 21. Pups were removed from dams after weaning at postnatal day 21. All animals were weighed at birth, and at weeks 5, 7, 9, 14, and 17 of age. At each time point, 1 male from each dose group was killed, and tested were removed for analysis. | Male pups all dose group had similar body weights with no apparent differences in litter size or sex ratio compared to that of the control group. At postnatal weeks 5, 7, 9, 14, and 17, testicular weights for the 10, 30, and 50 mg/kg/d dose groups were similar to that for the control, but the 100 mg/kg/d dose group had a statistically significant decrease in testicular weight after 9 wk of age. The number of Leydig cells in the 100 mg/kg/d dose group was significantly higher than that of the control at 9, 14, and 17 wk; all other dose groups at all ages were similar to that of the control. Compared to the control, distinctive morphological changes were observed in the 100 mg/kg/d dose group. At 5 - 7 wk of age, the 100 mg/kg dose group had abundant smooth endoplasmic reticulum with a complicated arrangement of non-dilated cisternae and little stripped rough endoplasmic reticulum. The amount of smooth endoplasmic reticulum in Leydig cells was significantly decreased at 9 - 14 wk of age, and smooth endoplasmic reticulum were not observed at 17 wk of age. Mitochondrial swelling was not encountered, and the Golgi apparatus displayed no observable change. From 5 - 17 wk of age, the 10, 30, and 50 mg/kg/d dose groups had testosterone and luteinizing hormone levels similar to the control, while the 100 mg/kg/d dose group had significantly lower testosterone levels. The luteinizing hormone level of the 100 mg/kg/d dose group was significantly lower at 5 and 7 wk of age and significantly higher at 9 - 17 wk of age compared to that of the control. | 67 |
| Dibutyl Phthalate and Diethyl Phthalate, each 99% pure | corn oil | groups of female Sprague-Dawley rats; 4 rats/dose group for Dibutyl Phthalate and 4-5 rats/dose group for Diethyl Phthalate (control had 9) | 0, 33, 50, 100, 300, or 600 mg/kg/d for Dibutyl Phthalate; 0, 100, 300, 600, or 900 mg/kg/d for Diethyl Phthalate | Pregnant rats received test materials via gavage on gestation days 8-18 and killed after the last dosing; uteri from dams were removed and number of live and dead fetuses and resorptions were counted; male fetuses had testes removed for analysis of testosterone production | No adverse effects were observed in dams exposed to Dibutyl Phthalate. Dibutyl Phthalate did not significantly increase the rate of fetal death at any dose level. Total resorptions and fetal mortality were significantly increased ($p < 0.05$) in the 600 mg/kg/d Diethyl Phthalate group, but not in the 900 mg/kg/d dose group. Fetal testicular testosterone production was significantly reduced ($p < 0.001$) at doses of 300 mg/kg/d or higher for Dibutyl Phthalate; however, Diethyl Phthalate did not affect testosterone productions at any dose level | 68 |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|---|----------|--|---|---|--|---------------|
| A mixture of 5 phthalates, including Dibutyl Phthalate (99% pure) | corn oil | groups of 4 - 8 female Sprague-Dawley rats | 0, 5, 10, 20, 40, 60, 80, or 100%; 300 mg/kg/d Dibutyl Phthalate at the top dose; total mixture contained 1300 mg of 5 phthalates (dose ratio was 3:3:3:3:1, with 1 being dipentyl phthalate); | Pregnant rats were dosed via gavage on gestation days 8 – 18; rats were killed after the last dosing; uteri from dams were removed and number of live and dead fetuses and resorptions were counted; male fetuses had testes removed for analysis of testosterone production | The phthalate mixture containing Dibutyl Phthalate significantly reduced maternal body weight gain at 40% of the top dose (containing 120 mg/kg/d Dibutyl Phthalate) and greater; no other adverse effects were noted in the dams. A significant increase in fetal mortality was observed at 40% of top dose and above. Testosterone production was reduced in a dose-additive manner starting at 20% of top dose (containing 60 mg/kg/d Dibutyl Phthalate) and above. Phthalate-induced underdevelopment of the testes was observed at 100% of the top dose. | ⁶⁸ |
| Dibutyl Phthalate | corn oil | groups of 6 male albino rats | 0, 200, 400, or 600 mg/kg/d | Adult rats received Dibutyl Phthalate via gavage for 15 d. Blood samples were collected 24 h after the last dose. The rats were killed and the testes were excised and weighed. The cauda epididymides were used for analysis of sperm count and motility assay. One testis from each group was used for histopathological exam. Remaining testes were homogenized for enzyme and other biochemical assays. | Testicular weight was significantly decreased at each dose tested when compared to control group. Sperm count and motility were significantly decreased in a dose-dependent manner. Serum follicle-stimulating hormone, testosterone levels, and testicular lactate dehydrogenase activity were significantly decreased at all doses tested. Serum total antioxidant capacity and the activities of testicular antioxidant enzymes, such as superoxide dismutase, catalase, and glutathione reductase were significantly decreased at all doses tested when compared to control group. Testes of the 200 mg/kg dose group showed degeneration with absence of spermatogenesis in some seminiferous. In the 400 mg/kg dose group, degeneration was also observed with absence of spermatogenesis and sperms from most of the seminiferous tubules. Necrosis of some of the seminiferous tubules was observed in the 600 mg/kg dose group. | ⁵⁸ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--|----------|--|--|---|---|---------------|
| Dibutyl Phthalate | corn oil | groups of 10 male Sprague-Dawley rats | 250 mg/kg | Rats were exposed to test material, BaP (5 mg/kg), or both substances via gavage for 4, 8, or 12 wk; animals were observed for toxic effects during exposure; at the end of the dosing period, blood was collected and the testes and epididymides were removed and weighed and studied for histopathology and sperm evaluation | No clinical signs of toxicity were observed during the experiment; feed consumption levels were similar between treatment and vehicle groups; weekly body weights increased with exposure time, but there were no significant changes between groups. No adverse effects were observed in the relative weights of the testes and epididymides. Sperm count increased significantly ($p < 0.05$) in the Dibutyl Phthalate and combined treatment groups at 4 wk, but were at control levels at 8 and 12 wk. Abnormal sperm ratio following an ascending trend in Dibutyl Phthalate treated rats, but this was not significantly different from the control. A significant increase ($p < 0.05$) was observed in serum T level from Dibutyl Phthalate exposure for 12 wk; combined exposure produced lower serum T levels than the Dibutyl Phthalate exposure. Vacuolization of Sertoli cells was observed in Dibutyl Phthalate-exposed rats after 12 wk. Combined exposure induced significant decreases in SOD and GSH-Px activities in the testis at 8 wk compared with control ($p < 0.05$). CAT activities significantly decreased following the Dibutyl Phthalate exposure after 8 wk ($p < 0.05$). Combined treatment resulted in a significant elevation in CAT activities at 12 wk control ($p < 0.05$). | ⁶⁰ |
| Dibutyl Phthalate, purity not reported | corn oil | groups of 5 male Sprague-Dawley rats | 0 or 500 mg/kg bw/d | Rats received treatment by gavage once daily for 3 wk. | Dibutyl Phthalate exposure caused a significant reduction in testes weight ($p < 0.05$), a decrease in the proportion of the germ cell layer in the cross-section of the seminiferous tubules, and a decrease of the sperm density in the lumen of epididymis tissues ($p < 0.05$). Dibutyl Phthalate exposure resulted in significantly increased MDA, Fe^{2+} , and GSSG levels and decreased GSH in testes homogenates ($p < 0.05$), and significant serum testosterone decrease ($p < 0.05$). Immunohistochemistry and q-PCR of testes tissue showed that the expression of PRDX6 and SP1 was upregulated. | ⁵² |
| Dibutyl Phthalate; > 99% pure | feed | 50 male and 50 female B6C3F1/N mice | 0, 1000, 3000, or 10,000 ppm, equivalent to 105 - 112, 329 - 347, and 1306 - 1393 mg/kg bw/d, respectively | 2-yr carcinogenicity study (see the Carcinogenicity Studies section for full details); body weights and feed consumption measured, clinical observations made through study period, and complete necropsies and microscopic examinations performed on all mice at study end | No gross lesions observed in male reproductive tract, but significantly ($p < 0.05$) increased incidences of germinal epithelium degeneration in the testes and exfoliated germ cells in the epididymal duct were observed microscopically, mainly in the 10,000 ppm dose group | ⁴⁴ |
| Diethyl Phthalate | feed | groups of male Wistar rats; no further details available | 0 or 2000 mg/kg/d (2%) | Rats received test material in the diet for 7 d. Testosterone levels were measured in testes and serum, and testes were examined histopathologically; no further details available | Lower levels of testosterone were measured in the testes and serum of the treated animals, but no testicular damage was observed by microscopic examination | ⁶ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|--|--|--|--|---|--|---------------|
| Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate, monobutyl phthalate, monomethyl phthalate, monomethyl phthalate, and phthalic acid | corn oil | groups of 6 male Sprague-Dawley rats | 250 mg/kg bw/d for monoesters and phthalic acid; 500 mg/kg bw/d for diesters | Rats received test materials orally via gavage for 4 wk; control group received corn oil only; (non-DART methods described in the repeated-dose table above); testes and epididymis were weighted at necropsy; and sperm was collected for sperm count analysis and motility analysis | Testes weights were significantly reduced in the Dibutyl Phthalate group compared to the control. Dibutyl Phthalate, monobutyl phthalate, and monoethyl phthalate significantly lowered sperm counts and sperm motility of epididymal sperm, detected by a change in the sperm motion parameters | ³⁶ |
| OTHER ROUTES | | | | | | |
| Dibutyl Phthalate | 25% (v/v) tetraglycol, 25% (v/v) polyethylene glycol 300, 25% (v/v) kolliphor EL, 15% (v/v) ethanol, and 10% (v/v) 1,2-propanediol | groups of 4 male C57BL/6J mice | 2.5 mg/kg/d | Mice were exposed to control, Dibutyl Phthalate, diethylhexyl phthalate (2.5 mg/kg/d), or a mixture for 40 d via surgically implanted osmotic pumps; caudal epididymal spermatozoa were extracted and analyzed for motility; sperm phosphorylation of protein kinase A substrates and tyrosine phosphorylation were also analyzed; in vitro fertilization was used to evaluate sperm fertilizing capacity | No significant difference in sperm motility and fertilization potential was observed. Abnormal sperm morphology was observed in all phthalate exposures, particularly in the mixture group. Significant differences in sperm concentration were observed between control and exposed groups. Protein phosphorylation of protein kinase A substrates was decreased in the in mixture group, but no significant changes in protein tyrosine phosphorylation were observed in any of the treatment groups. Assessment of the reproductive functionality did not reveal significant effects on in vitro fertilization and early embryo development rates, but showed wide variability in the mixture group. | ⁶⁹ |
| Dibutyl Phthalate | corn oil | groups of 12 neonatal male Sprague-Dawley rats | 0, 5, 10, or 20 mg/animal | Dibutyl Phthalate was injected subcutaneously in rats from postnatal day 5 to 14. Animals were observed for clinical signs of toxicity and body weights were measured. Animals were killed on postnatal day 31 or 42, and testes, epididymis, seminal vesicles, ventral prostate, LABC, and Cowper's glands were weighed. The expressions of androgen receptor, estrogen receptors, and steroidogenic factor-1 (SF-1) were also examined in the testes. | No clinical signs of toxicity were observed to any rats in the treatment groups. In the 20 mg dose group, significantly reduced body weight was observed on postnatal days 29 and 31, but changes recovered to control levels by postnatal day 42. No alterations in testicular descent were observed in any treated rats; however, significantly reduced testes, seminal vesicles, LABC, and Cowper's glands weights were observed in the 20 mg dose group when compared to the controls. No significant difference was observed in serum testosterone levels with the treated groups and the controls. In the 20 mg dose group, a marked decreased in seminiferous tubule and Leydig cell hyperplasia were observed. Epithelial cell height in the proximal vas deferens was reduced in the 20 mg dose group and epithelial cell proliferation was observed. At postnatal day 31, the 20 mg dose group had significantly reduced expression of androgen receptor and significantly increased estrogen and SF-1 expression. | ⁷⁰ |

Table 7. Developmental and reproductive toxicity studies

| Test Article | Vehicle | Animals/Group | Dose/Concentration | Procedure | Results | Reference |
|-------------------------------|---------|---|-----------------------|---|---|---------------|
| OTHER | | | | | | |
| Dibutyl Phthalate, > 99% pure | DMSO | <i>Danio rerio</i> ; 6 fish per tank; 3 replicates per dose | 11, 113, or 1133 µg/l | Study to investigate the effects of Dibutyl Phthalate alone or with diisobutyl phthalate (10, 103, or 1038 µg/l) on male reproduction and to explore potential toxicological mechanisms. Adult zebrafish were exposed to Dibutyl Phthalate, diisobutyl phthalate, or a mix for 30 d and the effects on plasma hormone secretion, testis histology, and transcriptomics were examined. Chemical exposure was performed in accordance with OECD TG 229. Test included control and solvent control (0.05% DMSO) groups | Highest concentrations of the exposure to the mixture caused greater imbalance testosterone to estradiol ratio and more severe structural damage to testis than exposure to just Dibutyl Phthalate. These effects were consistent with the testis transcriptome analysis for which 4570 genes were differentially expressed in the mixture exposure, while 2795 genes were differentially expressed in Dibutyl Phthalate. KEGG pathway analysis showed that both single and combined exposure of Dibutyl Phthalate and diisobutyl phthalate could affect cytokine-cytokine receptor interaction. The difference was that combined exposure could also affect steroid hormone synthesis, extracellular matrix receptor interaction, retinol metabolism, and PPAR signaling pathways. | ⁷² |

Table 8. Genotoxicity studies

| Test Article | Vehicle | Concentration/Dose | Test System | Protocol | Results | Reference |
|--|----------------------|------------------------------------|---|--|---|-----------|
| IN VITRO | | | | | | |
| Dibutyl Phthalate | not reported | 10 - 2000 µg/plate | <i>S. typhimurium</i> strains TA100 and TA1535 | Ames test; with and without metabolic activation | Not genotoxic | 5 |
| Diethyl Phthalate | DMSO | up to 5000 µg/plate | <i>S. typhimurium</i> strains TA98, TA100, TA1535, TA1537 and <i>E.coli</i> strain WP2 uvrA | Ames test in accordance with OECD TG 471, with and without metabolic activation | Not genotoxic | 6 |
| Dibutyl Phthalate, Diethyl Phthalate, Dimethyl Phthalate | DMSO | 100-2000 µg/plate | <i>S. typhimurium</i> strains TA98, TA100, TA1535, TA1537, TA1538, TA2637 | Ames test, with and without metabolic activation | Dibutyl Phthalate and Diethyl Phthalate were mildly genotoxic in strains TA100 and TA1535 without metabolic activation in a dose-dependent manner. Dimethyl Phthalate was mildly genotoxic in TA1535, and after normalization for cytotoxicity, was mildly genotoxic in TA100 without metabolic activation. Genotoxicity was not observed with metabolic activation in these strains with these test materials. | 73 |
| Dibutyl Phthalate | not reported | 0, 30, 50, or 80 µM | cultured bovine lymphocytes | Micronucleus test; cells incubated with test material for 24 h | Genotoxic; a significant ($p < 0.001$) effect of induction of micronuclei and binuclei observed at all 3 test concentrations compared to the negative control group | 74 |
| Dibutyl Phthalate | DMSO | 10, 15, or 30 µM | CHO cells | Micronucleus test; cells were treated with test material for 24 h | Dibutyl Phthalate had a non-linear induction of micronuclei percentage at all 3 concentrations, with 15 µM causing the highest percentage of micronuclei (4.90%, $p < 0.001$) compared with 10 µM (3.36% $p < 0.05$) and 30 µM (3.66%, $p < 0.01$). | 75 |
| Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate | methanol and ethanol | 0.05 or 0.2 ppm for each phthalate | TK6 human lymphoblast cells | MicroFlow® micronucleus assay; the assay also evaluated the genotoxicity of 42 perfumes that contained 5 phthalates including Dibutyl Phthalate (mean concentration 0.032 ppm), Diethyl Phthalate (mean concentration 1673 ppm), and Dimethyl Phthalate (mean concentration 32.25 ppm) | The percent frequency of micronuclei was more than twofold higher in the cells treated with the perfumes than those treated with the vehicle controls, but the differences were not significant ($p > 0.05$). Percent frequency of micronuclei by each of the 5 phthalates regardless of their concentration (0.05 or 0.2 ppm), did not differ significantly from those induced by the perfumes. No significant differences in percent frequency of micronuclei between the low and high concentrations of Diethyl Phthalate and Dimethyl Phthalate; the percent frequency of micronuclei for Dibutyl Phthalate was significantly lower at 0.2 than 0.05 ppm ($p < 0.001$). | 15 |
| Dibutyl Phthalate | DMSO | 10, 15, or 30 µM | CHO cells | Chromosome aberration test; cells were treated with test material for 24 h | Dibutyl Phthalate caused a higher occurrence of chromosomal aberrations at all tested concentration when compared to the vehicle control. Aberrations were predominantly with gaps in both chromatids and chromosomes. Effects occurred in a non-linear fashion with highest clastogenic effects observed at 15 µM | 75 |
| Diethyl Phthalate | DMSO | up to 1780 µg/ml | human lymphocytes | Chromosome aberration test in accordance with OECD TG 473; with and without metabolic activation | Not genotoxic | 6 |

Table 8. Genotoxicity studies

| Test Article | Vehicle | Concentration/Dose | Test System | Protocol | Results | Reference |
|--|----------------------|------------------------------------|---|---|---|---------------|
| Diethyl Phthalate | DMSO | up to 771 µg/ml | mouse L5178Y TK +/- lymphoma cells | Gene mutation assay in accordance with OECD TG 476; with and without metabolic activation | Not genotoxic | ⁶ |
| Dibutyl Phthalate | DMSO | 0 or 354 µmol/ml | human mucosal cells or human peripheral lymphocytes | Comet assay; cells were incubated with test material for 60 min. N-methyl-N-nitro-N-nitrosoguanidine served as positive control | Genotoxicity was observed in both cell types. Positive control yielded expected results. | ⁷⁶ |
| Dibutyl Phthalate | not reported | 0, 30, 50, or 80 µM | cultured bovine lymphocytes | Alkaline comet assay; cells incubated with test material for 24 h | Genotoxic; dose-dependent DNA damage observed | ⁷⁴ |
| Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate | methanol and ethanol | 0.05 or 0.2 ppm for each phthalate | TK6 human lymphoblast cells | Alkaline comet assay; the assay also evaluated the genotoxicity of 42 perfumes that contained 5 phthalates including Dibutyl Phthalate (mean concentration 0.032 ppm), Diethyl Phthalate (mean concentration 1673 ppm), and Dimethyl Phthalate (mean concentration 32.25 ppm) | The assay indicated that most of the perfumes caused DNA strand breaks. The average tail moment in the cells treated with perfume (55.05) was significantly higher than the tail moments for the negative and vehicle controls ($p < 0.01$), but not significantly different from the tail moments for each of the 5 phthalates ($p > 0.05$), regardless of the concentration. The extent of DNA damage by Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate was significantly ($p < 0.001$, except for Dimethyl at $p < 0.01$) higher (approximately twofold) at 0.2 than 0.05 ppm. DNA damage did not differ significantly between cells treated with either the low or high phthalate concentrations and cells treated with perfumes | ¹⁵ |

Table 9. Dermal irritation and sensitization studies

| Test Article | Vehicle | Concentration/Dose | Test Population/System | Protocol | Results | Reference |
|-------------------|-----------|--------------------|--|---|---|--------------|
| IRRITATION | | | | | | |
| ANIMAL | | | | | | |
| Dibutyl Phthalate | undiluted | undiluted | 2 male and 1 female Vienna White rabbits | Dermal irritation study performed in accordance with OECD TG 404; rabbits received 0.5 ml of test material on shaved skin on the back or flank with an area of 2.5 cm ² ; test sites were semi-occluded, washed after 4 h with lutrol and lutrol/water 1:1, and observed for 72 h | Not irritating; very slight (grade 1) erythema was observed in 2 anils after 4 and 24 h. Reactions were completely reversible after 48 h. | ⁵ |
| Diethyl Phthalate | undiluted | undiluted | 3 albino rabbits; sex not reported | Dermal irritation study; rabbits received 0.5 ml of test material on shaved and intact or abraded skin; test sites were 2 cm ² and occluded for 24 h; observation period was 72 h and evaluations for skin reactions were periods at the end of the 24 h exposure and again 48 h later | Not irritating; irritation scores were 0 for all rabbits | ⁶ |

Table 9. Dermal irritation and sensitization studies

| Test Article | Vehicle | Concentration/Dose | Test Population/System | Protocol | Results | Reference |
|--|---|---|--|---|---|----------------|
| SENSITIZATION | | | | | | |
| ANIMAL | | | | | | |
| Dibutyl Phthalate | olive oil DAB 9 | Intradermal induction: 5% Topical induction: 75% Challenge: 50% | 20 female Pirbright-Hartley guinea pigs | GPMT performed in accordance with OCED TG 406; induction with test material intradermally with Freund's adjuvant; topical induction 1 wk later on test site 2 cm x 4 cm that was occluded; after 2 wk rest period, challenge was performed with two 24-h patches (2 cm ²) applied a week apart | Not sensitizing; slight erythema was observed in 2 animals that received Dibutyl Phthalate after the first challenge patch 48 h after application. No reactions to the test material were observed in the second challenge. One reaction was observed in a control animal during the first challenge only. | ⁵ |
| Diethyl Phthalate | acetone, water, and ethanol; no further details available | up to 100% | Groups of 6 to 8 male and female Himalayan white spotted guinea pigs | Open epicutaneous test, Draize test, maximization test, and Freund's complete adjuvant test; no further details available | Not sensitizing; no further details available | ⁶ |
| Diethyl Phthalate | acetone: olive oil (4:1 v/v) | 25, 50, or 100% v/v | groups of 4 female CBA/Ca mice | LLNA in accordance with OECD TG 429 | Not sensitizing at any concentration tested; SI were 1.0, 1.3, and 1.5 for 25, 50, and 100%, respectively | ⁶ |
| Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate | acetone | 1:1 v/v with acetone; no further details provided | groups of female CD-1 mice, no further details provided | Ear swelling test to determine the adjuvant effect of phthalate esters on contact hypersensitivity; mice were epicutaneously sensitized with FITC in acetone with a phthalate ester; sensitization was evaluated after challenge with FITC; draining lymph node cells obtained 24 h after sensitization were examined for FITC fluorescence by flow cytometry; FITC-positive cells were characterized with anti-CD11c and anti-CD11b by 3-color flow cytometry; left ear of mice was treated with 20 µl of acetone/Dibutyl Phthalate alone as control | Mice sensitized with FITC in acetone containing Dibutyl Phthalate had strong enhancement of ear-swelling response; Diethyl Phthalate and Dimethyl Phthalate were less effective but produced some enhancement. With sensitization in the presence of Dibutyl Phthalate, the number of FITC-positive dendritic cells was increased in draining lymph nodes; no significant increase in FITC-positive cell number in the draining lymph nodes was observed with Diethyl Phthalate or Dimethyl Phthalate | ¹¹⁴ |

Table 10. Concentration of Use (2025) and Exposure by FDA Product Category – Diethyl Phthalate

| Product Category/Type of cosmetics exposure | Daily Exposure by Product Category (mg/d) | Retention factor | Maximum Concentration of Use | Daily Exposure Based on the Highest Use Concentration (mg/d) | Note |
|---|--|-------------------------|-------------------------------------|---|---|
| Skin cleansing (cold creams, cleansing lotions, liquids and pads) | 4810 | 0.1 | 0.1% | 0.481 | Estimated exposure to facial cleanser (lathering and non-lathering) at the 90 th percentile ¹⁴³ |
| Face and neck products (not spray) Leave-on | 3500 | 1 | 0.15% | 5.25 | Exposure amount of face cream/lotion applied ¹⁴⁴ |

REFERENCES

1. Nikitakis J, Kowcz A. 2024. Web-Based International Cosmetic Ingredient Dictionary and Handbook. <https://incipedia.personalcarecouncil.org/winci/>.
2. Elder RL (ed.). Final Report on the Safety Assessment of Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate. *J Am Coll Toxicol*. 1985;4(3):267–303.
3. Andersen FA (ed.). Annual Review of Cosmetic Ingredient Safety Assessment - 2002/2003. Dibutyl Phthalate, Diethyl Phthalate, and Dimethyl Phthalate. *Int J Toxicol*. 2005;24:34–42.
4. Andersen FA, Heldreth B. Dibutyl, Dimethyl, and Diethyl Phthalate and Butyl Benzyl Phthalate. *Int J Toxicol*. 2017;36:44S–45S.
5. European Chemicals Agency. 2023. Dibutyl Phthalate. https://chem.echa.europa.eu/100.001.416/dossier-view/5ac5a3cf-7792-48ce-ab71-72bad67bae5f/529f0536-1c0a-485d-b064-246b6f152cfd_529f0536-1c0a-485d-b064-246b6f152cfd. Date Accessed: May 20, 2025.
6. European Chemicals Agency. 2014. Diethyl Phthalate. https://chem.echa.europa.eu/100.001.409/dossier-view/d411946e-b58c-457c-8ae4-4e3603c33d51/d51e0e4b-f9fe-4869-9a0e-1417914d8167_d51e0e4b-f9fe-4869-9a0e-1417914d8167. Date Accessed: May 21, 2025.
7. European Chemicals Agency. 2023. Dimethyl Phthalate. https://chem.echa.europa.eu/100.004.557/dossier-view/b6f5f758-c41e-4a64-accb-6055e04d3a4f/d85e8588-2c7d-40de-b467-9b8bbd00207b_d85e8588-2c7d-40de-b467-9b8bbd00207b. Date Accessed: May 27, 2025.
8. Cosmetic Ingredient Review (CIR) Expert Panel. 2002. Dibutyl Phthalate Group. Washington, D.C. [Unpublished draft report for the re-review of Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate presented at the November 18–19, 2002 CIR Expert Panel meeting. Washington, D.C.].
9. Huang L, Zhu X, Zhou S, Cheng Z, et al. Phthalic acid esters: Natural sources and biological activities. *Toxins (Basel)*. 2021;13(7):495.
10. U.S. Food and Drug Administration. Federal Food, Drug, and Cosmetic Act Section 612 Title 21.
11. U.S. Food and Drug Administration Office of the Chief Scientist. 2024. Registration and Listing Data - Frequency of Use of Cosmetic Products. College Park, MD [Obtained under the Freedom of Information Act from the Division of Freedom of Information; requested as "Frequency of Use Data" July 17, 2024; received July 30, 2024].
12. Personal Care Products Council. 2025. Concentration of Use by FDA Product Category (new MoCRA categories): Dibutyl Phthalate, Diethyl Phthalate and Dimethyl Phthalate. [Unpublished data submitted by the Personal Care Products Council on March 27, 2025.].
13. U.S. Food and Drug Administration Center for Food Safety & Applied Nutrition. 2023. Voluntary Cosmetic Registration Program - Frequency of Use of Cosmetic Ingredients. College Park, MD. [Obtained under the Freedom of Information Act from CFSAN; requested as "Frequency of Use Data" January 4, 2023; received February 2, 2023.].
14. Lehner K, Santarelli F, Vasold R, Konig B, Landthaler M, Baumler W. Black tattoo inks are a source of problematic substances such as dibutyl phthalate. *Contact Derm*. 2011;65(4):231–238.
15. Al-Saleh I, AL-Rajudi T, Al-Qudaihi G, Manogaran P. Evaluating the potential genotoxicity of phthalates esters (PAEs) in perfumes using in vitro assays. *Environ Sci Pollut Res Int*. 2017;24(30):23903–23914.
16. Al-Saleh I, Elkhatib R. Screening of phthalate esters in 47 branded perfumes. *Environ Sci Pollut Res Int*. 2016;23(1):455–468.

17. Api AM. Toxicological profile of diethyl phthalate: A vehicle for fragrance and cosmetic ingredients. *Food Chem Toxicol.* 2001;39(2):97–108.
18. EUR-Lex. 2009. Regulation (EC) No 1223/2009 of the European Parliament and of the Council of 30 November 2009 on cosmetic products (recast). (Text with EEA relevance). <https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX%3A02009R1223-20231201&qid=1704209440026>.
19. Scientific Committee on Cosmetic Products and Non-Food Products. 2004. The Scientific Committee on Cosmetic Products and Non-Food Products Intended for Consumers Opinion Concerning Dibutylphthalate. http://ec.europa.eu/health/ph_risk/committees/sccp/documents/out287_en.pdf.
20. Scientific Committee on Cosmetic Products and Non-Food Products. 2002. Opinion of the Scientific Committee on Cosmetic Products and Non-Food Products Intended for Consumers Concerning Diethyl Phthalate. http://ec.europa.eu/health/ph_risk/committees/sccp/documents/out168_en.pdf.
21. Scientific Committee on Cosmetic Products and Non-Food Products. 2003. Scientific Committee on Cosmetic Products and Non-Food Products Intended for Consumers Opinion Concerning Diethyl Phthalate. http://ec.europa.eu/health/ph_risk/committees/sccp/documents/out246_en.pdf.
22. Scientific Committee on Consumer Products. 2007. Opinion on Phthalates in Cosmetic Products. https://ec.europa.eu/health/ph_risk/committees/04_sccp/docs/sccp_o_106.pdf.
23. da Silva Costa R, Sainara Maia Fernandes T, de Sousa Almeida E, et al. Potential risk of BPA and phthalates in commercial water bottles: A minireview. *J Water Health.* 2021;19(3):411–435.
24. Gao CJ, Kannan K. Phthalates, bisphenols, parabens, and triclocarban in feminine hygiene products from the United States and their implications for human exposure. *Environ Int.* 2020;136:105465.
25. U.S. Food and Drug Administration Center for Drug Evaluation and Research. 2012. Guidance for Industry Limiting the Use of Certain Phthalates as Excipients in CDER-Regulated Products. <http://www.fda.gov/downloads/Drugs/GuidanceComplianceRegulatoryInformation/Guidances/UCM294086.pdf>. Date Accessed: April 30, 2025.
26. European Food Safety Authority, Volk K, Castle L. 2019. Technical report of the public consultation on the 'Draft update of the risk assessment of di-butylphthalate (DBP), butyl-benzyl-phthalate (BBP), bis(2-ethylhexyl)phthalate (DEHP), di-isononylphthalate (DINP) and di-isodecylphthalate (DIDP) for use in food contact materials'. <https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1747>. Date Accessed: November 4, 2025.
27. Doan K, Bronaugh RL, Yourick JJ. In vivo and in vitro skin absorption of lipophilic compounds, dibutyl phthalate, farnesol and geraniol in the hairless guinea pig. *Food Chem Toxicol.* 2010;48(1):18–23.
28. Jung D, Yoon C, Um M, Hwang H, Baek J, Choi J. Characteristics of percutaneous absorption for three kinds of phthalate. *J Environ Health Sci.* 2013;39(4):360–368.
29. Jauregui EJ, Lock J, Rasmussen L, Craig ZR. Mono-n-butyl phthalate distributes to the mouse ovary and liver and alters the expression of phthalate-metabolizing enzymes in both tissues. *Toxicol Sci.* 2021;183(1):117–127.
30. Janjua NR, Frederiksen H, Skakkebaek NE, Wulf HC, Andersson A. Urinary excretion of phthalates and paraben after repeated whole-body topical application in humans. *Int J Androl.* 2008;31(2):118–130.
31. Lorber M, Weschler CJ, Morrison G, Beko G, et al. Linking a dermal permeation and an inhalation model to a simple pharmacokinetic model to study airborne exposure to di(n-butyl) phthalate. *J Expo Sci Environ Epidemiol.* 2017;27(6):601–609.
32. Koch HM, Christensen LKY, Harth V, Lorber M, Bruning T. Di-n-butyl phthalate (DnBP) and diisobutyl phthalate (DiBP) metabolism in a human volunteer after single oral doses. *Arch Toxicol.* 2012;86(12):1829–1839.

33. van Ravenzwaay B, Coelho-Palermo Cunha G, Strauss V, Wiemer J, et al. The individual and combined metabolite profiles (metabolomics) of dibutylphthalate and di(2-ethylhexyl)phthalate following a 28-day dietary exposure in rats. *Toxicol Lett.* 2010;198(2):159–170.
34. Zhou S, Han M, Ren Y, Yang X, et al. Dibutyl phthalate aggravated asthma-like symptoms through oxidative stress and increasing calcitonin gene-related peptide release. *Ecotoxicol Environ Saf.* 2020;199:110740.
35. Wang X, Han B, Wu P, Li S, et al. Dibutyl phthalate induces allergic airway inflammation in rats via inhibition of the Nrf2/TSLP/JAK1 pathway. *Environ Pollut.* 2020;267:115564.
36. Kwack SJ, Kim KB, Kim HS, Lee BM. Comparative toxicological evaluation of phthalate diesters and metabolites in Sprague-Dawley male rats for risk assessment. *Toxicol Environ Health A.* 2009;72(21-22):1446–1454.
37. Gardner ST, Wood AT, Lester R, Onkst PE, et al. Assessing differences in toxicity and teratogenicity of three phthalates, Diethyl phthalate, Di-n-propyl phthalate, and Di-n-butyl phthalate, using *Xenopus laevis* embryos. *J Toxicol Environ Health A.* 2016;79(2):71–82.
38. Arancio AL, Cole KD, Dominguez AR, Cohenour ER, et al. Bisphenol A, Bisphenol AF, Di-n-butyl Phthalate, and 17 β -estradiol have shared and unique dose-dependent effects on early embryo cleavage divisions and development in *Xenopus laevis*. *Reprod Toxicol.* 2019;84:65–74.
39. Xu Y, Jang J, Gye MC. The *Xenopus laevis* teratogenesis assay for developmental toxicity of phthalate plasticizers and alternatives. *Environ Pollut.* 2022;300:118985.
40. Mu X, Huang Y, Li J, Yang K, et al. New insights into the mechanism of phthalate-induced developmental effects. *Environ Pollut.* 2018;241:674–683.
41. Pu S, Hamid N, Ren Y, Pei D. Effects of phthalate acid esters on zebrafish larvae: Development and skeletal morphogenesis. *Chemosphere.* 2020;246:125808.
42. Dobrzynska MM, Tyrkiel EJ, Gajowik A. Three generation study of reproductive and developmental toxicity following exposure of pubescent F0 male mice to di-n-butyl phthalate. *Mutagenesis.* 2017;32(4):445–454.
43. Wolfe G. W., Patel R. V. 2002. Dibutyl Phthalate: Multigenerational Reproductive Assessment by Continuous Breeding When Administered to Sprague-Dawley Rats in the Diet. TherImmune No. 7244-201. NTP-RACB-97003. [Available from National Technical Information Service PB2003-100084].
44. National Toxicology Program. 2021. NTP Technical Report on the Toxicology and Carcinogenesis Studies of Di-n-butyl Phthalate (CASRN 84-74-2) Administered in Feed to Sprague Dawley (Hsd:Sprague Dawley® SD®) Rats and B6C3F1/N Mice. https://ntp.niehs.nih.gov/ntp/htdocs/lt_rpts/tr600_508.pdf.
45. Fujii S, Yabe K, Furukawa M, Hirata M, Kiguchi M, Ikka T. A two-generation reproductive toxicity study of diethyl phthalate (DEP) in rats. *J Toxicol Sci.* 2005;30:97–116.
46. Rasmussen LM, Sen N, Vera JC, Liu X, Craig ZR. Effects of in vitro exposure to dibutyl phthalate, mono-butyl phthalate, and acetyl tributyl citrate on ovarian antral follicle growth and viability. *Biol Reprod.* 2017;96(5):1105–1117.
47. Jauregui EJ, McSwain M, Liu X, Burns K, Craig ZR. Human relevant exposure to di-n-butyl phthalate tampers with the ovarian insulin-like growth factor 1 system and disrupts folliculogenesis in young adult mice. *Toxicol Sci.* 2023;195(1):42–52.
48. Liu X, Craig ZR. Environmentally relevant exposure to dibutyl phthalate disrupts DNA damage repair gene expression in the mouse ovary. *Biol Reprod.* 2019;101(4):854–867.
49. Ivell R, Heng K, Nicholson H, Anand-Ivell R. Brief maternal exposure of rats to the xenobiotics dibutyl phthalate or diethylstilbestrol alters adult-type Leydig cell development in male offspring. *Asian J Androl.* 2013;15(2):261–268.

50. Ahmad R, Verma Y, Gautam AK, Kumar S. Assessment of estrogenic potential of di-n-butyl phthalate and butyl benzyl phthalate in vivo. *Toxicol Ind Health*. 2015;31(12):1296–1303.
51. Zhou C, Gao L, Flaws JA. Exposure to an environmentally relevant phthalate mixture causes transgenerational effects on female reproduction in mice. *Endocrinology*. 2017;158(6):1739–1754.
52. Bu H, Wang B, Wu Y, Li P, et al. Curcumin strengthens a spontaneous self-protective mechanism-SP1/PRDX6 pathway, against di-n-butyl phthalate-induced testicular ferroptosis damage. *Environ Sci Pollut Res Int*. 2023;30(58):122165–122181.
53. Kaminska A, Pardyak L, Marek S, et al. Bisphenol A and dibutyl phthalate affect the expression of juxtacrine signaling factors in rat testis. *Chemosphere*. 2018;199:182–190.
54. Zhang J, jin S, Zhao J, Li H. Effect of dibutyl phthalate on expression of connexin 43 and testosterone production of leydig cells in adult rats. *Environ Toxicol Pharmacol*. 2016;47:131–135.
55. Pant N, Pant AB, Shukla M, Mathur N, et al. Environmental and experimental exposure of phthalate esters: The toxicological consequence on human sperm. *Hum Exp Toxicol*. 2010;30(6):507–514.
56. Xie F, Chen X, Weng S, et al. Effects of two environmental endocrine disruptors di-n-butyl phthalate (DBP) and mono-n-butyl phthalate (MBP) on human sperm functions in vitro. *Reprod Toxicol*. 2019;83:1–7.
57. Zhang Y, Jiang X, Chen B. Reproductive and developmental toxicity in F1 Sprague-Dawley male rats exposed to di-n-butyl phthalate in utero and during lactation and determination of its NOAEL. *Reprod Toxicol*. 2004;18(5):669–676.
58. Aly H, Hassan MH, El-Beshbishy HA, Alahdal AM, Osman A. Dibutyl phthalate induces oxidative stress and impairs spermatogenesis in adult rat. *Toxicol Ind Health*. 2015;32(8):1467–1477.
59. Fisher JS, Macpherson S, Marchetti N, Sharpe RM. Human 'testicular dysgenesis syndrome': A possible model using in-utero exposure of the rat to dibutyl phthalate. *Hum Reprod*. 2003;18(7):1383–1394.
60. Chen X, An H, Ao L, et al. The combined toxicity of dibutyl phthalate and benzo(a)pyrene on the reproductive system of male Sprague Dawley rats in vivo. *J Hazard Mater*. 2011;186(1):835–841.
61. Barlow NJ, Foster P. Pathogenesis of male reproductive tract lesions from gestation through adulthood following in utero exposure to di(n-butyl) phthalate. *Toxicol Pathol*. 2003;31(4):397–410.
62. Mahood IK, Scott HM, Brown R, Hallmark N, et al. In utero exposure to di(n-butyl) phthalate and testicular dysgenesis: Comparison of fetal and adult end points and their dose sensitivity. *Environ Health Perspect*. 2007;115(Suppl 1):55–61.
63. Scarano WR, Toledo FC, Guerra MT, Pinheiro P, et al. Functional and morphological reproductive aspects in male rats exposed to di-n-butyl phthalate (DBP) in utero and during lactation. *J Toxicol Environ Health A*. 2010;73(13-14):972–984.
64. de Souza NP, Ferragut Cardoso AP, Gomide L, Lima T, et al. Experimental cryptorchidism enhances testicular susceptibility to dibutyl phthalate or acrylamide in Sprague-Dawley rats. *Hum Exp Toxicol*. 2019;38(8).
65. Kleymenova E, Swanson C, Boekelheide K, Gaido KW. Exposure in utero to di(n-butyl) phthalate alters the vimentin cytoskeleton of fetal rat Sertoli cells and disrupts Sertoli cell-gonocyte contact. *Biol Reprod*. 2005;73(3):482–490.
66. Yin L, Yan L, He B, Fang Y, et al. The toxic effects of a plasticizer, dibutyl phthalate, on rat testis. *Int J Clin Exp Pathol*. 2016;9(11):11246–11253.
67. Shirai M, Wakui S, Wempe MF, Mutou T, et al. Male Sprague-Dawley rats exposed to in utero di(n-butyl) phthalate: Dose dependent and age-related morphological changes in Leydig cell smooth endoplasmic reticulum. *Toxicol Pathol*. 2013;41(7):984–991.

68. Howdeshell KL, Wilson VS, Furr J, Lambright CR, et al. A mixture of five phthalate esters inhibits fetal testicular testosterone production in the Sprague-Dawley rat in a cumulative, dose-additive manner. *Toxicol Sci.* 2008;105(1):153–165.
69. Mohanty G, Tourzani DA, Gerasi MG, et al. Effects of preconception exposure to phthalates on mouse sperm capacitation parameters. *Andrology.* 2023;11(7):1484–1494.
70. Kim HS, Kim TS, Shin J, Moon HJ, Kang IH, et al. Neonatal exposure to di(n-butyl) phthalate (DBP) alters male reproductive-tract development. *J Toxicol Environ Health A.* 2004;67(23-24):2045–2060.
71. Abdul-Ghani S, Yanai J, Abdul-Ghani R, Pinkas A, Abdeen Z. The teratogenicity and behavioral teratogenicity of di(2-ethylhexyl) phthalate (DEHP) and di-butyl phthalate (DBP) in a chick model. *Neurotoxicol Teratol.* 2012;34(1):56–62.
72. Chen H, Chen K, Qiu X, Xu H, et al. The reproductive toxicity and potential mechanisms of combined exposure to dibutyl phthalate and diisobutyl phthalate in male zebrafish (*Danio rerio*). *Chemosphere.* 2020;258:127238.
73. Agarwal DK, Lawrence WH, Nunez LJ, Autian J. Mutagenicity evaluation of phthalic acid esters and metabolites in *Salmonella typhimurium* cultures. *J Toxicol Environ Health.* 1985;16(1):61–69.
74. Ali MM, Sahar T, Firyal S, et al. Assessment of cytotoxic, genotoxic, and oxidative stress of dibutyl phthalate on cultured bovine peripheral lymphocytes. *Oxid med Cell Longev.* 2022;2022:9961513.
75. Javaji K, Mamilla J, Deshpande SS, Kanaka RY, et al. Clastogenic, aneugenic, and tubulin polymerization properties of di-(2-ethylhexyl) phthalate and dibutyl phthalate. *Toxicol Ind Health.* 2023;39(9):504–514.
76. Kleinsasser NH, Wallner BC, Kastenbauer ER, Weissacher H, Harreus UA. Genotoxicity of di-butyl-phthalate and di-isobutyl-phthalate in human lymphocytes and mucosal cells. *Teratog Carcinog Mutagen.* 2001;21(3):189–196.
77. Chen F, Chien M, Lee C. Regulation of the cell cycle and P13K/AKT/mTOR signaling pathway by phthalates in normal human breast cells. *Taiwan J Obstet Gynecol.* 2023;62(3):434–439.
78. Chen F, Chien M, Lee C. The no-observed-adverse-effect level of phthalates promotes proliferation and cell cycle progression in normal human breast cells. *Taiwan J OBstet Gynecol.* 2023;62(6):874–883.
79. Li E, Xu B, Wei H, Bai Y, et al. Molecular mechanism of di-n-butyl phthalate promotion of bladder cancer development. *Toxicol In Vitro.* 2023;86:105508.
80. Hsieh T, Tsai C, Hsu C, et al. Phthalates stimulate the epithelial to mesenchymal transition through an HDAC6-dependent mechanism in human breast epithelial stem cells. *Toxicol Sci.* 2012;128(2):365–376.
81. Cao W, Zhao M, Chen Y, et al. Low-dose phthalates promote breast cancer stem cell properties via the oncogene $\Delta Np63\alpha$ and the Sonic hedgehog pathway. *Ecotoxicol Environ Saf.* 2023;252:114605.
82. Song P, Jiang N, Zhang K, et al. Ecotoxicological evaluation of zebrafish liver (*Danio rerio*) induced by dibutyl phthalate. *J Hazard Mater.* 2022;425:128027.
83. Jiang N, Song P, Li X, Zhu L, Wang J, Yin X. Dibutyl phthalate induced oxidative stress and genotoxicity on adult zebrafish (*Danio rerio*) brain. *J Hazard Mater.* 2022;424(Pt D):127749.
84. Sicinska P, Jablonska E, Bukowska B, Balcerczyk A, Reszka E. The selected epigenetic effects of phthalates: DBP, BBP and their metabolites: MBP, MBzP on human peripheral blood mononuclear cells (In vitro). *Toxicol In Vitro.* 2022;82:105369.
85. Sohn J, Kim S, Koschorreck J, Kho Y, Choi K. Alteration of sex hormone levels and steroidogenic pathway by several low molecular weight phthalates and their metabolites in male zebrafish (*Danio rerio*) and/or human adrenal cell (H295R) line. *J Hazard Mater.* 2016;320:45–54.

86. Chen M, Zhao S, Guo W, Zhu Y, et al. Maternal exposure to di-n-butyl phthalate (DBP) aggravate gestational diabetes mellitus via FoxM1 suppression by pSTAT1 signalling. *Ecotoxicol Environ Saf*. 2020;205:111154.
87. Arzuaga X, Walker T, Yost E, Radke E, Hotchkiss A. Use of the Adverse Outcome Pathway (AOP) framework to evaluate species concordance and human relevance of Dibutyl Phthalate (DBP)-induced male reproductive toxicity. *Reprod Toxicol*. 2023;96:445–458.
88. Kim H, Nam KS, Oh S, Son S, et al. Toxicological assessment of phthalates and their alternatives using human keratinocytes. *Environ Res*. 2019;175:316–322.
89. Kismali G, Yurdakok-Dikmen B, Kuzukiran O, Arslan P, Filazi A. Phthalate induced toxicity in prostate cancer cell lines and effects of alpha lipoic acid. *Bratisl Lek Listy*. 2017;118(8):460–466.
90. Chen B, Hu X, Zhen X, Hu G. Effects of dibutyl phthalate and di(2-ethylhexyl) phthalate with their metabolites on CYP2C9*1 and CYP2C19*1 activities in vitro. *J Pharm Biomed Anal*. 2018;160:195–201.
91. Huo S, Li B, Du J, Zhang X, et al. Dibutyl phthalate induces liver fibrosis via p38MAPK/NF-κB/NLRP3-mediated pyroptosis. *Sci Total Environ*. 2023;897:165500.
92. Batool S, Batool S, Shameem S, Batool T, Batool S. Effects of dibutyl phthalate and di-(2-ethylhexyl) phthalate on hepatic structure and function of adult male mice. *Toxicol Ind Health*. 2022;38(8):470–480.
93. Cheng L, Li J, Cheng J, Wu Z. Dibutyl phthalate-induced activation of ROS and ERK1/2 causes hepatic and renal damage in Kunming mice. *Hum Exp Toxicol*. 2019;38(8):938–950.
94. Liang F, Xi J, Chen X, Huang J, Jin D, Zhu X. Curcumin decreases dibutyl phthalate-induced renal dysfunction in Kunming mice via inhibiting oxidative stress and apoptosis. *Hum Exp Toxicol*. 2021;40(9):1528–1536.
95. Zhiwen X, Yongqing Z, Wenlan S, Shan H, et al. Dibutyl phthalate induces epithelial–mesenchymal transition of renal tubular epithelial cells via the Ang II/AMPKα2/Cx43 signaling pathway. *Toxicology*. 2023;494:153584.
96. Wang X, Yan X, Yang Y, et al. Dibutyl phthalate-mediated oxidative stress induces splenic injury in mice and the attenuating effects of vitamin E and curcumin. *Food Chem Toxicol*. 2020;136:110955.
97. You M, Song Y, Chen J, Liu Y, et al. Combined exposure to benzo(a)pyrene and dibutyl phthalate aggravates pro-inflammatory macrophage polarization in spleen via pyroptosis involving cathepsin B. *Sci Total Environ*. 2023;881:163460.
98. Yang R, Zheng J, Qin J, Liu S, et al. Dibutyl phthalate affects insulin synthesis and secretion by regulating the mitochondrial apoptotic pathway and oxidative stress in rat insulinoma cells. *Ecotoxicol Environ Saf*. 2023;249:114396.
99. Deng T, Zhang Y, Wu Y, et al. Dibutyl phthalate exposure aggravates type 2 diabetes by disrupting the insulin-mediated P13K/AKT signaling pathway. *Toxicol Lett*. 2018;290:1–9.
100. Li L, Huang L, Lei R, Zhang P, et al. DEHP and DBP, common phthalates, induce glucose metabolism disorders in rats via oxidative damage of PI3K/Akt/GLUT4 signaling. *Environ Pollut*. 2024;341:122948.
101. Kokai D, Stanic B, Tesic B, Nenadov DS, et al. Dibutyl phthalate promotes angiogenesis in EA.hy926 cells through estrogen receptor-dependent activation of ERK1/2, PI3K-Akt, and NO signaling pathways. *Chem Biol Interact*. 2022;366:110174.
102. Oluranti OI, Alabi BA, Michael OS, Ojo AO, Fatokun BP. Rutin prevents cardiac oxidative stress and inflammation induced by bisphenol A and dibutyl phthalate exposure via NRF-2/NF-κB pathway. *Life Sci*. 2021;284:119878.
103. Xie X, Deng T, Duan J, Ding S, et al. Comparing the effects of diethylhexyl phthalate and dibutyl phthalate exposure on hypertension in mice. *Ecotoxicol Environ Saf*. 2019;174:75–82.

104. Xu H, Dong X, Zhang Z, Yang M, et al. Assessment of immunotoxicity of dibutyl phthalate using live zebrafish embryos. *Fish Shellfish Immunol.* 2015;45(2):286–292.
105. Larson RP, Zimmerli SC, Comeau MR, Itano A, et al. Dibutyl phthalate (DBP) induced thymic stromal lymphopoietin (TSLP) is required for Th2 contact hypersensitivity responses. *J Immunol.* 2010;184(6):2974–2984.
106. Maruyama T, Shiba T, Iizuka H, Matsuda T, et al. Effects of phthalate esters on dendritic cell subsets and interleukin-4 production in fluorescein isothiocyanate-induced contact hypersensitivity. *Microbiol Immunol.* 2007;51(3):321–326.
107. Kurohane K, Sekiguchi K, Ogawa E, Tsutsumi M, Imai Y. Dibutyl phthalate rather than monobutyl phthalate facilitates contact hypersensitivity to fluorescein isothiocyanate in a mouse model. *Biol Pharm Bull.* 2017;40(11):2010–2013.
108. Li J, Li L, Zuo H, Ke C, et al. T-Helper type-2 contact hypersensitivity of Balb/c mice aggravated by dibutyl phthalate via long-term dermal exposure. *PLoS One.* 2014;9(2):e87887.
109. Shigeno T, Katakuse M, Fujita T, Mukoyama Y, Watanabe H. Phthalate ester-induced thymic stromal lymphopoietin mediates allergic dermatitis in mice. *Immunology.* 2009;128(Suppl 1):849.
110. Matsuda T, Maruyama T, Iizuka H, Kondo A, et al. Phthalate esters reveal skin-sensitizing activity of phenethyl isothiocyanate in mice. *Food Chem Toxicol.* 2010;48(6):1704–1708.
111. Maestre-Batlle D, Huff RD, Schwartz C, Alexis NE, et al. Dibutyl phthalate augments allergen-induced lung function decline and alters human airway immunology. A randomized crossover study. *Am J Respir Crit Care Med.* 2020;202(5):672–680.
112. Maestre-Batlle D, Nygaard UC, Huff RD, Alexis NE, et al. Dibutyl phthalate exposure alters T-cell subsets in blood from allergen-sensitized volunteers. *Indoor Air.* 2022;32(4):e13026.
113. Pan W, Zeng D, Ding N, Kuo K, et al. Percutaneous penetration and metabolism of plasticizers by skin cell and its implication in dermal exposure to plasticizers by skin wipe. *Environ Sci Technol.* 2020;54(16):10181–10190.
114. Imai Y, Kondo A, Iizuka H, Maruyama T, Kurohane K. Effects of phthalate esters on the sensitization phase of contact hypersensitivity induced by fluorescein isothiocyanate. *Clin Exp Allergy.* 2006;36(11):1462–1468.
115. Kruger T, Cao Y, Kjaergaard SK, Knudsen LE, Bonefeld-Jorgensen EC. Effects of phthalates on the human corneal endothelial cell line B4G12. *Int J Toxicol.* 2012;31(4):364–371.
116. Geier J, Lessmann H, Dickel H, Frosch PJ, et al. Patch test results with the metalworking fluid series of the German Contact Dermatitis Research Group (DKG). *Contact Dermatitis.* 2004;51(13):118–130.
117. Chowdhury MM, Statham BN. Allergic contact dermatitis from dibutyl phthalate and benzalkonium chloride in Timodine cream. *Contact Dermatitis.* 2002;46(1):57.
118. Cleys P, Hardy E, Bamai YA, et al. HBM4EU e-waste study: Occupational exposure of electronic waste workers to phthalates and DINCH in Europe. *Int J Hyg Environ Health.* 2024;255:114286.
119. Kwapniewski R, Kozaczka S, Hauser R, Silva MJ, et al. Occupational exposure to dibutyl phthalate among manicurists. *J Occup Environ Med.* 2008;50(6):705–711.
120. Nassan FL, Coull BA, Skakkebaek NE, Williams MA, et al. A crossover-crossback prospective study of dibutyl-phthalate exposure from mesalamine medications and semen quality in men with inflammatory bowel disease. *Environ Int.* 2016;95:120–130.
121. Fang H, Wang H, Zeng C, Fu H, et al. A preliminary cumulative risk assessment of diethylhexyl phthalate and dibutyl phthalate based on the inhibition of embryonic development via the PPAR γ pathway. *Toxicol In Vitro.* 2022;84:105430.

122. Mint A, Hotchkiss SA, Caldwell J. Percutaneous absorption of diethyl phthalate through rat and human skin in vitro. *Toxicol In Vitro*. 1994;8(2):251–256.
123. Wu H, Just AC, Colicino E, Calafat AM, et al. The associations of phthalate biomarkers during pregnancy with later glycemia and lipid profiles. *Environ Int*. 2021;155:106612.
124. Braun JM, Sathyanarayana S, Hauser R. Phthalate exposure and children's health. *Curr Opin Pediatr*. 2013;25(2):247–254.
125. Kim Y, Ha E, Kim E, Park H, et al. Prenatal exposure to phthalates and infant development at 6 months: Prospective mothers and children's environmental health (mocch) study. *Environ Health Perspect*. 2011;119(10):1495–1500.
126. Yolton K, Xu Y, Strauss D, Altaye M, et al. Prenatal exposure to bisphenol A and phthalates and infant neurobehavior. *Neurotoxicol Teratol*. 2011;33(5):558–566.
127. Al-Saleh I, Elkhatib R, Alnuwaysir H, et al. Exposure of preterm neonates receiving total parenteral nutrition to phthalates and its impact on neurodevelopment at the age of 2 months. *Sci Rep*. 2023;13(1):6969.
128. Sprowles J, Dzwilewski K, Merced-Nieves FM, Musaad S, et al. Associations of prenatal phthalate exposure with neurobehavioral outcomes in 4.5- and 7.5-month-old infants. *Neurotoxicol Teratol*. 2022;92:107102.
129. Cho S, Bhang S, Hong Y, Shin M, et al. Relationship between environmental phthalate exposure and the intelligence of school-age children. *Environ Health Perspect*. 2010;118(7):1027–1032.
130. Engel SM, Miodovnik A, Canfield RL, Zhu C, et al. Prenatal phthalate exposure is associated with childhood behavior and executive functioning. *Environ Health Perspect*. 2010;118(4):565–571.
131. Miodovnik A, Engel SM, Zhu C, Ye X, et al. Endocrine disruptors and childhood social impairment. *Neurotoxicology*. 2011;32(2):261–267.
132. Whyatt RM, Liu X, Rauh VA, Calafat AM, et al. Maternal prenatal urinary phthalate metabolite concentrations and child mental, psychomotor, and behavioral development at 3 years of age. *Environ Health Perspect*. 2012;120(2):290–295.
133. Radke EG, Braun JM, Nachman RM, Copper GS. Phthalate exposure and neurodevelopment: A systemic review and meta-analysis of human epidemiological evidence. *Environ Int*. 2020;137:105408.
134. Zhang Y, Zheng L, Chen B. Phthalate exposure and human semen quality in Shanghai: A cross-sectional study. *Biomed Environ Sci*. 2006;19(3):205–209.
135. Nobles CJ, Mendola P, Kim K, et al. Preconception phthalate exposure and women's reproductive health: Pregnancy, pregnancy loss, and underlying mechanisms. *Environ Health Perspect*. 2023;131(12):127013.
136. Frederiksen H, Sorensen K, Mouritsen A, Aksglaede L, et al. High urinary phthalate concentration associated with delayed pubarche in girls. *Int J Androl*. 2012;35(3):216–226.
137. Lopez-Carrillo L, Hernandez-Ramirez RU, Calafat AM, et al. Exposure to phthalates and breast cancer risk in northern Mexico. *Environ Health Perspect*. 2010;118(4):539–544.
138. Reeves KW, Santana MD, Manson JE, et al. Urinary phthalate biomarker concentrations and postmenopausal breast cancer risk. *J Natl Cancer Inst*. 2019;111(10):1059–1067.
139. Segovia-Mendoza M, Palacios-Arreola MI, Monroy-Escamilla LM, et al. Association of serum levels of plasticizers compounds, phthalates and bisphenols, in patients and survivors of breast cancer: A real connection? *Int J Environ Res Public Health*. 2022;19(13):8040.

140. Duty SM, Singh NP, Silva MJ, Barr DB, et al. The relationship between environmental exposures to phthalates and DNA damage in human sperm using the neutral comet assay. *Environ Health Perspect.* 2003;111(9):1164–1169.
141. Derakhshan A, Shu H, Broeren MAC, et al. Association of phthalate exposure with thyroid function during pregnancy. *Environ Int.* 2021;157:106795.
142. Main KM, Mortensen GK, Kaleva MM, Boisen KA, et al. Human breast milk contamination with phthalates and alterations of endogenous reproductive hormones in infants three months of age. *Environ Health Perspect.* 2006;114(2):270–276.
143. Loretz LJ, Api AM, Babcock L, et al. Exposure data for cosmetic products: facial cleanser, hair conditioner, and eye shadow. *Food Chem Toxicol.* 2008;46(5):1516–1524.
144. Loretz LJ, Api AM, Barraj LM, et al. Exposure data for cosmetic products: lipstick, body lotion, and face cream. *Food Chem Toxicol.* 2005;43(2):279–291.

7

Final Report on the Safety Assessment of Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate

Dibutyl Phthalate (DBP), Dimethyl Phthalate (DMP), and Diethyl Phthalate (DEP) are dialkyl phthalates used primarily in cosmetics at concentrations of less than 10 percent as plasticizers, solvents, and perfume fixatives.

These phthalates are rapidly absorbed, metabolized, and excreted. Acute animal feeding studies indicate that these ingredients are nontoxic. The results of most subchronic and chronic tests indicate that these ingredients are relatively nontoxic to rats. The oral administration of DBP produced testicular atrophy in various test rodents. The available data are not adequate to prove that these ingredients are teratogenic agents to experimental animals. This was not observed after the administration of DMP and DEP. Undiluted DBP, DMP, and DEP produced only minimal irritation to eyes of rabbits.

The mutagenic activity of DBP, DMP, and DEP toward *Salmonella typhimurium* mutants is essentially negative, but some assays reported positive findings. Carcinogenesis was not observed in DBP feeding studies.

Limited clinical data on DBP, DMP, and DEP indicate that these ingredients are not human skin irritants, sensitizers, or phototoxic agents. On the basis of the available data, it is concluded that these compounds are safe for topical application in the present practices of use and concentration in cosmetics.

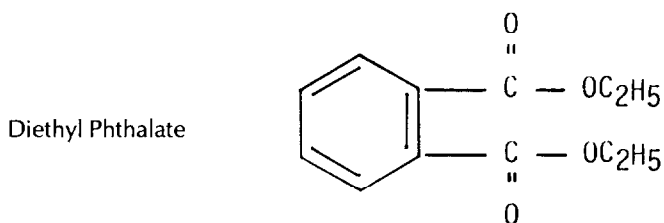
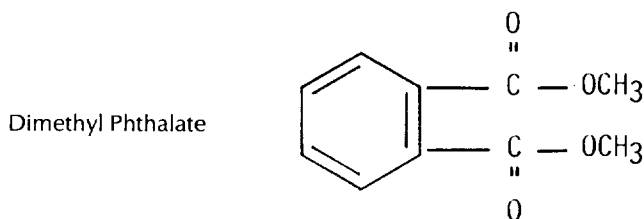
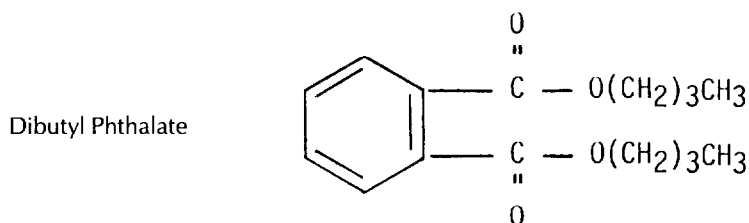
INTRODUCTION

This report reviews the published information and unpublished data supplied by the cosmetic industry on Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate. Di(2-ethylhexyl)phthalate, a compound currently of great concern, is not used in cosmetics.

CHEMICAL AND PHYSICAL PROPERTIES

Structure

Dibutyl Phthalate (CAS No. 84-74-2) (DBP), Dimethyl Phthalate (CAS No. 131-11-3) (DMP), and Diethyl Phthalate (CAS No. 84-66-2) (DEP) are dialkyl phthalates. DBP, DMP, and DEP are the aromatic diesters of butyl, methyl, and ethyl alcohol, respectively, and phthalic acid. The chemical formulas of these alkyl phthalates are as follows⁽¹⁾:



Properties

DBP, DMP, and DEP are colorless, oily liquids, soluble in alcohol, ether, and other common organic solvents and almost insoluble in water. DMP is insoluble in petroleum ether and other paraffin hydrocarbons. DBP is odorless. DMP and DEP have no to slight odors, and DEP has a bitter, disagreeable taste.⁽²⁻⁷⁾ DBP is soluble in a solution simulating human sweat (an aqueous solution containing 2.5 g sodium phosphate, 0.2 g triolein, and 2 drops Tween 85/1), and its solubility in this solution increases with an increase in pH.⁽⁸⁾ Chemical and properties of DBP, DMP, and DEP are presented in Table 1.

TABLE 1. Chemical and Physical Properties

| Property | DBP | DMP | DEP | Reference |
|----------------------------|-------------------|--------|--------|-----------|
| Molecular weight | 278.34 | 194.19 | 222.23 | |
| Specific gravity at: | | | | |
| 14/4°C | | | 1.232 | 7 |
| 15.6/15.6°C | | 1.196 | | 7 |
| 20°C | 1.0459, 1.0465 | | | 7 |
| 20/20°C | | 1.940 | | 7 |
| 20/20°C | 1.047, 1.049 | | | 2 |
| 20/20°C | | ~1.19 | | 3 |
| 20/20°C | | | ~1.12 | 4 |
| 20/20°C | 1.0484 | | | 6 |
| 25/25°C | | 1.189 | 1.120 | 6 |
| 25/25°C | | 1.189 | | 7 |
| Boiling point (°C) at: | | | | |
| 760 mm Hg | | 283.7 | | 7 |
| 400 mm Hg | | 257.8 | | 7 |
| 200 mm Hg | | 232.7 | | 7 |
| 100 mm Hg | | 210.0 | | 7 |
| 60 mm Hg | | 194.0 | | 7 |
| 40 mm Hg | | 182.8 | | 7 |
| 20 mm Hg | | 164.0 | | 7 |
| 10 mm Hg | | 147.6 | | 7 |
| 5 mm Hg | | 131.8 | | 7 |
| 1.0 mm Hg | | 100.3 | | 7 |
| Not specified | 340 | 282 | 295 | 5 |
| Not specified | 340.0 | 282 | 298 | 6 |
| Not specified | 340 | | 295 | 7 |
| Melting point (°C) | | | | |
| | -35 | | -40.5 | 6 |
| | | 5.5 | | 7 |
| Vapor pressure (mm Hg) at: | | | | |
| 20°C | | <0.1 | | 6 |
| 20°C | | <0.01 | | 7 |
| 150°C | 1.1 | | | 6 |
| 163°C | | | 14 | 6 |
| 182°C | | | 30 | 6 |
| 295°C | | | 734 | 6 |
| Refractive index at: | | | | |
| 14°C | | | 1.5049 | 7 |
| 20°C | 1.4900 | 1.5168 | | 7 |
| 25°C | 1.4915 | 1.5138 | 1.5002 | 6 |

Reactivity

The alkaline hydrolysis products of phthalate esters are mono- and diacids. The second-order alkaline hydrolysis rate constants in water at 30°C are 1.0×10^{-2} , 6.9×10^{-2} , and $2.5 \times 10^{-2} \text{ M}^{-1} \text{ sec}^{-1}$ for DBP, DMP, and DEP, respectively. Acid hydrolysis is generally slower than alkaline hydrolysis, and neutral hydrolysis is generally too slow to be detected.⁽⁹⁾ DBP is stable in solutions with a near neutral pH.⁽²⁾

The products of the thermal decomposition at 250 to 500°C of DBP are 1-butene, butanol, phthalic anhydride, and small amounts of benzoic acid, butyl

benzoate, phthalic acid, and monobutyl phthalate.⁽¹⁰⁾ The major products in the pyrolysis at 730°C of DBP are isobutene, butene, and propylene.⁽¹¹⁾

DBP can be degraded by radiolysis. The major product of a 1 ppm aqueous DBP solution at pH 7 after a dose of 3×10^4 rad of gamma radiation is monobutyl phthalate.⁽¹²⁾

Methods of Manufacture and Impurities

Phthalate esters can be prepared by the reaction of phthalic acid with alcohol. DBP, DMP, and DEP are produced industrially by the reaction of phthalic anhydride with butyl alcohol, methyl alcohol, and ethyl alcohol, respectively.^(6,7,13) DBP is manufactured by the esterification of phthalic anhydride with an excess of *n*-butyl alcohol. Vacuum stripping removes the unreacted *n*-butyl alcohol. Steam sparging ensures low odor. The phthalate is alkali refined to give a low acid number and is filtered to produce a clear product.⁽²⁾ The exact manufacturing processes for DMP and DEP are proprietary information. DEP may contain DMP or ethyl methyl phthalate as impurities.^(3,4)

DMP and DEP, for use in cosmetics, should contain minimums of 99 percent DMP and DEP, respectively, as determined by gas-liquid chromatography.^(2-4,14,15)

Analytical Methods

Qualitative and quantitative determinations of the phthalate esters are made by gravimetric procedures,⁽¹⁵⁻¹⁷⁾ titrimetric analysis,⁽¹⁵⁾ spectrophotometric methods,^(18,19) spectrophotofluorometric analysis,⁽²⁰⁾ the isotope dilution technique,⁽²¹⁾ thin-layer chromatography,^(22,23) liquid chromatography,⁽¹⁶⁾ liquid chromatography-mass spectrometry,⁽²⁴⁾ high-performance liquid chromatography,^(25,26) gas-liquid chromatography,^(3,4,27,28) gas chromatography,^(25,29,30) gas chromatography-mass spectrometry,^(31,32) high-resolution mass spectrometry, mass fragmentography,⁽³²⁾ gas chromatography with flame ionization,⁽²⁵⁾ vibration spectroscopy,⁽³³⁾ IR spectroscopy,^(14,16,17,34,35) UV spectroscopy,^(25,35,36) and NMR spectroscopy.^(34,35)

USE

Purpose in Cosmetics

DBP is used in cosmetics as a perfume solvent and fixative, as a suspension agent for solids in aerosols, as a lubricant for aerosol valves, as an antifoamer, as a skin emollient, and as a plasticizer in nail polish, fingernail elongators, and hair spray. DMP is used as a solvent, particularly for artificial musk, and as a plasticizer in fingernail elongators. DEP is used as a solvent for cellulose acetate in nail polish and dopes, as a fixative for perfume, as an alcohol denaturant in toilet preparations, and as a plasticizer in fingernail elongators.^(2-5,17,37)

Scope and Extent of Use in Cosmetics

Product types and the number of product formulations containing DBP,

DMP, or DEP and reported voluntarily to the Food and Drug Administration (FDA) in 1981 are presented in Table 2. Voluntary filing of this information by cosmetic manufacturers, packagers, and distributors conforms to the prescribed format of preset concentration ranges and product types as described in the Code of Federal Regulations (21 CFR 720.4)⁽³⁸⁾ Some cosmetic ingredients are supplied by the manufacturer at less than 100 percent concentration, and, therefore, the value reported by the cosmetic formulator or manufacturer may not necessarily reflect the true concentration of the finished product; the actual concentration in such a case would be a fraction of that reported to the FDA. The fact that data are only submitted within the framework of preset concentration ranges also provides the opportunity for overestimation of the actual concentration of an ingredient in a particular product. An entry at the lowest end of a concentration range is considered the same as one entered at the highest end of that range, thus introducing the possibility of a two- to ten-fold error in the assumed ingredient concentration. In 1981, DBP was reported as an ingredient in a total of 590 cosmetic formulations at concentrations ranging from ≤ 0.1 percent to between 10 and 25 percent. DMP was reported as an ingredient in 11 cosmetic formulations at concentrations ranging from ≤ 0.1 percent to between 10 and 25 percent. DEP was reported as an ingredient in 67 cosmetic formulations at concentrations ranging from ≤ 0.1 percent to between 25 and 50 percent.⁽³⁹⁾

Surfaces to which Commonly Applied

Cosmetic products containing DBP, DMP, or DEP may be applied to or come in contact with skin, eyes, hair, nails, mucous membranes, and respiratory epithelium (Table 2).⁽³⁹⁾

Frequency and Duration of Application

Product formulations containing DBP, DMP, or DEP may be applied as many as several times a day and may remain in contact with the skin for variable periods following application. Daily or occasional use may extend over many years (Table 2).⁽³⁹⁾

Potential Interactions with Other Cosmetic Ingredients

No interactions of DBP, DMP, or DEP with other cosmetic ingredients are reported. In typical formulations, the compounds are stable.⁽²⁻⁴⁾

Noncosmetic Uses

DBP, DMP, and DEP are used as solvents and plasticizers for nitrocellulose, cellulose acetate, and cellulose acetate-butyrate compositions. They are used in the manufacture of varnishes and plastics and in insecticides and insect repellents. DBP is used as a plasticizer in explosives and elastomers, such as polyvinyl, as a textile lubricating agent, as a resin solvent, and in safety glass, printing inks, paper coatings, and adhesives. DMP is used as a camphor substitute in the manufacture of celluloid, as a wetting agent, and as an alcohol denaturant.^(6,7)

DBP, DMP, and DEP may be used, at no specific concentration limits, in adhesives used as components of articles intended for packaging, transporting, or holding food (21 CFR 175.105).⁽³⁸⁾ DBP may be used as a catalyst and crosslinking

TABLE 2. Product Formulation Data⁽³⁹⁾

| Product Category | Total No. of Formulations in Category | Total No. Containing Ingredient | No. of Product Formulations within Each Concentration Range (percent) | | | | | |
|---|---------------------------------------|---------------------------------|---|--------|-------|------|--------|------|
| | | | >25-50 | >10-25 | >5-10 | >1-5 | >0.1-1 | ≤0.1 |
| <i>Dibutyl Phthalate</i> | | | | | | | | |
| Other hair preparations (noncoloring) | 177 | 3 | — | — | — | — | 3 | — |
| Other hair coloring preparations | 49 | 3 | — | — | — | — | 3 | — |
| Other makeup preparations (not eye) | 530 | 1 | — | — | — | — | 1 | — |
| Nail basecoats and undercoats | 44 | 36 | — | — | 8 | 28 | — | — |
| Nail polish and enamel | 767 | 522 | — | 3 | 61 | 168 | 127 | 163 |
| Nail polish and enamel remover | 41 | 3 | — | 1 | — | 1 | 1 | — |
| Other manicuring preparations | 50 | 14 | — | 1 | 2 | 9 | — | 2 |
| Other personal cleanliness products | 227 | 5 | — | — | — | 5 | — | — |
| Aftershave lotions | 282 | 3 | — | — | — | — | 3 | — |
| 1981 TOTALS | | 590 | — | 5 | 71 | 211 | 138 | 165 |
| <i>Dimethyl Phthalate</i> | | | | | | | | |
| Hair conditioners | 478 | 2 | — | — | — | — | 2 | — |
| Tonics, dressings, and other hair grooming aids | 290 | 2 | — | — | — | 1 | 1 | — |
| Wave sets | 180 | 2 | — | — | — | — | 2 | — |
| Other hair preparations (noncoloring) | 177 | 4 | — | — | — | — | 4 | — |
| Hair rinses (coloring) | 76 | 1 | — | — | — | — | 1 | — |
| 1981 TOTALS | | 11 | — | — | — | 1 | 10 | — |

| | | | | | | | | |
|---|------|----|---|---|---|---|----|----|
| <i>Diethyl Phthalate</i> | | | | | | | | |
| Bath oils, tablets, and salts | 237 | 3 | — | — | — | 1 | — | 2 |
| Other bath preparations | 132 | 2 | — | — | — | — | — | 2 |
| Eye shadow | 2582 | 1 | — | — | — | — | — | 1 |
| Colognes and toilet waters | 1120 | 19 | — | — | — | 1 | 10 | 8 |
| Perfumes | 657 | 23 | 1 | — | — | 1 | 7 | 14 |
| Fragrance powders (dusting and talcum, excluding aftershave talc) | 483 | 1 | — | — | — | — | 1 | — |
| Sachets | 119 | 3 | — | — | — | 1 | 2 | — |
| Other fragrance preparations | 191 | 2 | 1 | — | — | — | 1 | — |
| Hair sprays (aerosol fixatives) | 265 | 5 | — | — | — | 2 | 3 | — |
| Wave sets | 180 | 1 | — | — | — | — | 1 | — |
| Nail polish and enamel remover | 41 | 1 | — | — | — | 1 | — | — |
| Bath soaps and detergents | 148 | 1 | — | — | — | — | 1 | — |
| Aftershave lotions | 282 | 3 | — | — | — | — | 3 | — |
| Face, body, and hand skin care preparations (excluding shaving preparations) | 832 | 1 | — | — | — | — | — | 1 |
| Other skin care preparations | 349 | 1 | — | — | — | — | 1 | — |
| 1981 TOTALS | | 67 | 2 | — | — | 7 | 30 | 28 |

agent for epoxy resins, and DEP may be used as a plasticizer, at no specific concentration limits, in the resinous and polymeric coatings of the food-contact surface of articles intended for use in producing, manufacturing, packing, processing, preparing, treating, packaging, transporting, or holding food. DBP may be used in coatings of containers having a capacity of ≥ 1000 gallons and intended for repeated use with alcoholic beverages of less than or equal to 8 percent alcohol by volume (21 CFR 175.300, 175.320).⁽³⁸⁾ There are no concentration limits for the use of DBP as a component of the uncoated or coated food-contact surface of paper and paperboard intended for use in producing, manufacturing, packaging, processing, preparing, treating, packing, transporting, or holding aqueous and fatty foods (21 CFR 176.170).⁽³⁸⁾ DBP may be used in the base sheet or coating of cellophane used in packaging food, but total phthalates must not exceed 5 percent by weight of the finished cellophane (21 CFR 177.1200).⁽³⁸⁾ DBP and DMP may be used, at no specific concentration limits, as solvents for inhibitors, accelerators, and catalysts in crosslinked polyester resins used as articles or components of articles intended for repeated use in contact with food (21 CFR 177.2420).⁽³⁸⁾ DBP may be used as a plasticizer in rubber articles intended for repeated use in producing, manufacturing, packing, processing, preparing, transporting, or holding food. Total DBP may not exceed 30 percent by weight of the rubber product (21 CFR 177.2600).⁽³⁸⁾ There is no concentration limit for the use of DMP in semirigid and rigid acrylic and modified acrylic plastics used as articles intended for use in contact with food (21 CFR 177.1010).⁽³⁸⁾ There is no limit in the amount of DEP that may be used in surface lubricants used in the manufacture of metallic articles that contact food (21 CFR 178.3910).⁽³⁸⁾ DEP may be used as a plasticizer in the manufacturer of food-packaging materials with no specific limits. This DEP will not be considered a "food additive" if of good commercial grade, suitable for association with food, and used in accordance with good manufacturing practice; the amount of DEP that migrates into food as a result of its use in food-packaging materials should not be intended to accomplish any physical or technical effect in the food itself and should be reduced to the least amount reasonably possible (21 CFR 181.22, 181.27).⁽³⁸⁾

GENERAL BIOLOGY

Microbial Metabolism and Toxicity

A variety of bacteria can use DBP or DMP as a carbon source. The corresponding monoesters, phthalic acid, and protocatechuic acid are intermediates in the degradation of these chemicals.^(40,41)

The growth of *Pseudomonas aeruginosa* was not inhibited by concentrations of up to 1000 ppm DMP. A 1500 ppm solution slightly inhibited the growth of the organism. After a 24-hour incubation, the concentration of a 98 ppm DMP solution decreased to 88 ppm, suggesting some bacterial utilization of the compound.⁽⁴¹⁾ The concentration of neutralized DEP that inhibited the multiplication of *Pseudomonas putida* was greater than 400 ppm.⁽⁴²⁾ The minimum inhibitory concentration of a 10 percent solution of DEP in 95 percent ethanol was 1000 ppm for *Corynebacterium* sp. and greater than 1000 ppm for *Staphylococcus aureus* and *Escherichia coli*.⁽⁴³⁾

The growth of the blue-green alga, *Microcystis aeruginosa*, was inhibited by 100 to 300 ppm of DMP and suppressed for 3 days by 400 ppm DMP. After 4 days, cellular lysis was observed in the 400 ppm DMP culture. Concentrations of DMP from 500 ppm to 800 ppm completely destroyed the cells within 72 hours.^(44,45) Neutralized DEP inhibited the multiplication of *M. aeruginosa* at a concentration of 15 ppm and inhibited the multiplication of the green alga, *Scenedesmus quadricauda*, at a concentration of 10 ppm.^(42,46)

A 10 ppm solution of DBP in phosphate buffer at pH 7 decreased the percent survival of the yeast, *Saccharomyces cerevisiae*, throughout a 48-hour incubation; a 20 ppm solution was even more toxic.⁽⁴⁷⁾ The minimum inhibitory concentration of a 10 percent (w/v) solution of DEP in 95 percent ethanol was 500 ppm for the fungus, *Candida albicans*.⁽⁴³⁾

A concentration of 50 ppm of DBP completely inhibited the growth of cells of the protozoan, *Tetrahymena pyriformis*. Other phthalate esters were inhibitory as well.⁽⁴⁸⁾ A concentration of 1000 ppm of DMP markedly inhibited the growth rate of *T. pyriformis*.^(44,49) Neutralized DEP inhibited the multiplication of the flagellate protozoan, *Entosiphon sulcatum* at a concentration of 19 ppm.⁽⁴²⁾

In Vitro Cell Toxicity

The metabolism and toxicity of DBP, DMP, and DEP in cultures of mouse fibroblast and rat cerebellum and various human cell lines have been investigated.

Dose-response curves were produced, and the ID_{50} for the mouse fibroblast cultures, defined as the dose required to inhibit growth by 50 percent, was determined for the phthalates. The ID_{50} s for DBP, DMP, and DEP were 1×10^{-4} , 7×10^{-3} , and 3×10^{-3} mole/l, respectively. DMP was highly toxic to the cells when they were undergoing significant protein turnover.⁽¹³⁾ The effect of DMP on a replicating mouse fibroblastic cell culture was investigated. A radioactively labeled amino acid mixture (^{14}C) was added to the cultures, and the radioactivity was followed over a 96-hour incubation. Cells were relatively insensitive to growth inhibition by DMP, as measured by uptake of radioactivity, for the first 24 hours. However, between 24 and 96 hours, the uptake of radioactivity decreased continuously.⁽⁵⁰⁾

Toxicity to mouse fibroblasts was also investigated using the cell overlay method. Pads containing 0.05 ml of a 50 mg/ml emulsion of the phthalates were placed on the agar surface (2.5 mg phthalates/pad), and the cells were observed for 48 hours. DMP and DEP were toxic to the cells and DBP was not.⁽⁵¹⁾ In another study, mouse fibroblastic cells were incubated for 24 hours with paper discs containing pure DMP and DEP or saline solutions saturated with DMP and DEP at pH 6. Only the pure DMP was toxic to the cells.⁽⁵²⁾ Other researchers have reported that all three phthalates were toxic in a 24-hour incubation of mouse fibroblastic cells.⁽⁵³⁾ The response of mouse fibroblastic cells to 1, 5, 10, and 50 percent suspensions of DBP, DMP, and DEP was studied by Oser et al.⁽⁵⁴⁾ All the suspensions were toxic except the 1 and 5 percent suspensions of DBP. In cell suspensions with DBP and DEP, the cellular ATP concentrations decreased over a 6-hour incubation.

The effects of DBP, DMP, and DEP on the outgrowth of nerve fibers and fibroblasts in primary cultures of rat cerebellum were investigated. The phthalates were added directly to the nutrient media. DBP and DEP completely inhibited

outgrowth at concentrations greater than or equal to 1.17×10^{-3} and 1.53×10^{-3} M, respectively. DMP did not completely inhibit outgrowth at concentrations less than or equal to 3.05×10^{-3} M.⁽⁵⁵⁾

Human embryonic lung cell cultures were studied after the addition of 40 μ g/ml of DBP to the culture medium. DBP inhibited cell growth and caused morphological changes in the cells, the appearance of lipid drops in the cytoplasm, and the accumulation of triacylglycerol in the cytosol.⁽⁵⁶⁾

Thelestam et al.⁽⁵⁷⁾ found that DBP and DEP were inactive in a test in which the extent of membrane damage in human lung fibroblasts was determined by measuring the amount of a radioactively labeled cytoplasmic marker released into the media. The ID_{50} of DBP, defined as the concentration that caused 50 percent growth inhibition, for human diploid cell strain WI-85 was 1.35×10^{-4} M.⁽⁵⁸⁾

Guess and Haberman⁽⁵²⁾ studied the effects of DBP, DMP, and DEP on human amnion and KB human cancer cells in culture. All three compounds killed and lysed the cells. Saline solutions saturated with DMP and DEP at pH 6 did not cause hemolysis of human erythrocytes.

HeLa cells were incubated for 7 days after the addition of DBP, DMP, and DEP to the culture medium. The 7-day IC_{50} s, the geometrical mean values between the totally inhibitory concentrations and the maximal completely noninjurious ones, were 3.1×10^{-2} M for DBP, 7.7×10^{-2} M for DMP, and 6.3×10^{-2} M for DEP.⁽⁵⁹⁾

Effects on Enzymes

Phthalate esters have a variety of different effects on mammalian enzymes, both in vivo and in vitro. DBP and DMP affect drug-metabolizing enzymes in mammalian liver. Single-dose intraperitoneal administration of 3.05 ml/kg of DBP and 3.6 ml/kg of DMP to rats inhibited the activity of hepatic aminopyrine N-demethylase and aniline hydroxylase and had no effect on glucose-6-phosphatase, NADPH-cytochrome c reductase, and tyrosine aminotransferase activity. The activities of these enzymes were not decreased when the phthalates were administered intraperitoneally every day for 7 days.^(60,61) Results of another study indicated that DBP weakly enhanced the activity of aminopyrine N-demethylase from rat hepatic 10,000 g supernatant.⁽⁶²⁾ The oral administration of 5 mmole/kg per day of DBP for 6 days to male rats increased the hepatic cytochrome P-450, had no effects on glutathione-S-transferase activity or the monooxygenase activities dependent on cytochrome P-450, increased the epoxide hydratase activity, and increased the conjugation of o-aminophenol and 4-methylumbelliferone with glucuronic acid. Rat liver incubated in vitro with 2×10^{-3} M DBP had no effect on epoxide hydratase or glutathione-S-transferase activities, decreased the monooxygenase activities, and decreased the conjugation of o-aminophenol and 4-methylumbelliferone with glucuronic acid.⁽⁶³⁾

DBP, DMP, and DEP inhibited mitochondrial respiration. Concentrations of 5×10^{-5} to 1×10^{-3} M of the phthalates inhibited the respiration of isolated mitochondria from rat liver primarily by uncoupling oxidative phosphorylation rather than by inhibiting electron transport or energy transfer.^(64,65) Other researchers using the same concentrations have suggested that the contrary is probably true; the phthalates inhibited electron transport or energy transfer.⁽⁶⁶⁾ In some studies,

DBP and DMP inhibited the activities of succinate dehydrogenase and ATPase, enzymes of the rat liver inner mitochondrial membrane, after intraperitoneal administration, and in in vitro assays at concentrations of 1×10^{-4} to 1.5×10^{-3} M.^(62,65,67) DBP stimulated ATPase activity and induced swelling of rat liver mitochondria.⁽⁶⁸⁾

Administration of 0.7 percent DBP or 0.5 percent DMP in the diet of male rats for 21 days increased hepatic weights and reduced serum cholesterol concentrations. Acetate incorporation into triglycerides and the steryl ester plus squalene and mevalonate incorporation into squalene plus sterols in liver minces were inhibited by dietary DBP. These results were not observed with DMP. DMP administration resulted in a decrease in total hepatic cholesterol and lipid. This was not observed with DBP.⁽⁶⁹⁾ The intraperitoneal administration of 20 mg/kg per day of DBP to mice for 16 days did not significantly lower serum cholesterol but did lower serum triglycerides. DBP, at a concentration of 2.5×10^{-6} M inhibited mouse liver homogenate acetyl-CoA synthetase, citrate lyase, and acetyl-CoA carboxylase but not fatty acid synthetase. These enzymes are involved in the cholesterol and triglyceride synthesis pathways.⁽⁷⁰⁾ A 5×10^{-6} M concentration of DBP and DEP inhibited in vitro human blood lecithin/cholesterol acyltransferase. DMP, at the same concentration, inhibited the enzyme slightly.⁽⁷¹⁾

DBP elevated the activities of mouse and rat serum lactate dehydrogenase, glutamic-oxalacetic transaminase, and glutamic-pyruvate transaminase.⁽⁷²⁻⁷⁴⁾ DBP increased the activity of alkaline phosphatase in mice⁽⁷²⁾ but had no effect on this enzyme in rats.⁽⁷³⁾

ABSORPTION, DISTRIBUTION, METABOLISM, AND EXCRETION

DMP was absorbed through human skin, and some of its metabolites were detected in human urine.⁽⁷⁵⁾

A homogenate of rat epidermis metabolized DMP at approximately 1.5 percent of the rate of the metabolism of DMP by a homogenate of rat liver. The homogenates were compared on a mg wet weight basis. DMP was bound to the epidermis in quantities seven to eight times greater than those in which it was bound to an equal dry weight of hepatic tissue.⁽⁷⁶⁾

DEP was absorbed through the skin of rabbits. Labeled DEP (^{14}C) was applied topically, and the application sites were covered with cotton patches. Analysis of urine indicated that approximately 9 percent of the radioactivity was excreted after 24 hours, 14 percent after 48 hours, and 16 to 20 percent within 72 hours. After 3 days of topical exposure, tissue distribution was determined by autoradiography. Radioactivity was detected in the lung, heart, liver, kidney, gonads, spleen, and brain. It was not detected in the skin and subdermal fatty tissue at the site of application.⁽¹³⁾

DBP was administered by gavage to male rats in two doses of 0.2 ml 24 hours apart. Urine was collected for 48 hours after the first dose, and DBP and its metabolites were quantitated. A total of 24.6 percent of the phthalate moiety was recovered in the urine. The recovered phthalate moiety consisted of 89.8 percent monobutyl phthalate (MBP), 2.7 percent phthalic acid (PA), 0.4 percent intact DBP, and four other metabolites in very small amounts. The researchers suggested that DBP was metabolized by hydrolysis of one ester bond and both terminal and subterminal oxidation of the remaining alkyl chain. The resulting primary

and secondary alcohols were, presumably, further oxidized to acids and ketones, respectively. DMP was administered by gavage to male rats in a single dose of 0.1 ml, and the urine was collected for 24 hours. A total of 44.6 percent of the phthalate moiety was recovered in the urine, and it consisted of 77.5 percent mono-methyl phthalate, 14.4 percent PA, and 8.1 percent intact DMP. DMP appeared to be metabolized only by hydrolysis of one or both ester groups.⁽⁷⁷⁾

Male mice were administered labeled DBP (¹⁴C) orally or intravenously. The radioactivity accumulated in the liver and kidney within 6 hours of oral administration and within 1 hour of intravenous administration. The radioactivity was rapidly excreted in the urine and feces.⁽⁷⁸⁾

DBP interacted with DNA *in vitro*, but after oral administration of labeled DBP (¹⁴C) to mice, no radioactivity was recovered from hepatic DNA. DBP and its metabolites appeared not to be transported into the nuclei.⁽⁷⁹⁾

Labeled DBP (¹⁴C) was administered orally in dimethyl sulfoxide in a dose of 60 mg/kg or intravenously in saline in a dose of 10 mg/kg to male rats. Urine and feces were collected, and the amount of radioactivity excreted was determined. The percentage of administered radioactivity excreted varied from 81.4 to 97.7 in the urine and from 1.0 to 8.2 in the feces in the first 24 hours after oral or intravenous administration of DBP. Several rats were killed, and tissue distribution of radioactivity was determined. Brain, heart, liver, lung, spleen, muscle, adipose, stomach, prostate, and thymus tissues, blood, and the intestinal contents were examined 24 hours after oral or intravenous administration of DBP. Very little radioactivity was recovered. The elimination of DBP from tissues and organs was rapid, and no organ had any significant affinity for accumulation. Rats were administered labeled DBP (¹⁴C) orally, and bile was collected. From 27.6 to 52.8 percent of radioactivity was excreted in the bile within 24 hours after oral administration of DBP. Since more radioactivity was excreted in the bile than in the feces, there was apparently good absorption of DBP and its metabolites from rat intestine. Urinary metabolites were identified in male rats, male hamsters, and male guinea pigs given a single oral dose of 60 mg/kg DBP. All 24-hour urine samples contained MBP as the major product, intact DBP, PA, MBP glucuronide, and two other MBP oxidation products. The hamster urine contained an additional oxidation product. The livers from rats were examined 1 hour after intravenous dosing of DBP, and the data obtained indicated that DBP was rapidly hydrolyzed to MBP by the microsomal fraction. No PA was detected. The bile contained MBP and intact DBP but not PA. Since PA was detected in the urine, it was suggested that its formation must occur at other sites than the liver. It was concluded that the hydrolysis of DBP to MBP occurred in the liver, that there was entero-hepatic circulation of DBP and its metabolites and good absorption from the intestine, and that MBP was the main metabolite of DBP and was primarily excreted in urine.⁽⁸²⁾

DBP and DMP, in concentrations of 0.4 mg/ml, were incubated at 37°C with rat liver and kidney homogenates. DBP and DMP almost completely disappeared after 2 hours of incubation with rat liver homogenates. The action of rat kidney homogenates was slower; however, approximately 90 percent of the DBP and 95 percent of the DMP disappeared during a 5-hour incubation. The phthalates were found not to be degraded spontaneously under these experimental conditions.⁽⁸⁰⁾

A 500 mg/kg dose of labeled (¹⁴C) DBP in ethanol was administered by gastric

intubation to male rats and the bile was collected every hour for 6 hours. Six hours after oral administration of DBP, 4.5 percent of the radioactivity was recovered in the bile. Five hours after intravenous injection of DBP, 10 percent of the radioactivity was detected in the bile. DBP bile metabolites included MBP, intact DBP, PA, an MBP glucuronide, and traces of other glucuronides. A small amount of DBP appears to be absorbed unaltered from the intestine, and the excretion of DBP through the biliary route has a role in its metabolic fate.⁽⁸⁰⁾

Labeled DBP, DMP, and DEP (¹⁴C) were incubated with rat, ferret, and baboon hepatic postmitochondrial supernatant and with intestinal-mucosal cell homogenates. All of the diesters were hydrolyzed by cell homogenates. They were all hydrolyzed by all the preparations, and greater than 90 percent of the total metabolite formed was the corresponding monoester. Baboon liver preparations hydrolyzed the diesters faster than rat liver preparations; ferret liver preparations were the least active. Baboon intestinal-mucosal cell preparations hydrolyzed the diesters faster than rat intestinal-mucosal cell preparations, and ferret intestinal-mucosal cell preparations were the least active.⁽⁸¹⁾ Hepatic preparations from humans also catalyzed the monohydrolysis of DBP, DMP, and DEP. The toxic effects of phthalates administered orally may depend on the properties of the corresponding monoesters and/or alcohols.⁽⁸²⁾

DBP, DMP, and DEP, in concentrations of 1 mg/ml, were incubated for 16 hours at 37°C with the contents of rat stomach, small intestines, or cecum or with suspensions of human feces. The phthalates were metabolized rapidly to the corresponding monoesters when incubated with the contents of rat small intestine. Metabolism was slower in the presence of rat cecal contents and only DMP was metabolized to any extent by rat stomach contents. Human feces were almost inactive in metabolizing the phthalates; DBP and DMP were metabolized faster than DEP. The intestinal contents of younger male rats metabolized DBP and DMP at a slower rate than intestinal contents from more mature male rats. Among adults, intestinal contents from male rats metabolized DBP at a faster rate than intestinal contents from female rats. The monoesters were the only products of metabolism; complete hydrolysis to PA did not occur. It may be significant toxicologically that there is a good correlation between rate of phthalate hydrolysis and the acute oral toxicity to rats that is reported in the literature. The more rapidly hydrolyzed phthalate esters are more toxic. In another experiment, rat intestinal contents were incubated at 37°C for 90 minutes or centrifuged or filtered before addition of DMP. Preincubation reduced the ability of the small intestine contents to degrade DMP. The enzymes involved in DMP metabolism appeared to be labile *in vitro*. Both centrifugation and filtration reduced the rate of DMP hydrolysis. The effect of antibiotics was studied by adding antibiotics to the incubation mixture or to the intestinal contents during the 90-minute preincubation period. The antibiotics used in the experiments were antibacterial enzymes. They had no effect on the rate of metabolism of DMP by small intestine contents, suggesting that the involved enzymes are not bacterial and more probably are mammalian in origin. Mucosal cell enzymes may be involved in DMP metabolism. The low rate of phthalate hydrolysis by rat cecal contents and human feces might be explained by the presence of a low number of active intestinal mucosal cells.⁽⁸³⁾ DBP was hydrolyzed by crude pancreatic lipase solution.⁽⁸⁴⁾

The *in vitro* intestinal absorption of DBP and DMP was studied using an everted gut-sac preparation from the rat small intestine. In one experiment with

DBP, S,S,S,-tributylphosphorotrithioate (DEF), administered orally before gut-sac preparation, was used as an esterase inhibitor. Most of the DBP and DMP was hydrolyzed to the corresponding monoester before crossing the intestinal mucosa. Only 4.5 percent of the DBP and 18.8 percent of the DMP crossed the intestine intact. Inhibition of mucosal esterases by DEF reduced the amount of DBP hydrolyzed to MBP. Approximately the same amount of intact DBP was absorbed by the intestine with and without DEF, and DEF did not affect MBP absorption. Intestinal absorption of these compounds may be controlled by the hydrolysis of DBP to MBP.⁽⁸⁵⁾

Labeled DEP (¹⁴C) was administered intravenously to pregnant rats on Day 5 or Day 10 of gestation. Diester and/or metabolic products were present in maternal blood, fetal tissue, amniotic fluid, and placentas after Day 8 or Day 11, respectively, and throughout gestation.⁽⁸⁶⁾

Phthalates are ubiquitous in the environment, and human exposure is likely. DBP was found in normal and diseased kidneys,^(87,88) adipose tissue at autopsy,⁽⁸⁹⁾ in the blood of pregnant women, and in umbilical cords.⁽⁹⁰⁾ Possible routes of exposure to phthalates for humans are by oral or dermal contact, inhalation, or as a result of the use of medical devices, such as blood storage bags.⁽⁴⁴⁾

ANIMAL TOXICOLOGY

Oral Studies

Acute Toxicity

The acute oral toxicity of DBP, DMP, and DEP was studied in rats,^(84,91-97) mice,^(72,91,98-100) rabbits,^(91,101) guinea pigs, and chicks⁽⁹¹⁾ (Table 3). The LD₅₀ for rats administered DBP orally ranged from approximately 8 g/kg to 23.0 g/kg. The LD₅₀ value for rats administered DMP orally was 6.9 ml/kg. In the Hodge and Sterner⁽¹⁰²⁾ classification of single-dose oral toxicity for rats, DBP and DMP would be classified as practically nontoxic to relatively harmless and as practically nontoxic, respectively.

The acute oral toxicity of two nail preparations, one containing 9 percent DBP and one containing 6 percent DBP, was studied in rats^(104,105) (Table 3). Both preparations were practically nontoxic.

Subchronic and Chronic Toxicity

DBP, DMP, and DEP in corn oil were administered by oral intubation for 4 days to groups of 12 rats in doses of 7.2 mmole/kg per day (approximately 2.0 g/kg per day DBP, 1.4 g/kg per day DMP, and 1.6 g/kg per day DEP). There were no significant changes in food intake or body weight. DMP and DEP administration did not result in significant changes in testes weight, no testicular atrophy was observed, and urinary zinc excretion was unaffected. Administration of DBP decreased weight of testes and produced severe atrophy of the seminiferous tubules. Most of the tubules had complete loss of spermatocytes and spermatids. DBP administration was accompanied by an increase in the urinary excretion of zinc, and there was a decrease in the zinc content of testes on an absolute and relative weight basis.⁽¹⁰⁶⁻¹⁰⁸⁾ The administration of zinc, concurrently with DBP, provided substantial protection against DBP-produced testicular damage.⁽¹⁰⁶⁾

TABLE 3. Acute Oral Toxicity

| <i>Material Tested</i> | <i>Method</i> | <i>Species of Animal</i> | <i>LD₅₀</i> | <i>Comments</i> | <i>Reference</i> |
|---------------------------------|---|--------------------------|----------------------------|---|------------------|
| DBP | — | Rats | 23.0 g/kg | — | 92,97 |
| DBP | — | Rats | 12.5 g/kg | — | 93,96 |
| DBP | — | Rats | 14.95 g/kg | — | 103 |
| DBP | — | Mice | 9 g/kg | — | 95 |
| DBP | Animals were observed for 7 days following DBP administration | Rats | >20 ml/kg | — | 72,98 |
| DBP | Oral administration of 200 mg DBP to 10 mice | Male mice | 9.77 g/kg | — | 99 |
| Undiluted DBP | Oral administration of 4, 8, 16, and 32 g/kg, of DBP to 3, 9, 6, and 6 rats, respectively | Mice | — | 6/10 of the mice died within 7 hours. | 84 |
| DBP | Animals were observed for 7 days following DBP administration | Rats | ~8 g/kg | 0/3, 4/9, 6/6, and 6/6 rats died, respectively. The 4 g/kg dose had no effect on growth, the 8 g/kg dose slightly inhibited growth, and the 16 and 32 g/kg dose groups succumbed too quickly to exhibit significant changes in growth | 98 |
| DBP | Animals were observed for 7 days following DBP administration | Male mice | Between 14.8 and 17.0 g/kg | — | 98,100 |
| Undiluted DMP | 40 rats, 80 guinea pigs, and 120 chicks were fasted prior to DMP administration. 110 mice and 80 rabbits were not fasted. Animals were observed for 6 days following DMP administration. DMP was given to 10 animals/dose | Rats | 6.9 ml/kg | — | 91 |
| | | Mice | 7.2 ml/kg | | |
| | | Rabbits | 4.4 ml/kg | | |
| | | Guinea pigs | 2.4 ml/kg | | |
| | | Chicks | 8.5 ml/kg | | |
| DEP | — | Rabbits | 1.0 g/kg | — | 101 |
| DEP | — | Rats | 8.2 ml/kg | — | 94 |
| Nail polish, 9 percent DBP | 5 male and 5 female animals/dose were fasted 16 hours prior to oral intubation and were observed for 14 days after | Rats | >5 ml/kg | No signs of gross pathology on necropsy of rats receiving 5 ml/kg | 104 |
| Nail preparation, 6 percent DBP | Preparation administered by oral intubation to 10 animals | Rats | >5 g/kg | "Nontoxic" | 105 |

DBP produced testicular atrophy in the rat, mouse, guinea pig, and ferret, but not in the hamster after oral administration in a dose of 2.0 g/kg per day for 10 days.⁽¹⁰⁸⁾

Long-term oral toxicity of DBP, DMP, and DEP was studied in rats,^(84,94-96,103,109-113) mice,⁽¹⁰³⁾ and rabbits⁽¹¹⁴⁾ (Table 4). Except at dietary concentrations of 1.25 percent DBP for 1 year, 8.0 percent DMP for 2 years, and 5.0 percent DEP for 16 weeks, the phthalates were relatively nontoxic to rats in subchronic and chronic oral tests.

Dermal Studies

Acute Toxicity

The acute dermal toxicity of DMP to rabbits was determined by placing DMP in contact with the clipped skin and holding it in place with a rubber cuff. The rabbits were exposed for 24 hours and then observed for 2 weeks. The acute dermal LD₅₀ of DMP to rabbits was greater than 10 ml/kg.⁽⁹¹⁾

Subchronic and Chronic Toxicity

DBP and DMP were tested for long-term dermal toxicity by applying 0.5, 1.0, 2.0, and 4.0 ml/kg per day for 90 days to the clipped, intact skin of rabbits. The chemicals were applied to approximately 10 percent of the body surface. The subchronic dermal LD₅₀ of DBP to rabbits was greater than 4 ml/kg per day for 90 days. DBP was slightly irritating to skin and very irritating to rabbit penile mucosa. A slight dermatitis was observed, and in the 4 ml/kg dosed rabbits, slight renal damage (not further described) was observed.⁽⁹⁵⁾ The subchronic dermal LD₅₀ of DMP to rabbits was also greater than 4 ml/kg per day for 90 days. No skin irritation or dermatitis was observed, although DMP was irritating to rabbit penile mucosa. Pulmonary edema and slight renal damage were observed in the rabbits that died during the study. Rabbit survivors had varying degrees of nephritis (not further described) at the two highest doses.⁽⁹¹⁾

Primary Irritation

DBP and DMP were applied to the clipped, intact, and abraded skin of 3 rabbits. The rabbits were exposed to 0.5 ml of the chemicals for 24 hours with an occluded patch. DBP caused "very slight irritation." DMP was not irritating except in molting areas and the Primary Irritation Index (PII) was 0.7.^(91,95)

DMP was treated for primary irritation to rabbits using a pill box device. Pill boxes were affixed to shaved rabbit skin, and 0.1 ml of a 20 percent solution of formalin ("as the primary irritant") was painted onto the skin and allowed to dry. Discs containing 0.2 ml of DMP were placed in the pill box and the box was closed. A 0.25 ml volume of a 0.5 percent sterile Evans blue solution was injected intravenously. After 18 hours, the blue color at the pill box sites was evaluated and correlated with irritancy. Ten to 15 separate observations were made. DMP had an irritation score of 0.8 on a scale of 0 to 3; DMP was less than slightly irritating.⁽¹¹⁵⁾

Sensitization

No evidence of sensitization was observed in rabbits receiving daily topical applications of DBP and DMP at doses of up to 4.0 ml/kg per day for 90 days.⁽⁹⁵⁾

Intradermal Irritation

The intradermal irritation of phthalates to rabbits was measured by injecting the phthalates into the skin of the shaven backs. A trypan blue solution was injected into the marginal ear vein, and the extravasated trypan blue at the injection site was used as a measure of the extent of the inflammatory response. In one study, 0.2 ml of 100 mg/ml phthalate emulsions was injected. DBP gave a mild inflammatory response after 10 minutes and a moderate response after 26 minutes. A rapid and marked inflammatory response to DMP and DEP was noted.⁽⁵¹⁾ Other researchers used cottonseed oil as a diluent for DBP, DMP, and DEP. DBP was not irritating, but DMP and DEP produced a significant degree of irritation.⁽¹¹⁶⁾ In another study, saline solutions saturated with the phthalates were administered. No response was observed to DMP and DEP.⁽⁵²⁾

Eye Irritation

The eye irritation potential of DBP, DMP and DEP was studied in rabbits.⁽¹¹⁶⁻¹¹⁸⁾ The eye irritation potential of nail preparations containing 9 percent DBP and 6 percent DBP also was investigated^(105,119) (Table 5). DBP, DEP, and nail preparations containing DBP were relatively nonirritating to the rabbit eye. With long contact time, undiluted DMP may be injurious to the eyes of rabbits.

Inhalation Studies

Male rats were exposed to 1.5 mg/m³ of DBP vapor for 6 hours per day and 6 days per week for approximately 1 month. There were no significant effects on body or organ weights when the rats were compared to controls. No significant toxic effects were observed.⁽⁷³⁾ Rats were exposed to 0.5 mg/m³ and 50 mg/m³ of DBP mist for 6 hours per day for 6 months. Rats exposed to either concentration had smaller weight gains and greater brain and lung weights than control rats. The higher concentration had a greater effect than the lower concentration.⁽¹²⁰⁾

Intraperitoneal Studies

Acute Toxicity

Acute intraperitoneal toxicity of DBP, DMP, and DEP was studied in mice^(51, 52, 116, 121) and in rats⁽¹²²⁾ (Table 6). The acute intraperitoneal LD₅₀s for rats for DBP, DMP, and DEP were 3.05 ml/kg, 3.38 ml/kg, and 5.06 ml/kg, respectively.

Subchronic and Chronic Toxicity

DEP was administered intraperitoneally in a dose of 2 ml/kg per day to rabbits for 8 days. "Temporary distress" was observed during and after administration. There was no paralysis or other abnormal effect. The intraperitoneal administration of 1.5 ml/kg per day of DEP to guinea pigs for 8 days did not result in any permanent ill effects during or after the experiment.⁽¹¹⁴⁾ A DEP emulsion was administered intraperitoneally in a dose of 125 mg/kg per day for 6 weeks to 20 to 30 mice. There was slight retardation in weight gain and some evidence of peritonitis. The organ:body weight ratios for liver, heart, lungs, kidneys, spleen, and testes of treated mice were not different from the control mice ratios. No abnormal hematological patterns were observed.⁽⁵¹⁾

TABLE 4. Subchronic and Chronic Oral Toxicity

| <i>Material Tested</i> | <i>Dose and Vehicle</i> | <i>Length of Study</i> | <i>Number and Species of Animals</i> | <i>Results</i> | <i>Reference</i> |
|------------------------|--|------------------------|---|--|------------------|
| DBP | 1 ml/kg in oil 2 times a week | 6 weeks | Rats | No adverse effects were reported | 109,113 |
| DBP | 20 mg/kg | 11 weeks and 3 days | Rats | Leukocytosis was observed in rats. Mouse growth was inhibited | 103 |
| DBP | 0.12 and 1.2 g/kg per day suspensions in olive oil | 3 months | Mice 10 male and 10 female rats/dose, 40 control rats given only olive oil | 1/10 rat from the high dose group died. No specific cause of death was determined. Both DBP doses produced a statistically significant increase in the animals' mean liver weight. No histological evidence of any pathologic changes were found in the liver, kidneys, and spleen | 96 |
| DBP | 2.5 mg/kg per day | 6 months | Rats | No adverse effects were observed | 94,112 |
| DBP | 0.125 percent in the feed | 1 year | 20 male and 20 female rats in dosed and in control groups | 6/40 rats from the dosed group died. No specific cause of death was determined. No "remarkable" alterations were observed upon gross and histological examination of liver, kidneys, and spleen of dosed rats | 96 |
| DBP | 0.01, 0.05, 0.25, and 1.25 percent in the feed | 1 year | 10 rats/dose, 10 control rats | At 0.25 percent in the diet or lower, there was no effect on growth or survival. At 1.25 percent in the diet, 5/10 rats died during the first week. The remaining rats gained weight as did the controls. No rats exhibited significant changes in the number or distribution of elements in the peripheral blood or specific gross pathological changes | 84 |
| DBP | 1 ml/kg in oil 2 times a week | 1½ years | Rats | No pathological changes observed. No effects on hematological parameters or on organ weights | 109,111, 133 |

| | | | | | |
|-----|---------------------------------------|--------------------|---|---|-----|
| DMP | 2.0, 4.0, and 8.0 percent in the feed | 2 years | 10 female rats/dose | 2.0 percent in the feed had no effect on growth. 4.0 and 8.0 percent had a slight but significant effect on growth. Chronic nephritis seen in rats on 8.0 percent in diet. Mortality rates were not different from those for control rats | 95 |
| DEP | 3 ml/kg per day | 8 days | Rabbits | The rabbits appeared normal for the 8 days and for 2 weeks afterwards. "Temporary distress" was observed after DEP administration | 114 |
| DEP | 0.2, 1.0, and 5.0 percent in the feed | 2, 6, and 16 weeks | 5 male and 5 female rats in dosed and control groups on diet for 2 and 6 weeks. 6 rats of each sex, litter mate-paired, in 5.0 percent diet and control groups for 16 weeks. 15 rats of each sex in dosed and control groups on diet for 16 weeks | No changes in behavioral patterns or clinical signs of toxicity were observed. Both sexes on 5.0 percent feed and females on 1.0 percent feed consumed less food and gained less weight than the controls. There was a pattern of reduction in absolute weight and an increase in relative weight of the brain, spleen, heart, kidneys, adrenal glands, gonads, and pituitary of rats on the 5.0 percent diet. A pattern of increases in absolute and relative weights was observed in livers and various parts of the GI tract in these rats. Both liver and kidneys were enlarged but histologically normal | 110 |

TABLE 5. Rabbit Eye Irritation

| <i>Material Tested</i> | <i>Method</i> | <i>Results</i> | <i>Reference</i> |
|---------------------------------|--|---|------------------|
| DBP | Undiluted DBP instilled into eyes. Eyes examined at 0.5, 1.0, 1.5, 2.0, 2.5, 3.0, 24.0, and 48.0 hours postinstillation | No grossly observable irritation at any examination time | 116 |
| DMP | 0.5 ml undiluted DMP applied to corneal center while eyelids are retracted. Lids released after 1 minute. Eye injury scored on a scale of 0–20 points after 18–24 hours | Injury score was >0.1 and <5.0. 5.0 is the level representative of severe injury; necrosis visible after staining and covering ~75 percent of the surface of the eye | 117 |
| DMP | 0.1 ml undiluted DMP instilled into the conjunctival sac of the eyes. Injury scored on a scale of 0–110 points after 1 and 24 hours | Score was 3.3 after 1 hour and 2.2 after 24 hours | 118 |
| DMP | Undiluted DMP instilled into eyes. Eyes examined at 0.5, 1.0, 1.5, 2.0, 2.5, 3.0, 24.0, and 48.0 hours postinstillation | No grossly observable irritation at any examination time | 116 |
| DEP | 0.1 ml undiluted DEP instilled into the conjunctival sac of the eyes. Injury scored on a scale of 0–110 points after 1 and 24 hours | Score was 3.2 after 1 hour and 1.5 after 24 hours | 118 |
| DEP | Undiluted DEP instilled into eyes. Eyes examined at 0.5, 1.0, 1.5, 2.0, 2.5, 3.0, 24.0, and 48.0 hours postinstillation | No grossly observable irritation at any examination time | 116 |
| Nail polish, 9 percent DBP | 0.1 ml instilled into the conjunctival sac of one eye of 9 rabbits. Lids held together for 1 second. In 3 rabbits, treated eye washed at 30 seconds with 20 ml water. Scored on a scale of 0–110 at 24, 48, and 72 hours and 4 and 7 days postinstillation | Unwashed eyes' average score were 11.3, 9.7, 6.8, 4.8, and 0.5 and washed eyes' average scores were 8.3, 7.7, 4.0, 2.7, and 0.3 at 24, 48, and 72 hours and 4 and 7 days postinstillation, respectively | 119 |
| Nail preparation, 6 percent DBP | 0.1 ml instilled into conjunctival sac of one eye of 6 rabbits. Lids held together for 1 second. Ocular reactions recorded at 24, 48, and 72 hours | No positives for conjunctival redness or chemosis, keratitis, or iritis. "Nonirritating" | 105 |

A series of doses of DBP, DMP, and DEP was injected intraperitoneally into groups of 10 male mice 5 days a week. The apparent LD_{50} was calculated each week until it remained constant for 3 weeks; this was the chronic LD_{50} . DBP, DMP, and DEP reached chronic LD_{50} s in 25, 18, and 14 weeks, respectively. The chronic LD_{50} values were 0.85 ml/kg per day 5 days a week for DBP, 1.18 ml/kg per day 5 days a week for DMP, and 1.39 ml/kg per day 5 days a week for DEP.⁽¹¹⁶⁾

Other Studies

The acute intravenous LD_{50} of DBP to male mice was 0.72 g/kg.^(72,98) DEP, in a 3 percent acacia suspension, was administered to an anesthetized rabbit through the jugular vein. The DEP was administered in repeated doses of 50 mg/kg to a total dose of 650 mg/kg (time between doses was not given). The first six doses caused a transient fall in blood pressure. The total dose of 650 mg/kg did not cause death or significant change in the animal. Five doses of the 3 percent acacia vehicle did not produce any blood pressure changes.⁽⁵¹⁾ A 0.25 ml/kg dose of DEP in saline was injected slowly into the femoral vein of a dog. At first, respiration was stimulated and then it was paralyzed. The intravenous administration of 0.5 ml of DEP into a rabbit ear vein caused convulsions "similar to those produced by strychnine" within a few minutes. The symptoms "soon" disappeared and the rabbit appeared normal. A larger dose was fatal to rabbits by causing paralysis of respiration.⁽¹¹⁴⁾

The intramuscular administration of DBP in a dose of 4 g/kg to 3 rats and 8 g/kg to 3 rats did not result in any deaths, and there was no effect on the growth of the rats.⁽⁸⁴⁾

The subcutaneous LD_{50} of DEP to guinea pigs was greater than or equal to 3 g/kg.^(44,101)

SPECIAL STUDIES

Animal Reproduction and Teratology

DBP in doses of 2 and 4 ml/kg, DMP in doses of 0.5, 1, and 2 ml/kg, and saline in a dose of 4 ml/kg were administered intraperitoneally on Days 3, 6, and 9 of gestation to groups of 5 pregnant female rats. Day 1 of gestation was the day sperm were found in vaginal smears. Five control rats survived, and four of those implanted. Five and four rats survived, and four and three implanted, respectively, in the 2 and 4 ml/kg DBP groups. DBP administration resulted in a 50 percent reduction in the number of pups weaned per litter. Two male pups, one from each of two litters in the 2 ml/kg DBP group, had no eyes. In the 0.5, 1, and 2 ml/kg DMP groups, 5, 2, and 5 rats survived, and 4, 1, and 5 implanted, respectively. The numbers of pups weaned were not significantly different from the controls.⁽¹²³⁾ In another study in which Day 1 of gestation was the day after sperm were found in vaginal smears, groups of 5 pregnant female rats were administered DBP, DMP, and DEP intraperitoneally, in doses of 1/3, 1/5, and 1/10 of a previously determined acute intraperitoneal LD_{50} (3.05 ml/kg for DBP, 3.4 ml/kg for DMP, and 5.06 ml/kg for DEP), on Days 5, 10, and 15 of gestation (Table 7). Control rats were untreated or were administered distilled water, normal saline, or cottonseed oil. The rats were killed on Day 20, 1 day before expected

TABLE 6. Acute Parenteral Toxicity

| <i>Material Tested</i> | <i>Method</i> | <i>No. and Species of Animals</i> | <i>LD₅₀</i> | <i>Comments</i> | <i>Reference</i> |
|------------------------|--|--|---------------------------|---|------------------|
| DBP | Single IP injection of 4 dose levels ranging from 0.5 to 16 g/kg | Mice | 4.00 g/kg | — | 51 |
| DBP | Single IP injection. Pathological changes observed up to 72 hours. Deaths recorded for 7 days | Female mice | 14.9 mmole/kg | Pulmonary congestion, edema, and petechial hemorrhage, toxic reaction in spleen, and renal tubular degeneration observed after 72 hours | 121 |
| DBP | Undiluted DBP administered and animals observed 7 days for deaths; 2 ml/kg administered IP, 2 mice sacrificed at Days 1, 2, 4, 7, and 10, and heart, lung, kidneys, spleen, liver, pancreas, and bowel examined histologically for irritation | 10 male mice/dose to determine LD ₅₀ ; 10 mice of unspecified sex for histology | 3.57 g/kg (3.41 ml/kg) | No evidence of significant intra-peritoneal irritation | 116 |
| DBP | Animals observed for 7 days after IP injection | Female rats | 3.05 ml/kg | — | 122 |
| DMP | Single IP injection of 4 doses ranging from 0.5 to 16 g/kg | Mice | 1.58 g/kg | — | 51 |
| DMP | Single IP injection of saline saturated with DMP. 25 ml/kg DMP | Mice | — | No deaths observed | 52 |
| DMP | Single IP injection. Pathological changes observed up to 72 hours. Deaths recorded for 7 days | Female mice | 18.8 mmole/kg | Pulmonary congestion and atelectasis, toxic reaction in spleen and lymph nodes, and renal tubular necrosis observed after 72 hours | 121 |
| DMP | Undiluted DMP administered and animals observed 7 days for deaths; 2 ml/kg administered IP, 2 mice sacrificed at Days 1, 2, 4, 7, and 10, and heart, lungs, kidneys, spleen, liver, pancreas, and bowel examined histologically for irritation | 10 male mice/dose to determine LD ₅₀ ; 10 mice of unspecified sex for histology | 3.98 g/kg (3.35 ml/kg) | No evidence of significant intra-peritoneal irritation | 116 |
| DMP | Animals observed for 7 days after IP injection | Female rats | 3.38 ml/kg | — | 122 |

| | | | | | |
|-----|--|--|------------------------|--|-----|
| DEP | Single IP injection of 4 doses ranging from 0.5 to 16 g/kg | Mice | 2.83 g/kg | — | 51 |
| DEP | Single IP injection of saline saturated with DEP. 25 ml/kg DEP | Mice | — | No deaths observed | 52 |
| DEP | Single IP injection. Pathological changes observed up to 72 hours. Deaths recorded for 4 days | Female mice | 12.4 mmole/kg | Pulmonary congestion, edema and petechial hemorrhage, toxic reaction in spleen, and renal tubular degeneration observed after 72 hours | 121 |
| DEP | Undiluted DEP administered and animals observed 7 days for deaths; 2 ml/kg administered IP, 2 mice sacrificed at Days 1, 2, 4, 7, and 10, and heart, lungs, kidneys, spleen, liver, pancreas, and bowel examined histologically for irritation | 10 male mice/dose to determine LD ₅₀ ; 10 mice of unspecified sex for histology | 3.22 g/kg (2.87 ml/kg) | No evidence of significant intra-peritoneal irritation | 116 |
| DEP | Animals observed for 7 days after IP injection | Female rats | 5.06 ml/kg | — | 122 |

TABLE 7. Embryotoxic and Teratogenic Effects of Phthalates⁽¹²²⁾

| <i>Treatment Groups</i> | <i>Volume Injected* (ml/kg)</i> | <i>Number of Corpora Lutea</i> | <i>Number of Resorption†</i> | <i>Number of Dead Fetuses†</i> | <i>Number of Live Fetuses†</i> | <i>Number of Cross Abnormalities‡</i> | <i>Number of Skeletal Abnormalities**</i> |
|-------------------------|---------------------------------|--------------------------------|------------------------------|--------------------------------|--------------------------------|---------------------------------------|---|
| Untreated controls | None | 60 | 0 | 0 | 59 (100) | 0 | 0 |
| Distilled water | 10.00 | 59 | 4 (6.8) | 0 | 55 (93.2) | 0 | 0 |
| Normal saline | 10.00 | 62 | 7 (11.5) | 0 | 54 (88.5) | 1 (1.9) | 4 (14.3) |
| Cottonseed oil | 10.00 | 59 | 4 (6.8) | 0 | 55 (93.2) | 1 (1.8) | 3 (10.7) |
| | 5.00 | 54 | 3 (6.4) | 0 | 44 (93.6) | 0 | 0 |
| DBP | 1.017 | 64 | 23 (36.5) | 0 | 40 (63.5) | 0 | 8 (33.3) |
| | 0.610 | 56 | 2 (3.6) | 0 | 53 (96.4) | 0 | 7 (24.1) |
| | 0.305 | 56 | 4 (7.3) | 0 | 51 (92.7) | 0 | 6 (20.7) |
| DMP | 1.125 | 55 | 17 (32.1) | 5 (9.4) | 31 (58.5) | 4 (11.1) | 9 (75.0) |
| | 0.675 | 55 | 0 | 1 (1.9) | 52 (98.1) | 4 (7.5) | 6 (35.3) |
| | 0.338 | 65 | 21 (33.3) | 0 | 42 (66.7) | 4 (9.5) | 4 (25.0) |
| DEP | 1.686 | 57 | 2 (3.6) | 0 | 54 (96.4) | 0 | 13 (81.3) |
| | 1.012 | 59 | 0 | 0 | 57 (100.0) | 0 | 8 (47.1) |
| | 0.506 | 65 | 28 (44.4) | 0 | 35 (55.6) | 0 | 5 (26.3) |

*5 pregnant female rats injected IP on Days 5, 10, and 15 of gestation and sacrificed on Day 20.

†Numbers in parentheses are percent values based on total number of implantations.

‡Numbers in parentheses are percent values based on total number of viable and nonviable fetuses.

**Numbers in parentheses are percent values based on total number of stained fetuses. Generally 30–50 percent of the fetuses were stained.

parturition. Phthalate administration did not interfere with fertility, as reflected by corpora lutea:implantation site ratio. However, there were significant effects upon embryonic and/or fetal development. The average weights of the fetuses from the treated groups and those administered saline were significantly lower than the average weight of the fetuses from the untreated controls. The investigator normally selected 30 to 50 percent of the fetuses for visualization of skeletal abnormalities. There was a significantly higher number of skeletal abnormalities in the fetuses from the test group as compared to the controls.⁽¹²²⁾ The failure to include historical control data, as well as a positive control in the test program, makes it difficult to evaluate the significance of the results.

DBP was administered in the feed to pregnant mice throughout gestation, and the mice were killed on Day 18. Day 0 of gestation was the day on which a vaginal plug was found. DBP was administered in five dietary concentrations from 80 to 2100 mg/kg. Implantation was not affected, but resorptions and fetal deaths increased with dosage. Maternal weight gain was depressed at the higher dosages and was due to increased embryonic or fetal death. Two of three live fetuses from the 2100 mg/kg DBP group had neural tube defects. Ossification was depressed, but malformation and resorption rates and fetal weights were not significantly affected by DBP administration up to 350 mg/kg per day.⁽¹²⁴⁾ In another study, 120 and 600 mg/day of DBP in olive oil were administered by gavage to groups of 10 female rats for approximately 3 months prior to their being mated. Additional groups of female rats received the same doses for 21 days following fertilization. The uteri and fetuses from all the rats were removed on Day 21 of gestation. Fetuses from treated and control rats did not differ significantly in number of sternum ossification foci, in development of the bones of the base of the skull or in the paws of the front and hind extremities, and in rib fusion. The administration of DBP before gestation did not cause any significant changes in other measured parameters. Administration of DBP to pregnant rats did result in lower placental weights, and fetal weights were significantly lower in the high DBP dose group. There were 4, 2, and 22 resorptions in the control, 120, and 600 mg/day DBP groups, respectively.⁽⁹⁶⁾

The dietary administration of DBP, in doses of 10 and 100 mg/kg per day, to two mouse strains for three generations increased the formation of renal cysts in the F₁ and F₂ generations.^(98,125) In another three-generation reproduction study, female rats were dosed daily for 6 weeks with 50 percent DBP solution in oil, at a dose of 1 ml/kg, and then were paired with untreated males. The offspring were bred to produce two additional generations; it is not known whether the second and third generations were dosed with DBP. No impairment of reproductive performance was noted. Development, growth, and fertility were normal for all three generations.^(109,111,113)

Mutagenesis

The mutagenic activity of DBP, DMP, and DEP for *Salmonella typhimurium* mutants depended on the assay protocol. In the standard Ames test,⁽¹²⁶⁾ DBP and DEP were negative in *S. typhimurium* strains TA98, TA100, TA1535, and TA1537 with and without metabolic activation.^(98,127,128) DEP was also negative in these strains when using a preincubation protocol.⁽¹²⁹⁾ In a liquid suspension assay with a 4-hour incubation, DBP, DMP, and DEP were positive in strain TA100

without metabolic activation and negative with metabolic activation.⁽¹³⁰⁾ In a modified Ames test in which histidine and biotin were incorporated into the bottom agar, DBP, DMP, and DEP were negative with and without metabolic activation in strain TA98, and DBP was negative with and without metabolic activation in strain TA100. DMP and DEP were positive in strain TA100 without metabolic activation, and the response was dose related. They were negative in strain TA100 with metabolic activation.^(76,131) DBP and DEP were not mutagenic to *E. coli*.^(98,132)

DNA repair enzyme-deficient *Bacillus subtilis* and *E. coli* were equally or less sensitive to DBP and DEP than the wild-type bacteria.^(98,128) Rosenkranz and Leifer⁽¹³³⁾ reported that DBP did not affect either the wild-type or the DNA repair enzyme-deficient *E. coli*. There were no measurable zones of growth inhibition for either strain.

DBP was tested for mutagenic activity by reversion analysis of the yeast, *S. cerevisiae*. DBP had no mutagenic effect on the yeast whether the test was conducted with or without metabolic activation.⁽⁴⁷⁾

The effects of DBP on Chinese hamster cell chromosome aberrations and sister chromatid exchanges (SCEs) have been investigated in several studies. In one study DBP was negative for chromosome aberrations and SCEs.⁽¹²⁸⁾ In another study, the mitotic index was not appreciably decreased when the cells were exposed to DBP in ethanol. A significant increase over the vehicle for the number of SCEs was found, but no dosage effect was found for chromosome aberrations or SCEs.⁽¹³⁴⁾ DBP in a 0.2 percent bovine albumin solution was examined in a third study. The percentages of chromosome aberrations with DBP and with bovine albumin were 6 and 1.8 percent, respectively. These results did not conclusively prove that DBP caused chromosome aberrations; DBP was called a suspicious compound by the researchers.⁽¹³⁵⁾ DBP, DMP, and DEP, in doses of 0.25 mg/ml, had no effect on chromatid aberrations in human leukocyte cultures compared to controls.⁽¹³⁶⁾

Carcinogenesis

DBP was noncarcinogenic, but specific details of experiments are lacking.^(128,134) Carcinogenesis has not been observed in 18-month or longer DBP feeding studies in rats.⁽¹³⁷⁾

Di(2-ethylhexyl)phthalate (DEHP), a compound currently of great concern, is not used in cosmetics. DEHP was tested in a National Toxicology Program carcinogenesis bioassay and was carcinogenic in both rats and mice.⁽¹³⁸⁾

CLINICAL ASSESSMENT OF SAFETY

Dermal Studies

Patch tests have been performed on human subjects with the phthalates.^(139,140) The cosmetic industry has conducted studies on the skin irritation, sensitization, and photosensitization of a variety of products containing DBP⁽¹⁴¹⁻¹⁴⁸⁾ (Table 8). DBP, DMP, and DEP in concentrations of 2 percent in petrolatum and DBP at 5 percent in petrolatum were nonirritating in 48-hour closed patch tests; the 2 percent concentrations were tested on 1532 subjects with 1

positive reaction, and the 5 percent DBP was tested on 53 subjects with no positive reactions. Products containing DBP in concentrations ranging from 4.5 to 9 percent were tested at a concentration of 100 percent. A nail polish containing 9 percent DBP was slightly irritating in a 23-hour patch test on 13 subjects and not irritating in a 48-hour patch test on 25 subjects. The nail polish was tested in a modification of the maximization test on 25 subjects, and no contact sensitization was observed. A deodorant containing 4.5 percent DBP was tested in an antiperspirant efficacy test on 43 subjects; the deodorant was not irritating. It was slightly irritating in a 21-day cumulative irritancy test on 12 subjects. The deodorant was not an allergen in a modification of the repeated insult patch test on 200 subjects. A nail preparation containing 6 percent DBP was tested on 99 subjects in a prophetic patch test, on 48 subjects in a repeated insult patch test, and on 47 subjects in a controlled use study; the nail preparation was nonirritating and non-sensitizing. The nail preparation was also tested for photosensitization in the prophetic patch test and the repeated insult patch test; it was nonphotosensitizing.

Other Studies

A chemical worker accidentally swallowed approximately 10 g of DBP. The worker's symptoms included nausea, vomiting, dizziness, headache, pain and irritation in the eyes, conjunctivitis, and toxic nephritis. He recovered completely after 2 weeks.^(94,149)

Proper treatment of a human corneal burn caused by DMP resulted in healing within 48 hours and no loss of vision.⁽¹⁵⁰⁾

The health status of 147 workers subjected to prolonged occupational exposure to mixtures of phthalate plasticizers (including DBP) was investigated; many workers had a moderately pronounced toxic polyneuritis.⁽¹⁵¹⁾

SUMMARY

DBP, DMP, and DEP are dialkyl phthalates. They are primarily used in cosmetics at concentrations of less than 10 percent as plasticizers, solvents, and perfume fixatives.

Some bacteria can use DBP and DMP as carbon sources. These two phthalates and DEP may inhibit the growth of or be toxic to bacteria, algae, yeast, and protozoa. The phthalates may also inhibit the growth of or be toxic to mouse fibroblast, rat cerebellum, and various human cell lines. The phthalate esters have a variety of different effects on mammalian enzymes, both in vivo and in vitro.

Radioactive DBP, after oral administration to rats, hamsters, and guinea pigs, is rapidly metabolized to monobutyl phthalate and other products, and these metabolites are excreted in the urine and feces. In rats, the biliary route seems to be important in the metabolic fate of DBP. Only small amounts of radioactivity are found in rat tissues and organs after oral administration of labeled DBP (¹⁴C). DMP is absorbed through human skin. Labeled DEP (¹⁴C) was absorbed through the skin of rabbits, and the radioactivity was distributed throughout the body and excreted in the urine. Within several days of the intravenous administration of DEP to pregnant rats, DEP and its metabolic products were found in maternal blood, fetal tissue, amniotic fluid, and placentas.

TABLE 8. Skin Irritation and Sensitization

| <i>Material Tested</i> | <i>Concentration (percent)</i> | <i>Method</i> | <i>Number of Subjects</i> | <i>Results</i> | <i>Reference</i> |
|----------------------------|---|--|---|--|------------------|
| DBP | 5 percent in petrolatum | 48-hour closed patch test on back. Readings 48 and 72 hours after patch application | 7 men and 46 women who wore dentures and suffered from "burning mouth syndrome" | No positive reactions | 139 |
| DBP, DMP, and DEP | All phthalates at 2 percent in petrolatum | 48-hour closed patch test. Joint study by International Contact Dermatitis Research Group | 1532 | 1 positive reaction | 140 |
| Nail polish, 9 percent DBP | 100 | 21 23-hour patches on same site on back. 1 hour rest. Scored (0-3) 1 hour after patch removal | 2 men and 11 women | Composite total score was 247 (out of a possible maximum of 819). "Slightly irritating" | 147 |
| Nail polish, 9 percent DBP | 100 | 48-hour occluded patch to back or forearm | 25 | No irritation observed | 148 |
| Nail polish, 9 percent DBP | 100 | Modification of the maximization test. ⁽¹⁵²⁾ 5 48-hour occluded induction patches on back or forearm with 24-hour rests in between them; 24-hour sodium lauryl sulfate (SLS) pretreatment before first patch. 10-day rest period. 1-hour SLS pretreatment followed by 48-hour challenge patch; scored at patch removal and 24 hours later | 25 | "No instances of contact sensitization" | 148 |
| Deodorant, 4.5 percent DBP | 100 | Antiperspirant efficacy test, normal use conditions (Federal Register 43:46694-732, October 10, 1978). Applied 0.5 g/day for 2 days | 43 | "No irritation observed" in the axillary region | 146 |
| Deodorant, 4.5 percent DBP | 100 | Modification of the repeated insult patch test. ^(153,154) 8 48 hour induction patches of 0.2 g on upper arms. 2-week rest 72-hour challenge at original and new sites. Scored at 48 and 96 hours (0-3) | 41 | 9 reactions of 1 (mild erythema) at induction and 1 equivocal reaction at original site at 96-hour challenge observation. "Not an allergen under conditions of the test" | 144 |

| | | | | | |
|---|-----|--|--------------------|--|-----|
| Deodorant, 4.5 per- cent DBP | 100 | 21-day cumulative irritancy test. ⁽¹⁵⁵⁾ 21 24-hour occluded patches of 0.3 g applied to the back over 21 days. Each patch scored at removal (0-4) | 1 man and 11 women | Total score calculated on the basis of 10 subjects was 140.8 out of a possible maximum of 840. "Slightly irritating" | 143 |
| Deodorant, 4.5 per- cent DBP | 100 | Modification of the repeated insult patch test. ^(153,154) 10 24-48 hour occlusive induction patches of 0.2 g ~ 3 times a week. Patches applied to the back. 10-day rest. 72-hour challenge patch. Scored after patch removal (0-4) | 159 | 4 equivocal reactions and 1 score of 1 (erythema) during induction. One equivocal and 1 score of 1 at challenge. "Does not appear to be an allergen under test conditions" | 145 |
| Nail prepara- tion, 6 percent DBP | 100 | Prophetic patch test. ⁽¹⁵⁶⁾ 2 24-hour open and closed patches 10 to 14 days apart. 1 open patch irradiated for 1 minute at a distance of 12 in with a Hanovia Tanette Mark I lamp (UV). Scored at patch removal and daily for 5 days thereafter (1+ to 3+) | 99 | No positive reactions were observed. "Nonirritating, nonsensitizing, nonphotosensitizing" | 141 |
| Nail prepara- tion, 6 percent, DBP | 100 | Repeated insult patch test. ⁽¹⁵⁷⁾ 10 24-hour open and closed induction patches 24 hours apart. 2-3 week rest. 48-hour challenge patches. Open and closed patches at inductions 1,4,7, and 10 and at the challenge were irradiated for 1 minute at a distance of 12 in with a Hanovia Tanette Mark I lamp (UV). Scored at patch removal (1+ to 3+) | 48 | Five 1+ and one 2+ reactions to open patches at induction. One reaction to an open patch at UV challenge. "Nonirritating, nonsensitizing, nonphotosensitizing" | 141 |
| Nail prepara- tion, 6 percent DBP | 100 | Controlled use study for 4 weeks. Fingernails and eyes were examined each week (1+ to 3+) | 47 | No positive reactions were observed. "Nonirritating" | 142 |

DBP and DMP are degraded by renal homogenates from rats. Both of these phthalates and DEP are hydrolyzed by rat, ferret, and baboon liver and intestinal-mucosal cell homogenates. The phthalates are also metabolized by human liver homogenates and rat intestinal contents. The enzymes involved in DMP metabolism by rat intestinal contents are labile *in vitro* and mammalian in origin. DMP is hydrolyzed by the gastric contents of rats. Human feces are relatively inactive in degrading the phthalates.

Phthalates are ubiquitous in the environment, and human exposure is likely. DBP has been found in the kidneys, adipose tissue, blood, and umbilical cords of humans.

The acute oral LD₅₀ value of DBP for rats ranged from approximately 8 g/kg to 23.0 g/kg; DBP was practically nontoxic to relatively harmless. DMP was practically nontoxic; it had an acute oral LD₅₀ value for rats of 6.9 ml/kg. The LD₅₀ for rabbits for DEP administered orally was 1.0 g/kg. No adverse effects were reported after the oral administration of doses of DBP of 2.5 mg/kg per day for 6 months or 1 mg/kg two times a week for 1½ years to rats, or of a DBP concentration of 0.25 percent in the feed of rats for 1 year. At doses of 20 mg/kg of DBP for 80 days, growth was inhibited in mice and leukocytosis was observed in rats. At a concentration of 1.25 percent DBP in the diet, 5 of 10 rats died within a week, but the remaining rats survived the diet for a year and appeared normal. A 2.0 percent dietary concentration of DMP fed for 2 years to rats had no effect on growth, 4.0 and 8.0 percent inhibited growth, and rats fed 8.0 percent had chronic nephritis. Doses of DEP of 3 ml/kg per day for 8 days and a 0.2 percent concentration of DEP in the diet for up to 16 weeks had no adverse health effects in rabbits and rats, respectively. Concentrations of 1.0 and 5.0 percent DEP in the diet for up to 16 weeks reduced the growth of rats.

The oral administration of DBP produced testicular atrophy in the rat, mouse, guinea pig, and ferret but not in the hamster in a dose of 2.0 g/kg per day for 10 days. The simultaneous administration of zinc provided substantial protection against testicular damage in the rat. Testicular atrophy was not observed after the administration of DMP and DEP in oral doses of 7.2 mmole/kg per day (approximately 1.4 g/kg per day DMP and 1.6 g/kg per day DEP) to rats.

The acute dermal LD₅₀ for DMP for rabbits was greater than 10 ml/kg. The subacute (90-day) dermal LD₅₀s of DBP and DMP to rabbits were greater than 4 ml/kg per day. At doses of 0.5 to 4.0 ml/kg per day, DBP was slightly irritating to skin, DMP was irritating in molting areas only, and there was no evidence of sensitization by either of the phthalates. Renal damage was observed in rabbits that died during this study, and survivors receiving 2.0 to 4.0 ml/kg per day of DBP and DMP had varying degrees of nephritis. Phthalate emulsions were injected intradermally into rabbits; DBP produced a mild to moderate inflammatory response, and DMP and DEP produced a marked inflammatory response. The results of other such experiments varied with the vehicle used.

Undiluted DBP, DMP, and DEP were instilled into the eyes of rabbits; irritation was minimal. However, with long contact time, DMP may be irritating to the rabbit eye.

The inhalation and intraperitoneal, intravenous, intramuscular, and subcutaneous administration of the phthalates have been studied in a variety of laboratory animals. Results depended on the route, the species, and the dose.

Several studies suggest that the administration of DBP, DMP, and DEP to

pregnant rats may increase the number of resorptions and have significant effects upon embryonic and fetal development. Gross and skeletal abnormalities in offspring have been observed in some cases. The dietary administration to three generations of mice of doses up to 100 mg/kg per day DBP has increased the formation of renal cysts in the second and third generations.

The mutagenic activity of DBP, DMP, and DEP toward *S. typhimurium* mutants depends on the assay protocol; studies have been both negative and positive in the same strains. DBP and DEP were not mutagenic for *E. coli*. DNA repair enzyme-deficient *B. subtilis* and *E. coli* were not more sensitive to DBP and DEP than the wild-type bacteria; one study reported that DBP did not affect either the DNA repair enzyme-deficient *E. coli* or the wild-type. DBP was negative in a *S. cerevisiae* reversion analysis with and without metabolic activation. DBP was both negative and positive for chromosome aberrations and sister chromatid exchanges in Chinese hamster cells; it has been called a "suspicious compound." DBP, DMP, and DEP have no effect on chromosome aberrations in human leukocyte cultures.

DBP was not carcinogenic in chronic (18-month or longer) feeding studies in rats.

There were no positive reactions among 53 human subjects patch tested with 5 percent DBP. One positive reaction was observed when 1532 subjects were patch tested with DBP, DMP, and DEP at a 2 percent concentration. Cosmetic formulations containing up to 9 percent DBP were tested in a variety of patch test procedures; in some procedures some of the formulations were slightly irritating. In other cases, no irritation was observed. Sensitization and photosensitization were not observed.

DISCUSSION

A comparison of the chemical structures of the phthalates suggests that DBP may have the greatest toxicological significance. Data are limited for both DMP and DEP, and, in particular, there are clinical phototoxicity and photosensitivity data only for a preparation containing DBP. However, the Panel believes that the information contained in this report is adequate for a safety assessment of all three phthalates.

DBP but not DMP and DEP caused testicular injury in laboratory animals. The combined teratogenic test data available to the Expert Panel are not adequate to conclude that DBP, DMP, or DEP are proven teratogens. The concentrations used in cosmetic products and the rapid metabolism and elimination of these ingredients, as indicated by experimental studies, minimize the significance of the observations of testicular damage by DBP and the conflicting teratogenic test results. The Panel notes that the information provided in the literature on the carcinogenicity of DBP is limited and does not permit an evaluation of the assays performed and the results obtained. The results of mutagenesis studies, however, are essentially negative.

CONCLUSION

On the basis of the available data, the Panel concludes that Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate are safe for topical application in the present practices of use and concentration in cosmetics.

ACKNOWLEDGMENT

Karen Brandt, Scientific Analyst and Writer, prepared the literature review and technical analysis used by the Expert Panel in developing this report.

REFERENCES

1. ESTRIN, N.F. (ed.). (1977). *CTFA Cosmetic Ingredient Dictionary*, 2nd ed. Washington, DC: Cosmetic, Toiletry and Fragrance Assoc.
2. CTFA. (October 19, 1981). Submission of data by CTFA (2-14-7). Cosmetic ingredient chemical description of dibutyl phthalate.*
3. CTFA. (October 19, 1981). Submission of data by CTFA (2-14-6). Cosmetic ingredient chemical description of dimethyl phthalate.*
4. CTFA. (October 19, 1981). Submission of data by CTFA (2-14-5). Cosmetic ingredient chemical description of diethyl phthalate.*
5. GREENBERG, L.A., and LESTER, D. (1954). *Handbook of Cosmetic Materials*. New York: Interscience Publishers.
6. HAWLEY, G.G. (ed.). (1971). *The Condensed Chemical Dictionary*, 8th ed. New York: Van Nostrand Reinhold Co.
7. WINDHOLZ, M. (ed.). (1976). *The Merck Index*, 9th ed. Rahway, NJ: Merck & Co.
8. ITO, M., MIYAGAWA, T., ASAYAMA, M., and OGAWA, T. (1979). Solubility of organic compounds used for household articles in human sweat. *Aichi Ika Diagaku Igakkai Zasshi* **7**, 183-9.
9. WOLFE, N.L., STEEN, W.C., and BURNS, L.A. (1980). Phthalate ester hydrolysis linear free energy relationships. *Chemosphere* **9**, 403-8.
10. SAIDO, K., WATABE, T., MOTOHASHI, S., KUROKI, T., IKEMURA, T., SATOMI, M., and KIRISAWA, M. (1981). Thermal decomposition of phthalic acid esters. 5. Thermal decomposition of di-*n*-butylphthalate and analysis of its decomposed products. *Eisei Kagaku* **27**, 16-22.
11. HAGEN, E. (1966). Analytical investigations of plasticizer pyrolysis. *Plaste Kautschuk* **13**, 140-2.
12. TEZUKA, M., OKADA, S., and TAMEMASA, O. (1978). Radiolytic decontamination of di-*n*-butyl phthalate from water. *Radioisotopes* **27**, 306-10.
13. AUTIAN, J. (1973). Toxicity and health threats of phthalate esters: review of the literature. *Environ. Health Perspec.* **4**, 3-26.
14. ESTRIN, N.F. (ed.). (1974). *CTFA standards: Cosmetic ingredient descriptions*. Washington, DC: Cosmetic, Toiletry and Fragrance Assoc.
15. JAPAN COSMETIC INDUSTRY ASSOCIATION (JCIA). (1967; trans. in 1979). *Japanese Standards of Cosmetic Ingredients*. Tokyo, Japan: Yakuji Nippo, Ltd.
16. PERSIANI, C., and CUKOR, P. (1975). Liquid chromatographic method for the determination of phthalate esters. *J. Chromatogr.* **109**, 413-7.
17. SENZEL, A.J. (1977). *Newburger's Manual of Cosmetic Analysis*, 2nd ed. Washington, DC: Assoc. of Official Analytical Chemists.
18. DRUYAN, E.A. (1972). Spectrophotometric determination of tricresylphosphate and dibutylphthalate jointly present in air. *Gig. Sanit.* **37**, 100-2.
19. RAPAPORT, L.I., GORTSEVA, L.V., and ANTONYUK, O.K. (1972). Determination of phthalate esters liberated from polyvinyl chloride materials into the air. *Gig. Sanit.* **37**, 60-1.
20. LEHMANN, W. (1980). Detection of ketonox in gastric contents after acute poisoning. *Bromatol. Chem. Toksykol.* **13**, 433-4.
21. ST. JOHN, G.A., McREYNOLDS, J.H., and ANBAR, M. (1975, 1976). Determination of the concentration of explosives in air by isotope dilution. *Edgewood Arsenal Spec. Publ. (US Dept. Army): ISS EO-SP-76001, Proc. Annu. Symp. Trace Anal. Detect Environ.*, **6**, 180-95.
22. DYAKOVA, G.A. (1970). Separate determination of dibutyl-, dioctyl-, and didodecylphthalate in air by means of thin-layer chromatography. *Gig. Tr. Prof. Zabol.* **14**, 45-7.
23. NELSON, D.A. (1969). TLC detection of phthalate esters with anisaldehyde and Dragendorff reagents. *Anal. Biochem.* **29**, 171-3.

*Available upon request: Administrator, Cosmetic Ingredient Review, 1110 Vermont Ave., NW, Suite 810, Washington, DC 20005.

24. McADAMS, M.J., and VESTAL, M.L. (1980). Analysis of priority pollutants by crossed beam liquid chromatography mass spectrometry. *J. Chromatogr. Sci.* **18**, 110-1.
25. SCHOUTEN, M.J., PEERBOOM, J.W.C., and BRINKMAN, U.A.T. (1979). Liquid chromatographic analysis of phthalate esters in Dutch river water. *Int. J. Environ. Anal. Chem.* **7**, 13-24.
26. SCHWARTZ, H.E., ANZION, C.J.M., VAN VLIET, H.P.M., COPIUS PEERBOOMS, J.W., and BRINKMAN, U.A.T. (1979). Analysis of phthalate esters in sediments from Dutch rivers by means of high performance liquid chromatography. *Int. J. Environ. Anal. Chem.* **6**, 133-44.
27. RAMSEY, J.D., LEE, T.D., OSSELTON, M.D., and MOFFAT, A.C. (1980). Gas-liquid chromatographic retention indexes of 296 nondrug substances on SE-30 or OV-1 likely to be encountered in toxicological analyses. *J. Chromatogr.* **184**, 185-206.
28. TAKESHITA, R., TAKABATAKE, E., MINAGAWA, K., and TAKIZAWA, Y. (1977). Micro-determination of total phthalate esters in biological samples by gas-liquid chromatography. *J. Chromatogr.* **133**, 303-10.
29. AUSTERN, B.M., DOBBS, R.A., and COHEN, J.M. (1975). Gas-chromatographic determination of selected organic compounds added to waste water. *Environ. Sci. Technol.* **9**, 588-90.
30. TIROSH, N., SIEGMANN, A., and NARKIS, M. (1974). Pesticides: recovery from plasticized PVC and analysis by GLC. *J. Chromatogr. Sci.* **12**, 149-51.
31. BOUWER, E.J., McCARTY, P.L., and LANCE, J.C. (1981). Trace organic behavior in soil columns during rapid infiltration of secondary wastewater. *Water Res.* **15**, 151-9.
32. SHINOHARA, R., KIDO, A., ETO, S., HORI, T., KOGA, M., and AKIYAMA, T. (1981). Identification and determination of trace organic substances in tap water by computerized gas chromatography-mass spectrometry and mass fragmentography. *Water Res.* **15**, 535-42.
33. REZAEV, N.I., MAKOLKIN, I.A., and DAVYDOVA, Z.H.V. (1975). Vibration spectra of biologically active compounds. *Mosc. Univ. Biol. Sci. Bull.* **30**, 20-3.
34. NEERGAARD, J., NIELSEN, B., FAURBY, V., CHRISTENSEN, D.H., and NIELSEN, O.F. (1975). On the exudation of plasticizers from PVC haemodialysis tubings. *Nephron* **14**, 263-74.
35. WEAST, R.C. (ed.). (1978). *CRC Handbook of Chemistry and Physics*, 5th ed. West Palm Beach, FL: CRC Press.
36. MORI, S. (1976). Identification and determination of phthalate esters in river water by high-performance liquid chromatography. *J. Chromatogr.* **129**, 53-60.
37. BALSAM, M.S., and SAGARIN, E. (1972). *Cosmetics: Science and Technology*, 2nd ed. New York: Wiley & Sons, Vol. 2.
38. CODE OF FEDERAL REGULATIONS (CFR). (1981). Title 21, Parts 175.105, 175.300, 175.320, 176.170, 177.1010, 177.1200, 177.2420, 177.2600, 178.3910, 181.22, 181.27, 720.4. Washington, DC: US Government Printing Office.
39. FOOD AND DRUG ADMINISTRATION (FDA). (December 22, 1981). Cosmetic product formulation data: (a) Ingredients used in each product category and (b) Number of brand name products in each product code. Two computer printouts. Washington, DC.
40. ENGELHARDT, G., WALLNOEFER, P.R., and HUTZINGER, O. (1975). Microbial metabolism of dibutyl phthalate and related dialkyl phthalates. *Bull. Environ. Contam. Toxicol.* **13**, 342-7.
41. PEREZ, J.A., DOWNS, J.E., and BROWN, P.J. (1976). The effects of dimethyl phthalate on the growth of *Pseudomonas aeruginosa*. *Bull. Environ. Contam. Toxicol.* **16**, 486-90.
42. BRINGMANN, G., and KUEHN, R. (1980). Comparison of the toxicity thresholds of water pollutants to bacteria, algae, and protozoa in the cell multiplication test. *Water Res.* **14**, 231-41.
43. MORRIS, J.A., KHETTRY, A., and SEITZ, E.W. (1979). Antimicrobial activity of aroma chemicals and essential oils. *J. Am. Oil Chem. Soc.* **56**, 595-603.
44. LAWRENCE, W.H. (1978). Phthalate esters: The question of safety. *Clin. Toxicol.* **13**, 89-139.
45. RUIZ, R.G., and BROWN, P.J. (April 7-8, 1976). The effects of dimethyl phthalate on the yields of *Microcystis aeruginosa*. Presented at the Minority Biomedical support symposium, New Orleans, LA.
46. BRINGMANN, G., and KUEHN, R. (1978). Testing of substances for their toxicity threshold: model organisms *Microcystis (Diplocystis) aeruginosa* and *Scenedesmus quadricauda*. *Mitt. Internat. Verein. Limnol.* **21**, 275-84.
47. SHAHIN, M.M., and VON BORSTEL, R.C. (1977). Mutagenic and lethal effects of α -benzene hexachloride, dibutyl phthalate and trichloroethylene in *Saccharomyces cerevisiae*. *Mutat. Res.* **48**, 173-80.
48. YOSHIZAWA, T., TERAURA, M., and MOROOKA, N. (1977). Inhibitory effect of phthalic acid esters on multiplication of *Tetrahymena pyriformis* strain W protozoa. *Kagawa Daigaku Nogakubu Gakujutsu Hokoku* **28**, 149-55.
49. GUTIERREZ, Y., BARRERA, E., and WENIGER, D. (April 7-8, 1976). Effects of dimethyl phthalate and di-2-ethylhexyl phthalate on the growth rate of the protozoan, *Tetrahymena pyriformis*, presented at the minority Biomedical Support Symposium, New Orleans, LA.

50. DILLINGHAM, E.O., and AUTIAN, J. (1973). Teratogenicity, mutagenicity and cellular toxicity of phthalate esters. *Environ. Health Perspec.* **3**, 81-89.
51. CALLEY, D., AUTIAN, J., and GUESS, W.L. (1966). Toxicology of a series of phthalate esters. *J. Pharm. Sci.* **55**, 158-62.
52. GUESS, W.L., and HABERMAN, S. (1968). Toxicity profiles of vinyl and polyolefinic plastics and their additives. *J. Biomed. Mater. Res.* **2**, 313-35.
53. NEMATOLLAHI, J., GUESS, W.L., and AUTIAN, J. (1967). Plasticizers in medical application. I. Analysis and toxicity evaluation of dialkyl benzenedicarboxylates. *J. Pharm. Sci.* **56**, 1146-53.
54. OSER, Z., ABODEELY, R.A., and MCGUNNIGLE, R.G. (1977). Evaluation of elastomers for biomedical applications utilizing in vivo and in vitro model systems-state of the art and future trends. *Int. J. Polymeric Mater.* **5**, 177-87.
55. KASUYA, M. (1976). Toxicity of phthalate esters to tissue cultures. *Proc. 6th Int. Congr. Pharmacol.* **6**, 215-24.
56. SUZUKI, Y., and MATSUMOTO, M. (1980). Accumulation of triacylglycerol in tissue culture cells derived from human embryonic lung (L-132 cells) on administration of di-n-butyl phthalate. *Japan J. Exp. Med.* **50**, 253-62.
57. THELESTAM, M., CURVALL, M., and ENZELL, C.R. (1980). Effect of tobacco smoke compounds on the plasma membrane of cultured human lung fibroblasts. *Toxicology* **15**, 203-17.
58. JONES, A.E., KAHN, R.H., GROVES, J.T., and NAPIER, E.A. (1975). Phthalate ester toxicity in human cell cultures. *Toxicol. Appl. Pharmacol.* **31**, 283-9.
59. EKWALL, B., NORDENSTEN, C., and ALBANUS, L. (1982). Toxicity of 29 plasticizers to HeLa cells in the MIT-24 system. *Toxicology* **24**, 199-210.
60. SETH, P.K., AGARWAL, D.K., and AGARWAL, S. (1981). Effects of phthalic acid esters on drug metabolizing enzymes. *Bull. Environ. Contam. Toxicol.* **26**, 764-8.
61. SETH, P.K. (1982). Hepatic effects of phthalate esters. *Environ. Health Perspect.* **45**, 27-34.
62. TANAKA, A., MATSUMOTO, A., and YAMAHARA, T. (1978). Biochemical studies of phthalate esters. III. Metabolism of dibutyl phthalate (DBP) in animals. *Toxicology* **9**, 109-23.
63. AITIO, A., and PARKKI, M. (1978). Effect of phthalate esters on drug metabolizing enzyme activities in rat liver. *Arch. Int. Pharmacodyn. Ther.* **235**, 187-95.
64. INOUE, B., OGINO, Y., ISHIDA, T., OGATA, M., and UTSUMI, K. (1978). Effects of phthalate esters on mitochondrial oxidative phosphorylation in the rat. *Toxicol. Appl. Pharmacol.* **43**, 189-98.
65. MELNICK, R.L., and SCHILLER, C.M. (1982). Mitochondrial toxicity of phthalate esters. *Environ. Health Perspect.* **45**, 51-6.
66. OHYAMA, T. (1976). Effects of phthalate esters on the respiration of rat liver mitochondria. *J. Biochem.* **79**, 153-8.
67. SRIVASTAVA, S.P., MUSHTAQ, M., AGARWAL, D.K., and SETH, P.K. (1977). Effect of some phthalate plasticizers on rat liver. *Environ. Pollut. Human Health Proc. Int. Symp.* **1**, 465-75.
68. OHYAMA, T. (1977). Effects of phthalate esters on the latent ATPase and swelling of mitochondria. *J. Biochem.* **82**, 9-16.
69. BELL, F.P., PATT, C.S., BRUNDAGE, B., GILLIES, P.J., and PHILLIPS, W.A. (1978). Studies on lipid biosynthesis and cholesterol content of liver and serum lipoproteins in rats fed various phthalate esters. *Lipids* **13**, 66-74.
70. HALL, I.H., CHAPMAN, J.M., and COCOLAS, G.H. (1981). Effects of imide analogs on enzymes required for cholesterol and fatty acid synthesis. *J. Pharm. Sci.* **70**, 326-8.
71. LAGENTE, M., De La FARGE, F., and VALDIGUIE, P. (1979). The influence of phthalate esters on human plasma lecithin/cholesterol acyltransferase. *Lipids* **14**, 533-4.
72. MIYAHARA, C., KOGI, K., SATOH, S., SHIMIZU, H., KISHI, M., TANI, T., ARAI, K., HORIGUCHI, Y., and WADA, Y. (1973). Toxicity of phthalate plasticizers. *Kanagawa-Ken Kogyo Shikensho Kenkyu Hokoku* **3**, 19-34.
73. NISHIYAMA, K., SUZUKI, Y., KAWANO, M., and OKAMOTO, F. (1977). Effect of phthalic acid esters on biological systems. Effects on rats. *Shikoku Eisei Gakkai Zasshi* **22**, 19-21.
74. SAJIKI, J. (1978). Lactate dehydrogenase in serum and viscera of the mouse after administration of dibutyl phthalate. *Igaku No Ayumi* **104**, 463-5.
75. GLEIBERMAN, S.E., KOTOVA, I.A., NIKOLAEV, G.M., and YURCHENKO, V.V. (1978). Pharmacodynamics of dimethylphthalate. *Med. Parazit. Parazit. Bolezni* **47**, 58-63.
76. KOZUMBO, W.J., KROLL, R., and RUBIN, R.J. (1982). Assessment of the mutagenicity of phthalate esters. *Environ. Health Perspect.* **45**, 103-9.
77. ALBRO, P.W., and MOORE, B. (1974). Identification of the metabolites of simple phthalate diesters in rat urine. *J. Chromatogr.* **94**, 209-18.

78. YAMAGUCHI, T., KANESHIMA, H., OKUI, T., OGAWA, H., and YAMAGISHI, T. (1976). Effects of phthalate esters on the biological system. Part 1. Distribution of 14C-dibutylphthalate in mice. *Nippon Eiseigaku Zasshi* **32**, 331-6.
79. OKADA, S., and TAMEMASA, O. (1978). Distribution of metabolism of di-(n-butyl)-phthalate in mice and its interaction with nucleic acids and proteins. *Yakugaku Zasshi* **98**, 1229-35.
80. KANESHIMA, H., YAMAGUCHI, T., OKUI, T., and NAITOH, M. (1978). Studies on the effects of phthalate esters on the biological system. Part 2. In vitro metabolism and biliary excretion of phthalate esters in rats. *Bull. Environ. Contam. Toxicol.* **19**, 502-9.
81. LAKE, B.G., PHILLIPS, J.C., HODGSON, R.A., SEVERN, B.J., GANGOLLI, S.D., and LLOYD, A.G. (1976). Studies on the hydrolysis in vitro of phthalate esters by hepatic and intestinal mucosal preparations from various species. *Biochem. Soc. Trans.* **4**, 654-5.
82. LAKE, B.G., PHILLIPS, J.C., LINNELL, J.C., and GANGOLLI, S.D. (1977). The in vitro hydrolysis of some phthalate diesters by hepatic and intestinal preparations from various species. *Toxicol. Appl. Pharmacol.* **39**, 239-48.
83. ROWLAND, I.R., COTTRELL, R.C., and PHILLIPS, J.C. (1977). Hydrolysis of phthalate esters by the gastrointestinal contents of the rat. *Food Cosmet. Toxicol.* **15**, 17-21.
84. SMITH, C.C. (1953). Toxicity of butyl stearate, dibutyl sebacate, dibutyl phthalate, and methoxyethyl oleate. *Arch. Ind. Hyg. Occup. Med.* **7**, 310-8.
85. WHITE, P.D., CARTER, D.E., EARNEST, D., and MUELLER, J. (1980). Absorption and metabolism of three phthalate diesters by the rat small intestine. *Food Cosmet. Toxicol.* **18**, 383-6.
86. SINGH, A.R., LAWRENCE, W.H., and AUTIAN, J. (1975). Maternal-fetal transfer of 14C-di-2-ethylhexyl phthalate and 14C-diethyl phthalate in rats. *J. Pharm. Sci.* **64**, 1347-50.
87. OVERTURF, M.L., DRUILHET, R.E., KIRKENDALL, W.M., LIEHR, J.G., and CAPRIOLI, R.M. (1978). Occurrence of phthalate esters in human kidney. *Clin. Res.* **26**, 781A.
88. OVERTURF, M.L., DRUILHET, R.E., LIEHR, J.G., KIRKENDALL, W.M., and CAPRIOLI, R.M. (1979). Phthalate esters in normal and pathological human kidneys. *Bull. Environ. Contam. Toxicol.* **22**, 536-42.
89. MES, J., COFFIN, D.E., and CAMPBELL, D.S. (1974). Dibutyl and di(2-ethylhexyl) phthalate in human adipose tissue. *Bull. Environ. Contam. Toxicol.* **12**, 721-5.
90. YAGI, Y., NAKAMURA, Y., and TOMITA, I. (1977). Embryotoxicity of phthalate esters in mouse. *Int. Congr. Toxicol.* **33**, 590.
91. DRAIZE, J., ALVAREZ, E., WHITESELL, M.F., WOODWARD, G., HAGAN, E., and NELSON, A.A. (1948). Toxicological investigations of compounds proposed for use as insect repellants. *J. Pharmacol. Exp. Ther.* **93**, 26-39.
92. GESLER, R.M. (1973). Toxicology of di-2-ethylhexyl phthalate and other phthalic acid ester plasticizers. *Environ. Health Perspec.* **3**, 73-9.
93. HOMROWSKI, S., and NIKONOROW, M. (1959). Toksycznosc ostra ftalanu dwubutylu oraz ftalanu dwu-2-etyloheksylu produkceji krwjowej. *Rocz. Panstw. Zakl. Hig.* **10**, 321-7.
94. KRAUSKOPF, L.G. (1973). Studies on the toxicity of phthalates via ingestion. *Environ. Health Perspec.* **3**, 61-72.
95. LEHMAN, A.J. (1955). Insect repellants. *Assoc. Food Drug Offic. US Quart Bull.* **19**, 87-99.
96. NIKONOROW, M., MAZUR, H., and PIEKACZ, H. (1973). Effect of orally administered plasticizers and polyvinyl chloride stabilizers in the rat. *Toxicol. Appl. Pharmacol.* **26**, 253-9.
97. RADEVA, M., and DINOVA, S. (1966). Toxicity of dibutyl phthalate by oral application in albino rats. *Khig. Zdraveopazvane* **9**, 510.
98. OMORI, Y. (1976). Recent progress in safety evaluation studies on plasticizers and plastics and their controlled use in Japan. *Environ. Health Perspect.* **17**, 203-9.
99. SAJIKI, J., FUKUDA, Y., and YAMAGIWA, J. (1979). Pathological observations of phthalate ester (DBP) intoxicated mouse. *Kankyo Osen Busshitsu to Sono Tokishikorojii Shinpojumu* **6**, 54-6.
100. YAMADA, A., et al. (1975). Toxicity studies on plasticizers. 1. Subacute toxicity of di-(2-ethylhexyl) phthalate. *Trans. Food Hyg. Soc. Japan*, 29th meeting, p. 36.
101. CLAYTON, G.D., and CLAYTON, F.E. (eds.). (1981). *Patty's Industrial Hygiene and Toxicology*, 3rd ed. New York: Interscience, Vol. 2.
102. HODGE, H.C., and STERNER, J.H. (1949). Tabulation of toxicity classes. *Am. Indus. Hyg. Assoc. Quart.* **10**, 93-6.
103. KOMAROVA, E.N. (1979). Materials on the toxicology of dibutyl phthalate, dioctyl phthalate, dibutyl sebacate, and butyl stearate. *Khim. Prom-st. Ser.: Taksikol. Sanit. Khim. Plastmass* **3**, 12.
104. BIODYNAMICS. (April 20, 1976). Submission of data by CTFA (2-14-1). Project No. 3089-76. Acute oral toxicity in rats, compound no. WNP-2-113, composite no. 3.*
105. CTFA. (1977). Submission of data by CTFA (2-14-13). CIR safety data test summary. Acute oral toxicity and

- ocular irritation with a nail preparation containing 6 percent dibutyl phthalate.*
106. CATER, B.R., COOK, M.W., GANGOLLI, S.D., and GRASSO, P. (1977). Studies in dibutyl phthalate-induced testicular atrophy in the rat: Effect on zinc metabolism. *Toxicol. Appl. Pharmacol.* **41**, 609-18.
 107. FOSTER, P., THOMAS, L.V., COOK, M.W., and GANGOLLI, S.D. (1980). Testicular effects and changes in zinc excretion produced by some n-alkyl phthalates in the rat. *Toxicol. Appl. Pharmacol.* **54**, 392-8.
 108. GANGOLLI, S.D. (1982). Testicular effects of phthalate esters. *Environ. Health Perspect.* **45**, 77-84.
 109. BORNMANN, G., LOESER, A., MIKULICZ, K., and RITTER, K. (1956). Behavior of the organism as influenced by various plasticizers. *Z. Lebensm. Unters. Forsch.* **103**, 413-24.
 110. BROWN, D., BUTTERWORTH, K.R., GAUNT, I.F., GRASSO, P., and GANGOLLI, S.D. (1978). Short-term oral toxicity study of diethyl phthalate in the rat. *Food Cosmet. Toxicol.* **16**, 415-22.
 111. DANIEL, J.W. (1978). Toxicity and metabolism of phthalate esters. *Clin. Toxicol.* **13**, 257-68.
 112. MASLENKO, A.A. (1968). Hazards of dibutyl phthalate and dioctyl phthalate as reservoir pollutants. *Gig. Sanit.* **33**, 102.
 113. SHIBKO, S.I., and BLUMENTHAL, H. (1973). Toxicology of phthalic-acid esters used in food packaging material. *Environ. Health Perspec.* **3**, 131-7.
 114. BLICKENDORFER, P., and TEMPLETON, L. (1930). A study of the toxic properties of diethyl phthalate. *J. Am. Pharm. Assoc. Sci. Ed.* **19**, 1179-81.
 115. SRIVASTAVA, R.K., DESHMUKH, M.B., and SHARMA, S.K. (1978). A convenient method for evaluation of dermal irritant effect on rabbits. *Indian J. Pharmacol.* **10**, 319-22.
 116. LAWRENCE, W.H., MALIK, M., TURNER, J.E., SINGH, A.R., and AUTIAN, J. (1975). Toxicological investigation of some acute, short-term, and chronic effects of administering di-2-ethylhexyl phthalate (DEHP) and other phthalate esters. *Environ. Res.* **9**, 1-11.
 117. CARPENTER, C.P., and SMYTH, H.F. (1946). Chemical burns of the rabbit cornea. *Am. J. Ophthalmol.* **29**, 1363-72.
 118. DRAIZE, J.H., WOODARD, G., and CALVERY, H.O. (1944). Methods for the study of irritation and toxicity of substances applied topically to the skin and mucous membranes. *J. Pharmacol. Exp. Ther.* **82**, 377-90.
 119. BIODYNAMICS. (April 27, 1976). Submission of data by CTFA (2-14-2). Project no. 3093-76. Rabbit eye irritation, compound no. WNP-2-113, nail polish, composite No. 3.*
 120. KAWANO, M. (1980). Toxicological studies on phthalate esters. I. Inhalation effects of dibutyl phthalate on rats. *Japan J. Hyg.* **35**, 684-92.
 121. KAREL, L., LANDING, B.H., and HARVEY, T.S. (1947). The intraperitoneal toxicity of some glycols, glycol ethers, glycol esters, and phthalates in mice. *Fed. Proc.* **6**, 342.
 122. SINGH, A.R., LAWRENCE, W.H., and AUTIAN, J. (1972). Teratogenicity of phthalate esters in rats. *J. Pharm. Sci.* **61**, 51-5.
 123. PETERS, J.W., and COOK, R.M. (1973). Effect of phthalate esters on reproduction in rats. *Environ. Health Perspec.* **3**, 91-4.
 124. SHIOTA, K., and NISHIMURA, H. (1982). Teratogenicity of di(2-ethylhexyl)phthalate (DEHP) and di-n-butyl phthalate (DBP) in mice. *Environ. Health Perspect.* **45**, 65-70.
 125. ONDA, S., KODAMA, H., YAMADA, N., and OTA, H. (1974). Studies on the toxic effect of phthalate. III. Teratogenicity in mice. *Japan J. Hyg.* **29**, 177.
 126. AMES, B.N., McCANN, J., and YAMASAKI, M. (1975). Methods for detecting carcinogens and mutagens with the *Salmonella*/mammalian-microsome mutagenicity test. *Mutat. Res.* **31**, 347-64.
 127. FLORIN, I., RUTBERG, L., CURVALL, M., and ENZELL, C.R. (1980). Screening of tobacco smoke constituents for mutagenicity using the Ames test. *Toxicology* **15**, 219-32.
 128. KAWACHI, T., YAHAGI, T., KADA, T., TAZIMA, Y., ISIDATE, M., SASAKI, M., and SUGIYAMA, T. (1980). Cooperative program on short-term assays for carcinogenicity in Japan. *IARC Sci. Publ.* **27**, 323-30.
 129. NATIONAL TOXICOLOGY PROGRAM (NTP). (December 1980). Annual Plan for Fiscal Year 1981 (NTP-80-62).
 130. SEED, J.L. (1982). Mutagenic activity of phthalate esters in bacterial liquid suspension assays. *Environ. Health Perspect.* **45**, 111-4.
 131. RUBIN, R.J., KOZUMBO, W., and KROLL, R. (1979). Ames mutagenic assay of a series of phthalic acid esters: Positive response of the dimethyl and diethyl esters in TA100. *Toxicol. Appl. Pharmacol.* **48**, A133.
 132. SZYBALSKI, W. (1958). Special microbiological systems. 2. Observations on chemical mutagenesis in microorganisms. *Ann. NY Acad. Sci.* **76**, 475-89.
 133. ROSENKRANZ, H.S., and LEIFER, Z. (1980). Determining the DNA-modifying activity of chemicals using DNA-polymerase-deficient *Escherichia coli*. *Chem. Mutagens Prin. Methods Detection* **6**, 109-47.
 134. ABE, S., and SASAKI, M. (1977). Chromosome aberrations and sister chromatid exchanges in Chinese hamster cells exposed to various chemicals. *J. Natl. Cancer Inst.* **58**, 1635-41.

135. ISHIDATE, M., and ODASHIMA, S. (1977). Chromosome tests with 134 compounds on Chinese hamster cells in vitro—A screening for chemical carcinogens. *Mutat. Res.* **48**, 337–53.
136. TSUCHIYA, K., and HATTORI, K. (1976). Chromosomal study on human leukocyte cultures treated with phthalic acid ester. *Hokkaidoritsu Eisei Kenkyusho Ho* **26**, 114.
137. DERNEHL, C.U. (March 20–22, 1973). Toxicity of phthalates via ingestion. *Tech. Pap. Reg. Tech. Conf. Soc. Plast. Eng. Palisades Sect.* 71–2.
138. NTP. (March 1982). Technical report on the carcinogenesis bioassay of Di(2-ethylhexyl)phthalate (CAS No. 117-81-7) in F344 rats and B6C3F₁ mice (feed study) PB82184011.
139. KAABER, S., THULIN, H., and NIELSON, E. (1979). Skin sensitivity to denture base materials in the burning mouth syndrome. *Contact Dermatitis* **5**, 90–6.
140. SCHULSINGER, C., and MOLLGAARD, K. (1980). Polyvinyl chloride dermatitis not caused by phthalates. *Contact Dermatitis* **6**, 477–80.
141. COSMETIC, TOILETRY AND FRAGRANCE ASSOCIATION (CTFA). (1976). Submission of data by CTFA (2-14-14). CIR safety data test summary. Patch tests with a nail preparation containing 6 percent DBP.*
142. CTFA. (1977). Submission of data by CTFA (2-14-15). CIR safety data test summary. Controlled use study with a nail preparation containing 6 percent DBP.*
143. CTFA. (January–February 1979). Submission of data by CTFA (2-14-11). CIR safety data test summary. 21-day cumulative irritancy test with a deodorant containing 4.5 percent dibutyl phthalate.*
144. CTFA. (May–June 1980). Submission of data by CTFA (2-14-10). CIR safety data test summary. Draize repeated insult patch test with a deodorant containing 4.5 percent dibutyl phthalate.*
145. CTFA. (November–December 1980). Submission of data by CTFA (2-14-12). CIR safety data test summary. Draize repeated insult patch test with a deodorant containing 4.5 percent dibutyl phthalate.*
146. CTFA. (November 1980 and February 1981). Submission of data by CTFA (2-14-9). CIR safety data test summary. Use test with a deodorant containing 4.5 percent dibutyl phthalate.*
147. HILL TOP RESEARCH. (May 20, 1976). Submission of data by CTFA (2-14-3). The study of cumulative irritant properties of a series of test materials.*
148. IVY RESEARCH LABORATORIES. (June 2, 1976). Report on appraisal of sensitizing potential of four products by maximization testing, Ivy prot. no. 3294.*
149. CAGIANUT, V.B. (1954). Keratitis erosiva und nephritis toxica nach Einnahme von Dibutylphthalat. *Schweiz. Med. Wochenschr.* **84**, 1243–4.
150. McLAUGHLIN, R.S. (1946). Chemical burns of the human cornea. *Am. J. Ophthalmol.* **29**, 1355–62.
151. MILKOV, L.E., ALDYREVA, M.V., POPOVA, T.B., LOPUKHOVA, K.A., MAKARENKO, Y.L., MALYAR, L.M., and SHAKHOVA, T.K. (1973). Health status of workers exposed to phthalate plasticizers in the manufacture of artificial leather and films based on PVC resins. *Environ. Health Perspec.* **3**, 175–8.
152. KLIGMAN, A.M. (1966). The identification of contact allergens by human assay. III. The maximization test: a procedure for screening and rating contact sensitizers. *J. Invest. Dermatol.* **47**, 393–409.
153. DRAIZE, J.H. (1959). Dermal toxicity. In: *Appraisal of the Safety of Chemicals in Foods, Drugs and Cosmetics*. Austin, TX: Association of Food and Drug Officials of the United States, Texas State Department of Health, pp. 46–59.
154. MARZULLI, F.N., and MAIBACH, H.I. (eds.). (1977). *Advances in Modern Toxicology. Dermatotoxicology and Pharmacology*. New York: Hemisphere Publishing Corp., Vol. 4.
155. PHILLIPS, L., STEINBERG, M., MAIBACH, H.I., and AKERS, W.A. (1972). A comparison of rabbit and human skin response to certain irritants. *Toxicol. Appl. Pharmacol.* **21**, 369–82.
156. SCHWARTZ, L., and PECK, S.M. (1944). The patch test in contact dermatitis. *Public Health Rep.* **59**, 546–57.
157. SHELANSKI, H.A., and SHELANSKI, M.V. (1953). A new technique of human patch tests. *Proc. Joint Conf. Cosmet. Sci.* **19**, 46–9.

Werner, U., T. Kissel, and M. Reers. 1996. Effects of permeation enhancers on the transport of a peptidomimetic thrombin inhibitor (CRC 220) in a human intestinal cell line (Caco-2). *Pharm. Res.* 13:1219–1227.

DIBUTYL PHTHALATE, DIETHYL PHTHALATE, AND DIMETHYL PHTHALATE

A safety assessment of Dibutyl Phthalate (DBP), Diethyl Phthalate (DEP), and Dimethyl Phthalate (DMP) was published in 1985 with the conclusion that these ingredients “are safe for topical application in the present practices of use and concentrations in cosmetics” (Elder 1985). Since then many additional studies have appeared in the scientific literature. These studies, along with the updated information in Table 8 regarding uses and use concentrations, were considered by the CIR Expert Panel. Based on its consideration of the data discussed below, the Panel decided not to reopen this safety assessment.

DBP, DEP, and DMP are phthalate diesters that are used in cosmetics as plasticizers, solvents and fragrance ingredients in a wide variety of cosmetic product types. DEP is also used as a denaturant. DBP is found primarily in nail care products (at concentrations up to 15%) and in some hair care formulations (up to 0.1%). DEP is found in certain bath preparations, fragrance products, deodorants, lotions, and other skin care products. The highest reported concentration of use of DEP is 11% in perfumes. DMP is an ingredient in some hair care products, including aerosol fixatives. The reported maximum concentration of use of DMP in cosmetics is 2% in aerosol hair sprays. Table 8 provides the frequency and concentration of use as a function of product type.

Recent studies document that DBP, DEP, and DMP all absorb readily through the skin and through the gastrointestinal (GI) tract. Once absorbed, most short-chain phthalate diesters are hydrolyzed to the corresponding monoester and alcohol. The phthalates and their metabolites distribute to most tissues, and cross the placenta, but they do not accumulate in any specific tissue type. Phthalates are quickly eliminated in the urine, usually as the corresponding monoester or its glucuronide conjugate. However, humans and primates metabolize longer-chain diester phthalates (e.g., DEHP) into the glucuronide-conjugated monoester forms to a much larger extent than do rats. Also, rats excrete three to four times more free unconjugated MBP than do hamsters given similar doses of DBP or MBP, possibly due to greater testicular β -glucuronidase activity in rats than in hamsters. Phthalates undergo some enterohepatic cycling, and some phthalate is eliminated in the feces.

New data on acute and short-term toxicity were consistent with previously available data.

In a NTP study, DBP, DEP, and DMP were not found to be dermal irritants or sensitizers, confirming previous data using human and animal subjects.

Although previous data had identified that orally administered (in feed or by gavage) DBP and its metabolite MBP have re-

productive and developmental effects in rodents, with impaired male development being the most sensitive effect, newly available data provided additional demonstration of such effects.

When pregnant rats and mice were exposed to 1.0% DBP in powdered feed throughout gestation, the pregnancy outcome showed reductions in fertility, number of pups per litter, number of live pups, and body weights of pups. Adult male rats exposed to 1.0% DBP showed signs of liver and kidney toxicity and reduced weights of the prostate, testes, and seminal vesicles. Pregnant rats exposed to 2% DBP in feed throughout pregnancy had a higher incidence of preimplantation loss and resorptions, and no male pups were born alive. Exposure to 1% or 2% DBP in feed only during the latter half of gestation did not show the preimplantation loss and resorption rate seen in rats exposed throughout pregnancy. However, the increased survivability of these fetuses allowed the morphological defects of developing fetuses to be observed. These defects included reduced body weights in both sexes at 2% DBP, reduced anogenital distance and undescended testes in male fetuses at 1% and 2% DBP, and increased incidence of cleft palate and fused sternebrae. Adverse fetal effects were not seen in this study in a 0.5% DBP feed group, or at 331 mg/kg/day, based on average food consumption.

Oral intubation (gavage) of DBP in rats during gestation produced similar effects to those seen in the feeding studies described above. Pregnant rats given oral doses of approximately 0.63 to 0.75 g/kg/day and higher on certain gestation days produced litters with higher incidences of fetal toxicity and malformations. Exposure to DBP on gestation days 7 through 9 or on days 13 through 15 results in increased incidence of skeletal malformations such as cleft palate, fused sternebrae, and vertebral anomalies, as well as dilatation of the renal pelvis and undescended testes. However, exposure to DBP on gestation days 10 through 12 did not produce these effects, suggesting that DBP teratogenicity may be age dependent. Prenatal exposure to MBP appears to produce fetotoxicity and teratogenicity similar to DBP, following the same patterns of age-dependent sensitivity and dose efficacy. This supports the proposal that it is the monoester metabolite that produces the developmental toxicity of DBP and other phthalates.

DEP fed to mice at concentrations up to 2.5% (calculated to be 3.64 g/kg/day) in a continuous breeding protocol produced no effects of DEP on fertility or pregnancy outcome in the F₀ generation. F₁ male mice of the 2.5% DEP group had enlarged prostates and reduced sperm counts, but sperm motility and morphology were not affected. The F₂ generation showed no treatment-related differences between DEP and control groups. Pregnant rats fed up to 5.0% DEP mixed in feed on gestation days 6 through 15 produced no treatment-related alterations in fetal viability or development.

Repeated dermal application of 2 ml/kg up to 50% DEP to pregnant rabbits on gestation days 6 through 18 did not produce maternal or fetal toxicity or affect fetal development.

DMP was not fetotoxic or teratogenic when administered dermally (in rats) or orally (in rats and mice) during gestation.

TABLE 8
Historical and current cosmetic product uses and concentrations for Dibutyl, Diethyl, and Dimethyl Phthalate

| Product category | 1981 use (Elder 1985) | 2001 use (FDA 2001) | 1981 concentrations (FDA 1981) (%) | 2001 concentrations (CTFA 2001a, 2001b, 2001c) (%) |
|--|--------------------------|------------------------|--|--|
| <i>Dibutyl Phthalate</i> | | | | |
| Perfumes | — | — | — | 38–890 ppm** |
| Hair sprays | — | — | — | 55–160 ppm** |
| Shampoos (noncoloring) | — | — | — | 0.007 |
| Hair preparations (other noncoloring) | 3 | — | >0.1–1 | — |
| Hair-coloring preparations (other) | 3 | — | >0.1–1 | 0.1 |
| Aftershave lotions | 3 | — | >0.1–1 | — |
| Hair bleaches | — | — | — | 0.1 |
| Makeup (other) | 1 | — | >0.1–1 | 0.5 |
| Nail basecoats and undercoats | 36 | 32 | >1–10 | 1–6; 15* |
| Nail creams and lotions | — | 2 | — | 5 |
| Nail extenders | — | — | — | 1; 1* |
| Nail polish and enamel | 522 | 88 | ≤25 | 0.5–15; 15* |
| Nail polish and enamel removers | 3 | — | 0.1–25 | 2 |
| Nail care preparations (other) | 14 | 25 | ≤25 | 5–7; 6* |
| Underarm deodorants | — | — | — | 140–200 ppm** |
| Personal cleanliness products (other) | 5 | 3 | >1–5 | — |
| Total uses/ranges for Dibutyl Phthalate | 590 | 150 | 0.1–25 | 0.0038–15 |
| <i>Diethyl Phthalate</i> | | | | |
| Baby shampoos | — | — | — | 0.03 |
| Baby lotions, oils, powders, and creams | — | — | — | 0.00003 |
| Baby products (other) | — | — | — | 0.05 |
| Bath oils, tablets, and salts | 3 | 1 | ≤5 | — |
| Bubble baths | — | — | — | 0.06 |
| Bath preparations (other) | 2 | 2 | ≤0.1 | 0.008–0.09 |
| Colognes and toilet waters | 19 | 24 | ≤5 | 0.2–2 |
| Perfumes | 23 | 7 | ≤50 | 1–11 |
| Powders | 1 | 5 | >0.1–1 | — |
| Sachets | 3 | 2 | >0.01–5 | — |
| Other fragrance preparations | 2 | 11 | >0.1–50 | 0.01–1 |
| | | | | 67–28,000 ppm** |
| Hair conditioners | — | — | — | 0.1–0.2 |
| Hair sprays (aerosol fixatives) | 5 | — | >0.1–5 | 0.4 |
| | | | | 17–1500 ppm** |
| Shampoos (noncoloring) | — | — | — | 0.0008–0.2 |
| Hair tonics, dressings, etc. | — | 1 | — | 14–220 ppm** |
| Wave sets | 1 | — | >0.1–1 | — |
| Face powders | — | — | — | 0.4 |
| Eye shadow | 1 | — | ≤0.1 | — |
| Eyebrow pencil | — | — | — | 0.007 |
| Mascara | — | — | — | 0.007–0.07 |
| Eye makeup preparations (other) | — | — | — | 0.07 |
| Foundations | — | — | — | 0.3 |
| Makeup (other) | — | — | — | 0.0003 |
| Nail polish and enamel | — | — | — | 0.1 |
| Nail polish and enamel remover | 1 | — | >1–5 | — |

(Continued on next page)

TABLE 8

Historical and current cosmetic product uses and concentrations for Dibutyl, Diethyl, and Dimethyl Phthalate (*Continued*)

| Product category | 1981 use (Elder 1985) | 2001 use (FDA 2001) | 1981 concentrations (FDA 1981) (%) | 2001 concentrations (CTFA 2001a, 2001b, 2001c) (%) |
|--|--------------------------|------------------------|--|--|
| Nail care preparations (other) | — | — | — | 0.2 |
| Bath soaps and detergents | 1 | — | >0.1–1 | 2 |
| Underarm deodorants | — | 4 | — | 0.3–1 20–3300 ppm** |
| Feminine hygiene deodorants | — | — | — | 0.4 |
| Other personal cleanliness products | — | — | — | 1 |
| Aftershave lotion | 3 | 4 | >0.1–1 | 0.5–2 |
| Shaving cream (aerosol, brushless, and lather) | — | — | — | 0.001 |
| Other shaving preparation products | — | — | — | 1 |
| Skin-cleansing creams, lotions, liquids and pads | — | — | — | 0.0002 |
| Face and neck skin care preparations | 1*** | — | ≤0.1*** | 0.3 |
| Body and hand skin care preparations | — | 2 | — | 0.008–0.5 26–190 ppm** |
| Foot powders and sprays | — | — | — | 1 |
| Night skin care preparations | — | — | — | 0.0004 |
| Paste masks (mud packs) | — | 1 | — | 0.1 |
| Skin fresheners | — | 4 | — | 0.1–0.9 |
| Skin care preparations (other) | 1 | 5 | >0.1–1 | 0.00003–0.9 |
| Total uses/ranges for Diethyl Phthalate | 67 | 73 | ≤ 0.1–50 | 0.00003–2 |
| <i>Dimethyl Phthalate</i> | | | | |
| Hair conditioners | 2 | — | >0.1–1 | — |
| Hair sprays (aerosol fixatives) | — | 8 | — | 0.00002–2 |
| Hair rinses | 1 | — | >0.1–1 | — |
| Shampoos (noncoloring) | — | — | — | 0.00002 |
| Hair tonics, dressings, etc. | 2 | — | >0.1–5 | — |
| Wave sets | 2 | — | >0.1–1 | — |
| Hair preparations (other noncoloring) | 4 | 3 | >0.1–1 | — |
| Hair color sprays (aerosol) | — | 1 | — | — |
| Blushers | — | — | — | 0.00008 |
| Face powders | — | — | — | 0.00008 |
| Foundations | — | — | — | 0.005 |
| Bath soaps and detergents | — | — | — | 0.004 |
| Underarm deodorants | — | — | — | 33 ppm**–0.2 |
| Aftershave lotions | — | — | — | 0.2 |
| Total use/ranges for Dimethyl Phthalate | 11 | 12 | >0.1–5 | .00002–2 |

*Maximum concentrations reported by Nail Manufacturers Council (NMC 2001).

**Concentrations found in off-the-shelf products (Houlihan et al. 2002).

***These categories were combined when the original safety assessment was performed and are now separate categories.

Exposure to some phthalates has been shown to cause impairments of normal male development in rodents. The documented male-specific effects of phthalates include malformations of the epididymis and vas deferens, undescended testes, hypospadias, retention of thoracic nipples, and reduced anogenital distance. DEP and DMP did not cause the dramatic effects on male development seen with longer-chain dialkyl phthalates. Many studies

have reviewed the mechanisms of the male-targeted toxicity of phthalates. DBP, DEP, and DMP have weak or no binding affinity for the estrogen receptor and do not affect estrogen-regulated developmental endpoints. An antiandrogenic mechanism has been proposed, but many studies show that these phthalates do not bind with androgen receptors, either. However, phthalate esters inhibit the synthesis of testosterone, which is an important

hormone in normal development in males. DBP has also been shown to inhibit the action of Müllerian Inhibiting Substance produced by Sertoli cells.

DBP, DEP, and DMP previously had been screened for mutagenicity in the Ames bacterial reverse mutation assay with no mutagenic potential found. Additional data were available reporting that DBP caused an increase in the number of TA100 revertants in the absence but not in the presence of S9 rat liver fraction. DEP caused increases in the numbers of TA100 and TA1535 revertants, but this effect was also eliminated by the presence of S9. DMP caused an increase in the number of TA1535 revertants, but S9 prevented the effect. Overall, DBP, DEP, and DMP continue to have little genotoxic potential. One study on males of subfertile couples examined the relationship between environmental exposures to phthalates and DNA damage in human sperm using the neutral comet assay which is said to measure at least two aspects of DNA integrity. Neither the monobutyl form of DBP nor DMP had a significant association with comet assay parameters, and a significant association with the monobutyl form of DEP was seen only with one measure of DNA integrity.

Phthalates are a matter of concern for those responsible for public health and have been (and continue to be) reviewed by many government and international organizations. Phthalates are ubiquitous in the modern environment. The monoester metabolites of phthalates have been detected in the urine of an adult reference population and in the urine of young children in a small pilot study.

The Centers for Disease Control and Prevention (2003) found that the urinary concentrations of the monoester metabolites of DBP and DEP in 2536 Americans were similar or slightly lower than those reported in a preliminary study of 289 adults (Blount et al. 2000). Environmental exposure to phthalates and other endocrine disruptors have been proposed to be linked to an increased incidence of hypospadias in humans. The developmental effects of phthalates seen in rodents raise questions about the potential for human health risks. However, these effects seen in rodents are at much higher exposure levels than humans are likely to encounter, and they are subject to the species differences in the metabolism of phthalate diesters. The estimated median exposure levels of DEP and DBP are 57 $\mu\text{g/kg/day}$ and 7 $\mu\text{g/kg/day}$, respectively, while the U.S. EPA reference doses (RfD) for DEP and DBP are 800 $\mu\text{g/kg/day}$ and 100 $\mu\text{g/kg/day}$, respectively. Thus, the human exposure is well below the safety limits set by the U.S. EPA. Even the median exposure levels of the highest-exposed group (women aged 20 to 40 years) are well below the RfDs. Exposure levels were not available for DMP.

Scientific committees with the governments of the United States and the European Union have evaluated the human risks of DBP and DEP and expressed minimal to no concern over consumer exposure to these compounds (NTP Center for the Evaluation of Risks to Human Reproduction 2000; Netherlands Organization for Applied Scientific Research and National Institute of Public Health and the Environment 2000; Scientific Committee on Cosmetic Products and Non-Food Products 2002).

As in the original safety assessment of these phthalate diesters in 1984, the primary safety issue regarding phthalate esters in this re-review is antiandrogenic activity and the potential effects on male development. The CIR Expert Panel noted that the free monoester metabolite appears to be the active agent in phthalate diester toxicity. Of the three compounds reviewed in this safety assessment, Dibutyl Phthalate raised the most concern.

The Panel reviewed the numerous studies that describe the developmental toxicity of DBP in rodents. The Panel noted that the no observed adverse effect level (NOAEL) of DBP in a gavage study was 50 mg/kg/day (Mylchreest et al. 2000). However, a feeding study reported a NOAEL of 331 mg/kg/day (Ema et al. 1998). Overall, the Panel felt that feeding studies better represent the type of exposure that humans would receive from cosmetics than do gavage studies, but agreed that a worst-case NOAEL of 50 mg/kg/day should be considered.

The Panel considered a Margin of Safety (MOS) approach to assess the risk of DBP exposure to human users of cosmetics based on calculated exposures and the animal developmental toxicity data. Exposure calculations were based on ingredient concentration of use in cosmetic products (CTFA, 2001a, 2001b, 2001c; Houlihan et al. 2002), extent of cosmetic use survey data (Environ Corporation 1985; CTFA 2002b), and dermal (Mint et al. 1994) and subungual penetration data (Jackson Research Association 2002). A conservative approach to penetration was used; i.e., an estimate of approximately 5% absorption of DEP in human skin was considered to be a conservative estimate of DBP absorption, because data suggest that DEP is more readily absorbed in rat skin than DBP (Scott et al. 1987). The Panel used an estimated consumer body weight of 60 kg.

The expected exposure was calculated as follows:

Nail Basecoat or Polish

- 280 mg/application to 10 fingernails (Environ Corporation 1985)
- 15% maximum DBP in nail basecoats and polish (CTFA 2001a, 2001b, 2001c; Houlihan et al. 2002)
- 8.5% penetration through nail in 14 days (Jackson Research Association 2002)

$$280 \text{ mg/day} \times 15\% \times 8.5\%/14 \text{ days} = 0.255 \text{ mg/day/60 kg} = 4.25 \text{ } \mu\text{g/kg/day} \text{ and } 4.25 \text{ } \mu\text{g/kg/day} \times 2 \text{ (for fingers and toes)} = \underline{8.5 \text{ } \mu\text{g/kg/day}}.$$

Hair Spray

- 5 g/day hair spray use (CTFA, 2002)
- 160 $\mu\text{g/g}$ DBP in hair spray (Houlihan et al. 2002)
- 20% skin contact, from CTFA maximum worst case
- 5% skin absorption (Mint et al. 1994)

$$5 \text{ g/day} \times 160 \text{ } \mu\text{g/g} \times 20\% \times 5\% = 8 \text{ } \mu\text{g/day/60 kg} = \underline{0.14 \text{ } \mu\text{g/kg/day}}.$$

Deodorant

- 0.52 g/day deodorant use (Environ Corporation 1984)
- 200 µg/g DBP in deodorant (Houlihan et al. 2002)
- 5% skin absorption (Mint et al. 1994)

$$0.52 \text{ g/day} \times 200 \text{ µg/g} \times 5\% = 5.2 \text{ µg/day/60 kg} = \underline{0.09 \text{ µg/kg/day}}$$

Perfume

- 0.53 g/day perfume use (CTFA 2002)
- 890 µg/g DBP in perfume (Houlihan et al. 2002)
- 5% skin absorption (Mint et al. 1994)

$$0.53 \text{ g/day} \times 890 \text{ µg/g} \times 5\% = 24 \text{ µg/day/60 kg} = \underline{0.4 \text{ µg/kg/day}}$$

Total Exposure

- Sum of each of the separate exposures

$$8.5 \text{ µg/kg/day} + 0.14 \text{ µg/kg/day} + 0.09 \text{ µg/kg/day} + 0.4 \text{ µg/kg/day} = \underline{9.13 \text{ µg/kg/day}}$$

The calculated estimated exposure level of DBP from the concurrent use of multiple cosmetic products was 9.13 µg/kg/day. This value is within the reported range of total human exposure to DBP from all sources in women, 32 µg/kg/day (upper 95th percentile for women of reproductive age) to 6.5 µg/kg/day (upper 95th percentile for rest of group). Therefore, the Panel accepted 9.13 µg/kg/day as a not unreasonable approximation of DBP exposure from cosmetic products.

The Panel calculated the MOS of DBP by dividing the NOAEL of 331 mg/kg/day (from a feeding study) by the expected exposure of 9.13 µg/kg/day, yielding an MOS of 36,254. If the more conservative NOAEL of 50 mg/kg/day (from a gavage study) is used, the MOS is 5476. The Panel also noted that both NOAEL figures were obtained from rat studies, and detoxification metabolism of DBP is faster in humans than in rats.

The Panel acknowledged the use of DBP, DEP, and DMP in hair sprays. The effects of inhaled aerosols depend on the specific chemical species, the concentration, the duration of exposure, and site of deposition (Jensen and O'Brien 1993) within the respiratory system. Particle size is the most important factor affecting the location of deposition. The mean aerodynamic diameter of pump hair spray particles is approximately 80 µm, and the diameter of anhydrous hair spray particles is 60 to 80 µm. Typically less than 1% are below 10 µm which is the upper limit for respirable particles (Bowen 1999). Based on the particle size, DBP, DEP, and DMP would not be respirable in formulation. Therefore, exposure of the lung by inhalation was not considered likely.

Based on the available information included in this report, the CIR Expert Panel concluded that Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate are safe for use in cosmetic products in the present practices of use and concentrations.

REFERENCES

- Agarwal, D. K., W. H. Lawrence, L. J. Nunez, and J. Autian. 1985. Mutagenicity evaluation of phthalic acid esters and metabolites in Salmonella Typhimurium cultures. *J. Toxicology Environ. Health* 16:61–69.
- Agency for Toxic Substances and Disease Registry. 1995. Toxicological profile for diethyl phthalate. NTIS Report No. PB95264214.
- Albro, P. W. 1986. Absorption, metabolism, and excretion of di(2-ethylhexyl)phthalate by rats and mice. *Environ. Health Perspect.* 65:293–298.
- Anderson, W. A., L. Castle, M. J. Scotter, R. C. Massey, and C. Springall. 2001. A biomarker approach to measure human dietary exposure to certain phthalate diesters. *Food Addit. Contam.* 18:1068–1074.
- Api, A. M. 1997. *In vitro* assessment of phototoxicity. *In Vitro Toxicol.* 10:339–350.
- Api, A. M. 2001. Toxicological profile of diethyl phthalate: A vehicle for fragrance and cosmetic ingredients. *Food Chem. News.* 39:97–108.
- Bagley, D. M., J. R. Garner, G. Holland, R. W. Lewis, J.-F. Regnier, D. A. Stringer, and A. P. Walker. 1996. Skin irritation: Reference chemicals data bank. *Toxicol. In Vitro.* 10:1–6.
- Barber, E. D., M. Cifone, J. Rundell, R. Pryzgoda, B. D. Astill, E. Moran, A. Mulholland, E. Robinson, and B. Schneider. 2000. Results of the L5178Y mouse lymphoma assay and the Balb/3T3 cell *in vitro* transformation assay for eight phthalate esters. *J. Appl. Toxicol.* 20:69–80.
- Baskin, L. S., K. Himes, and T. Colborn. 2001. Hypospadias and endocrine disruption: Is there a connection? *Environ. Health Perspect.* 109:1175–1183.
- BASF. no date. Di-n-butyl phthalate—subacute inhalation study in Wistar rats. 20 exposures as a liquid aerosol. Unpublished data submitted by Phthalate Esters Panel, American Chemistry Council. 57 pages.⁹
- BASF. 1992. Study on the oral toxicity of dibutyl phthalate in Wistar rats. Administration via the diet over 3 months. Unpublished data submitted by Phthalate Esters Panel, American Chemistry Council. 369 pages.⁹
- BASF. 1999. Di-n-butyl phthalate—subacute inhalation study in Wistar rats. 20 exposures as a liquid aerosol. Unpublished data submitted by Phthalate Esters Panel, American Chemistry Council. 189 pages.⁹
- Blair, R. M., H. Fang, W. S. Branham, B. S. Hass, S. L. Dial, C. L. Moland, W. Tong, L. Shi, R. Perkins, and D. M. Sheehan. 2000. The estrogen receptor relative binding affinities of 188 natural and xenochemicals: Structural diversity of ligands. *Toxicol. Sci.* 45:138–153.
- Blount, B. C., M. J. Silva, S. P. Caudill, et al. 2000. Levels of seven urinary phthalate metabolites in a human reference population. *Environ. Health Perspect.* 108:979–982.
- Bower, D. 1999. Unpublished information on hair spray particle size provided at the September 9, 1999 CIR Expert Panel meeting.⁹
- Brock, J. W., S. P. Caudill, M. J. Silva, L. L. Needham, and E. D. Hilborn. 2002. Phthalate monoesters levels in the urine of young children. *Bull. Environ. Contam. Toxicol.* 68:309–314.
- Butala, J., M. Moore, M. Cifone, J. Bankston, and B. Astill. 1996. Oncogenicity study of di(isononyl) phthalate in rats. *Toxicologist.* 30:202.
- Butala, J., M. Moore, M. Cifone, J. Bankston, and B. Astill. 1997. Oncogenicity study of di(isononyl) phthalate in mice. *Toxicologist.* 36:173.
- Castle, L., A. Mercer, J. R. Startin, and J. Gilbert. 1988. Migration from plasticized films into foods 3. Migration of phthalate, sebacate, citrate and phosphate esters from films used for retail food packaging. *Food Addit. Contam.* 5:9–20.
- Centers for Disease Control and Prevention. 2003. Second national report on human exposure to environmental chemicals. Available at <http://www.cdc.gov/exposurereport/pdf/SecondNER.pdf>
- Chu-Tse, W., P. Xue-Tao, C. Jun-Rong, and X. Hui-Hua. 1993. A new pharmacological activity of dibutyl phthalate (dbp) on selective elimination of tumor cells from bone marrow. *Leukemia Res.* 17 7:557–560.

⁹Available from Director, Cosmetic Ingredient Review, 1101 17th Street NW, Suite 310, Washington, DC 20036, USA.

- Colón, A., D. Caro, C. J. Bourdony, and O. Rosario. 2000. Identification of phthalate esters in the serum of young Puerto Rican girls with premature breast development. *Environ. Health Perspect.* 108 9:895–900.
- Consumer Product testing. 1978a. Oral LD50 (rat) study no. 78-032. Unpublished data submitted by CTFA. 13 pages.⁹
- Consumer Product testing. 1978b. Dermal LD50 (rat) study no. 78-033. Unpublished data submitted by CTFA. 13 pages.⁹
- Consumer Product testing. 1978c. Guinea pig sensitization study no. 78-034. Unpublished data submitted by CTFA. 13 pages.⁹
- Cosmetic, Toiletry, and Fragrance Administration (CTFA). 2001a. Unpublished data regarding current concentrations of cosmetic uses of Dibutyl Phthalate. January 4, 2001. 2 pages.⁹
- CTFA. 2001b. Unpublished data regarding current concentrations of cosmetic uses of Dibutyl Phthalate. April 20, 2001. 2 pages.⁹
- CTFA. 2001c. Unpublished data regarding current concentrations of cosmetic uses of Dimethyl and Diethyl Phthalate. October 18, 2001. 2 pages.⁹
- CTFA. 2001d. Unpublished data available on the CTFA International Cosmetic Legal & Regulatory Database Global Ingredient Search.⁹
- CTFA. 2002a. Unpublished data regarding physical properties of dibutyl phthalate. Submitted August 30, 2002. 3 pages.⁹
- CTFA. 2002b. Unpublished data regarding average hairspray and perfume use. June 26, 2002. 3 pages.⁹
- Cummings, A. M., and L. E. Gray. 1987. Dibutyl Phthalate: Maternal effects versus fetotoxicity. *Toxicol. Lett.* 39:43–50.
- David, R., M. Moore, M. Cifone, D. Finney, and D. Guest. 1999. Chronic peroxisome proliferation and hepatomegaly associated with the hepatocellular tumorigenesis of di(2-ethylhexyl)phthalate and the effects of recovery. *Toxicol. Sci.* 50:195–205.
- David, R. M. 2000. Exposure to phthalate esters. *Environ. Health Perspect.* 108:A440.
- Dirven, H., P. H. H. Vandenbroek, A. M. M. Arends, H. H. Nordkamp, A. Delepper, P. T. Henderson, and F. J. Jongeneelen. 1993. Metabolites of the plasticizer di(2-ethylhexyl)phthalate in urine samples of workers in polyvinylchloride processing industries. *Int. Arch. Occup. Environ. Health* 64:549–554.
- Duty, S. M., N. P. Singh, M. J. Silva, D. B. Barr, J. W. Brock, L. Ryan, R. F. Herrick, D. C. Christiani, and R. Hunter. 2002. The relationship between environmental exposures to phthalates and DNA damage in human sperm using the neutral comet assay. *Environ. Health Perspect.* available at <http://dx.doi.org/doi:10.1289/ehp.5756>.
- E. I. du Pont de Nemours and Company. 1946. Possibility of dermatitis from dimethyl phthalate when used as an insect repellent. Unpublished data collected Haskell Laboratory of Industrial Toxicology and submitted by Phthalate Esters Panel, American Chemical Council. 2 pages.⁹
- E. I. du Pont de Nemours and Company. 1970. Skin primary irritation and sensitization test on guinea pigs. Unpublished data collected Haskell Laboratory for Toxicology and Industrial Medicine and submitted by Phthalate Esters Panel, American Chemical Council. 2 pages.⁹
- E. I. du Pont de Nemours and Company. 1974. Dimethyl phthalate. Unpublished data collected Haskell Laboratory for Toxicology and Industrial Medicine and submitted by Phthalate Esters Panel, American Chemical Council. 2 pages.⁹
- Elder, R. L., ed. 1985. Final report on the safety assessment of Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate. *J. Am. Col. Toxicol.* 4:267–303.
- Elsisi, A. E., D. E. Carter, and I. G. Sipes. 1989. Dermal absorption of phthalate diester in rats. *Fundam. Appl. Toxic.* 12:70–77.
- Ema, M., H. Amano, T. Itami, and H. Kawasaki. 1993. Teratogenic evaluation of di-n-butyl phthalate in rats. *Toxicol. Lett.* 69:197–203.
- Ema, M., H. Amano, and Y. Ogawa. 1994. Characterization of the developmental toxicity of di-n-butyl phthalate in rats. *Toxicology* 86:163–174.
- Ema, M., A. Harazono, E. Miyawaki, and Y. Ogawa. 1997a. Developmental effects of di-n-butyl phthalate after a single administration in rats. *J. Appl. Toxicol.* 17:223–229.
- Ema, M., A. Harazono, E. Miyawaki, and Y. Ogawa. 1997b. Embryo lethality following maternal exposure to dibutyl phthalate during early pregnancy in rats. *Bull. Environ. Contam. Toxicol.* 58:636–643.
- Ema, M., R. Kurosaka, H. Amano, and Y. Ogawa. 1995a. Comparative developmental toxicity of n-butyl benzyl phthalate and di-n-butyl phthalate in rats. *Arch. Environ. Contam. Toxicol.* 28:223–228.
- Ema, M., R. Kurosaka, H. Amano, and Y. Ogawa. 1995b. Developmental toxicity evaluation of mono-n-butyl phthalate in rats. *Toxicology Lett.* 78:101–106.
- Ema, M., R. Kurosaka, A. Harazono, H. Amano, and Y. Ogawa. 1996. Phase specificity of developmental toxicity after oral administration of mono-n-butyl phthalate in rats. *Arch. Environ. Contam. Toxicol.* 31:170–176.
- Ema, M., and E. Miyawaki. 2001a. Adverse effects on development of the reproductive system in male offspring given monobutyl phthalate, a metabolite of dibutyl phthalate, during late pregnancy. *Reprod. Toxicol.* 15:189–194.
- Ema, M., and E. Miyawaki. 2001b. Effects of monobutyl phthalate on reproductive function in pregnant and pseudopregnant rats. *Reprod. Toxicol.* 15:261–267.
- Ema, M., E. Miyawaki, and K. Kawashima. 1998. Further evaluation of developmental toxicity of di-n-butyl phthalate following administration during late pregnancy in rats. *Toxicol. Lett.* 98:87–93.
- Ema, M., E. Miyawaki, and K. Kawashima. 2000a. Critical period for adverse effects on development of reproductive system in male offspring of rats given di-n-butyl phthalate during late pregnancy. *Toxicol. Lett.* 111:271–278.
- Ema, M., E. Miyawaki, and K. Kawashima. 2000b. Effects of dibutyl phthalate on reproductive function in pregnant and pseudopregnant rats. *Reprod. Toxicol.* 14:13–19.
- Environ Corporation. 1984. Summary of the Results of Surveys of the Amount and Frequency of Use of Cosmetic Products by Women. Unpublished report submitted by CTFA. 7 pages.⁹
- Environment Canada. 1994. Canadian environmental protection act priority substances list assessment report: dibutyl phthalate. NTIS Report No. PB94169059.
- Epstein, W. L. 1978. Human sensitization study of diethyl phthalate. Unpublished data submitted by CTFA. 4 pages.⁹
- Essex Testing Clinic, Inc. 1989. Clinical safety evaluation of diethyl phthalate in repeated insult patch test, with cover letter dated 05/09/94 (sanitized). NTIS Report No. OTS0557309.
- Essex Testing Clinic, Inc. 1990. Final report, clinical safety evaluation of test article diethyl phthalate control (neat); repeated insult patch test with cover letter dated 05/09/94 (sanitized). NTIS Report No. OTS0557311.
- European Commission. 2002. The rules governing cosmetic products in the European Union. Volume 1. Cosmetics legislation—Cosmetic products. Accessed April 22, 2002. <http://dg3.eudra.org/F3/home.html>.
- Field, E. A., C. J. Price, R. S. Sleet, J. D. George, M. C. Marr, C. B. Myers, B. A. Schwetz, and R. E. Morrissey. 1993. Developmental Toxicity evaluation of diethyl and dimethyl phthalate in rats. *Teratology*. 48:33–44.
- Food and Drug Administration (FDA). 2001. Frequency of use of cosmetic ingredients. *FDA database*. Washington, DC: FDA.
- Foster, P. M., M. W. Cook, L. V. Thomas, D. G. Walters, and S. D. Gangolli. 1983. Differences in urinary metabolic profile from di-n-butyl phthalate-treated rats and hamsters. A possible explanation for species differences in susceptibility to testicular atrophy. *Drug Metab. Dispos.* 11:59–61.
- Foster, P. M. D., R. C. Cattley, and E. Mylchreest. 2000. Effects of di-n-butyl phthalate (DBP) on male reproductive development in the rat: Implications for human risk assessment. *Food Chem. Toxicol.* 38:S97–S99.
- Fredricsson, B., L. Möller, Å. Pousette, and R. Westerholm. 1993. Human sperm motility is affected by plasticizers and diesel particle extracts. *Pharmacol. Toxicol.* 72:128–133.
- Fukuoka, M., T. Kobayashi, Y. Zhou, and T. Hayakawa. 1993. Mechanism of testicular atrophy introduced by di-n-butyl phthalate in rats. Part 4. Changes in the activity of succinate dehydrogenase and the levels of transferrin and ferritin in the Sertoli and germ cells. *J. Appl. Toxicol.* 13:241–246.
- Fukuoka, M., S. Nhmi, Y. Zhou, T. Kobayashi, and T. Hayakawa. 1996. Responsiveness of rat nonparenchymal liver cells and spleen cells to the decomposition of erythrocytes exposed to di-n-butyl phthalate and its metabolite. *Biol. Pharm. Bull.* 19:1154–1159.

- Fukuoka, M., T. Tanimoto, Y. Zhou, N. Kawasaki, and A. Tanaka. 1989. Mechanism of testicular atrophy by di-n-butyl phthalate in rats. Part 1. *J. Appl. Toxicol.* 9:277–283.
- Fukuoka, M., Y. Zhou, A. Tanaka, I. Ikemoto, and T. Machida. 1990. Mechanism of testicular atrophy induced by di-n-butyl phthalate in rats. Part 2. The effects on some testicular enzymes. *J. Appl. Toxicol.* 10:285–293.
- Gall, H., A. Köhler, and R. U. Peter. 1999. Anaphylactic shock reaction to dibutyl-phthalate-containing capsules. *Dermatology*. 199:169–170.
- Gray, T. J. B., I. R. Rowland, P. M. D. Foster, and S. D. Gangolli. 1982. Species differences in the testicular toxicity of phthalate esters. *Toxicology Lett.* 11:141–147.
- Gray, L. E., J. Ostby, J. Furr, M. Price, D. N. Veeramachaneni, and L. Parks. 2000. Perinatal exposure of the phthalates DEHP, BBP, and DINP, but not DEP, DMP, or DOTP, alters sexual differentiation of the male rat. *Toxicol. Sci.* 58:350–365.
- Gray, L. E., C. Wolf, C. Lambright, et al. 1999. Administration of potentially antiandrogenic pesticides (procymidone, linuron, iprodione, chlozolinate, *p,p'*-DDE, and ketoconazole) and toxic substances (dibutyl- and diethylhexyl phthalate, PCB 169, and ethane dimethane sulphonate) during sexual differentiation produces diverse profiles of reproductive malformations in the male rat. *Toxicol. Ind. Health* 15:94–118.
- Hall, M., A. Mathews, L. Webley, and R. Harling. 1999. Effects of diisononyl phthalate (DINP) on peroxisomal markers in the marmoset-DINP is not a peroxisome proliferator. *J. Toxicol. Sci.* 24:237–244.
- Hansen, E., and O. Meyer. 1989. No embryotoxic or teratogenic effect of dimethyl phthalate in rats after epicutaneous application. *Pharmacol. Toxicol.* 64:237–238.
- Harris, C. A., P. Henttu, M. G. Parker, and J. P. Sumpter. 1997. The estrogenic activity of phthalate esters *in vitro*. *Environ. Health Perspect.* 105:802–811.
- Harrison Research Laboratories. 1998. Repeated insult patch test (RIPT) on a nail enamel. Unpublished data submitted by CTFA. 13 pages.⁹
- Hill Top Research, Inc. 1968. Primary irritation patch test on 3239K. Unpublished data submitted by CTFA. 13 pages.⁹
- Hill Top Research, Inc. 1971. Repeated insult patch test of group number 69, A-40-DIETPH. Unpublished data submitted by CTFA. 11 pages.⁹
- Hill Top Research, Inc. 1998. Evaluation of phototoxicity in humans. HTR project no. 101649. Unpublished data submitted by CTFA. 56 pages.⁹
- Hill Top Research, Inc. 1999. Evaluation of phototoxicity in humans. HTR project no. 100688. Unpublished data submitted by CTFA. 62 pages.⁹
- Hill Top Research, Ltd. 1997a. Evaluation of human photoallergy by repeated insult patch test. Project no. 97-4182-74. Unpublished data submitted by CTFA. 58 pages.⁹
- Hill Top Research, Ltd. 1997b. Evaluation of human photoallergy by repeated insult patch test. Project no. 97-4193-74. Unpublished data submitted by CTFA. 84 pages.⁹
- Hoppin, J. A., J. W. Brock, B. J. Davis, and D. D. Baird. 2002. Reproducibility of urinary phthalate metabolites in first morning urine samples. *Environ. Health Perspect.* 110:515–518.
- Houlihan, J., C. Brody, and B. Schwan. 2002. Not too pretty. Phthalates, beauty products & the FDA. Unpublished report released by Environmental Working Group, Coming Clean, and Health Care Without Harm. Available at <http://www.nottoopretty.org>
- Imajima, T., T. Shono, O. Zakaria, and S. Suita. 1997. Prenatal phthalate causes cryptorchidism postnatally by inducing transabdominal ascent of the testes in fetal rats. *J. Pediatr. Surg.* 32:18–21.
- Inveresk Research International. 1990. 2 weeks dermal dose range finding study in rats. Unpublished data submitted by CTFA. 652 pages.⁹
- Ivy Research Laboratories. 1973. The primary irritation potential of one fragrance material in six different vehicle formulations. Unpublished data submitted by CTFA. 5 pages.⁹
- Ivy Research Laboratories. 1984. The determination of the contact-sensitizing potential of four materials by means of the maximization study. Unpublished data submitted by CTFA. 4 pages.⁹
- Jackson Research Associates. 2002. Analysis of the Subungual Penetration of 100% Dibutyl Phthalate (DBP) in Human Fingernails. Unpublished report submitted by the American Beauty Association, 10-25-02.⁹
- Japanese Plasticizer Industrial Association. 1997. Evaluation of estrogen activity of phthalate esters by receptor-binding test (Study no. 6L471). Unpublished data submitted by Phthalate Esters Panel, American Chemistry Council. 59 pages.⁹
- Japanese Plasticizer Industrial Association. 1998. Evaluation of estrogen activity and mechanism of action of DBP and DEHP by *in vivo* assay using rats (Study no.: 7L432). Unpublished data submitted by Phthalate Esters Panel, American Chemistry Council. 27 pages.⁹
- Jensen, P. A., and O'Brien, D. 1993. Industrial Hygiene. In *Aerosol measurement. principles techniques and applications*, ed. K. Willeke and P. A. Baron. New York: John Wiley and Sons.
- Jones, H. B., D. A. Garside, R. Liu, and J. C. Roberts. 1993. The influence of phthalate esters on Leydig cell structure and function *in vitro* and *in vivo*. *Exp. Molecul. Pathol.* 58:179–193.
- Jury, H. H., T. R. Zacharewski, and G. L. Hammond. 2000. Interaction between human plasma sex hormone-binding globulin and xenobiotic ligands. *J. Steroid Biochem. Mol. Biol.* 75:167–176.
- Kageyama, K., Y. Onoyama, T. Nakajima, S. Otani, I. Yano, H. Hotta, I. Matsui-Yuasa, H. Kogawa, and N. Miwa. 1994. DNA synthesis inhibition and transmembrane permeation into tumor cells by various dialkyl phthalates upon hyperthermia. *Anticancer Res.* 14:2769–2772.
- Kanerva, L., R. Jolanki, K. Alanko, and T. Esterlander. 1999. Patch-test reactions to plastic and glue allergens. *Acta. Derm. Venereol.* 79:296–300.
- Kanerva, L., R. Jolanki, and T. Esterlander. 1997. Allergic and irritant patch test reactions to plastic and glue allergens. *Contact Dermatitis.* 37:301–302.
- Karmin, M. A., and G. H. Mayor. 1991. Diethyl phthalate: A perspective. *J. Clin. Pharmacol.* 31:484–489.
- Kavlock, R., K. Boekelheide, R. Chapin, M. Cunningham, E. Faustman, P. Foster, M. Golub, R. Henderson, I. Hinberg, R. Little, J. Seed, K. Shea, S. Tabacova, R. Tyl, P. Williams, and T. Zacharewski. 2002. NTP Center for the Evaluation of Risks to Human Reproduction: Phthalates expert panel report on the reproductive and developmental toxicity of di-n-butyl phthalate. *Reprod. Toxicol.* 16:489–527.
- Khalik, M. A., M. S. Alam, and S. P. Srivastava. 1992. Implications of physicochemical factors on the migration of phthalate esters from tubing commonly used for oral/nasal feeding. *Bull Environ Toxicol.* 48:572–578.
- Kim, S. H., S. S. Kim, O. Kwon, K. H. Sohn, et al. 2002. Effects of dibutyl phthalate and monobutyl phthalate on cytotoxicity and differentiation in cultured rat embryonic limb bud cells; protection by antioxidants. *Toxicol. Environ. Health.* 65:461–472.
- Kleinsasser, N. H., E. R. Kastenbauer, H. Weissacher, R. K. Muezenrieder, and U. A. Harréus. 2000. Phthalates demonstrate genotoxicity on human mucosa of the upper aerodigestive tract. *Environ. Mol. Mutagen.* 35:9–12.
- Kluwe, W. M. 1982. Overview of phthalate ester pharmacokinetics in mammalian species. *Environ. Health Perspect.* 45:3–10.
- Kohn, M. C., F. Parham, S. A. Masten, C. J. Portier, M. D. Shelby, J. W. Brock, and L. L. Needham. 2000. Human exposure estimates for phthalates. *Environ. Health Perspect.* 108:A440–A442.
- Koo, J. W., F. Parham, M. C. Kohn, S. A. Masten, J. W. Brock, L. L. Needham, and C. J. Portier. 2002. The association between biomarker based exposure estimates for phthalates and demographic factors in a human reference population. *Environ. Health Perspect.* 110:405–410.
- Kozombo, W. J., R. Kroll, and R. J. Rubin. 1982. Assessment of the mutagenicity of phthalate esters. *Environ. Health Perspect.* 45:103–109.
- Kurata, Y., F. Kidachi, M. Yokoyama, N. Toyota, M. Tsuchitani, and M. Katoh. 1998. Subchronic toxicity of di(2-ethylhexyl) phthalate in common marmosets: Lack of peroxisomal proliferation, testicular atrophy, or pancreatic acinar cell hyperplasia. *Toxicol. Sci.* 42:49–56.
- Lamb, J. C., R. E. Chapin, J. Teague, A. D. Lawton, and J. R. Reel. 1987. Reproductive effects of four phthalic acid esters in the mouse. *Toxicol. Appl. Pharmacol.* 88:255–269.

- Lamb, J. C., J. Reel, and A. D. Lawton. 1997a. Diethylphthalate. *Environ. Health Perspect.* 105:245–246.
- Lamb, J. C., J. Reel, and A. D. Lawton. 1997b. Di-n-butylphthalate, mice. *Environ. Health Perspect.* 105:247–248.
- Leberco Laboratories. 1963. Rabbit eye irritation study of 12.5% diethyl phthalate. Unpublished data submitted by CTFA. 5 pages.⁹
- Lington, A., M. Bird, R. Plutnick, W. Stubblefield, and R. Scala. 1997. Chronic toxicity and carcinogenic evaluation of di-isononylphthalate in rats. *Fundam. Appl. Toxicol.* 36:79–89.
- Lisheng, W., W. Chutse, P. Xuetao, et al. 1996a. Dibutyl phthalate purged autologous bone marrow transplant in the treatment of leukemia. *Leukemia Res.* 20:343–348.
- Lisheng, W., W. Chutse, P. Xuetao, X. Li, C. Jurong, and H. Fuchu. 1996b. Purgative effect of dibutyl phthalate on leukemic cells associated with apoptosis. *Leukemia Res.* 20:989–992.
- Liss, G. M., P. W. Albro, R. W. Hartle, and W. T. Stringer. 1985. Urine phthalate determinations as an index of occupational exposure to phthalic anhydride and di(2-ethylhexyl)phthalate. *Scand. J. Work Environ. Health.* 11:381–387.
- Lovekamp, T. N., and B. J. Davis. 2001. Mono-(2-ethylhexyl) phthalate suppresses aromatase transcript levels and estradiol production in cultured rat granulosa cells. *Toxicol. Appl. Pharmacol.* 172:217–224.
- Lygre, H., E. Solheim, and N. R. Gjerdet. 1995. Leaching from denture base materials *in vitro*. *Acta Odontol. Scand.* 53:75–80.
- McKee, R. H. 2000. The role of inhibition of gap junction intercellular communication in rodent liver tumor induction by phthalates: Review of data on selected phthalates and their potential relevance to man. *Regula. Toxicol. Pharmacol.* 32:51–55.
- Melnick, R. L., and C. M. Schiller. 1985. Effect of phthalate esters on energy coupling and succinate oxidation in rat liver mitochondria. *Toxicology* 34:13–27.
- Milligan, S. R., A. V. Balasubramanian, and J. C. Kalita. 1998a. Relative potency of xenobiotic estrogens in an active *in vivo* mammalian assay. *Environ. Health Perspect.* 106:23–26.
- Milligan, S. R., O. Khan, and M. Nash. 1998b. Competitive binding of xenobiotic oestrogens to rat alpha-fetoprotein and to sex steroid binding proteins in human and rainbow trout (*Oncorhynchus mykiss*) plasma. *Gen. Comp. Endocrinol.* 112:89–95.
- Ministry of Health, Labor and Welfare. 2001a. Unofficial translation of MHW Ordinance No. 331, Attached Table 1 [Negative List]. Ministry of Health, Labor and Welfare, Pharmaceutical and Medical Safety Bureau, Inspection and Guidance Division, 2-2, 1-chome, Kasumigaseki, Chiyoda-ku, Tokyo 100-8045, Japan.
- Ministry of Health, Labor and Welfare. 2001b. Unofficial translation of MHW Ordinance No. 331, Attached Table 2 [Restricted List]. Ministry of Health, Labor and Welfare, Pharmaceutical and Medical Safety Bureau, Inspection and Guidance Division, 2-2, 1-chome, Kasumigaseki, Chiyoda-ku, Tokyo 100-8045, Japan.
- Mint, A., S. A. M. Hotchkiss, and J. Caldwell. 1994. Percutaneous absorption of diethyl phthalate through rat and human skin *in vitro*. *Toxicol. In Vitro.* 8:251–256.
- Monsanto Europe. No date. Study to evaluate the effect of monobutyl phthalate on uterine growth in immature female rats after oral administration. Unpublished data submitted by Phthalate Esters Panel, American Chemistry Council. 16 pages.⁹
- Moura, C., M. Dias, and T. Vale. 1994. Contact dermatitis in painters, polishers and varnishers. *Contact Dermatitis.* 31:51–52.
- Murakami, K., K. Nishiyama, and T. Higuti. 1986a. Mitochondria effect of orally administered dibutyl phthalate in rats. *Jpn. J. Hyg.* 41:769–774.
- Murakami, K., K. Nishiyama, and T. Higuti. 1986b. Toxicity of dibutyl phthalate and its metabolites in rats. *Jap. J. Hyg.* 41:775–781.
- Murature, D. A., S. Y. Tang, G. Steinhardt, and R. C. Dougherty. 1987. Phthalate esters and semen quality parameters. *Biomed. Environ. Mass Spectrom.* 14:473–477.
- Mylchreest, E., R. C. Cattley, and P. M. D. Foster. 1998. Male reproductive tract malformations in rats following gestational and lactational exposure to di(n-butyl) phthalate: An androgenic mechanism? *Toxicol. Sci.* 43:47–60.
- Mylchreest, E., M. Sar, R. C. Cattley, and P. M. D. Foster. 1999. Disruption of androgen-regulated male reproductive development by di(n-butyl) phthalate during late gestation in rats is different from flutamide. *Toxicol. Appl. Pharmacol.* 156:81–95.
- Mylchreest, E., M. Sar, D. G. Wallace, and P. M. D. Foster. 2002. Fetal testosterone insufficiency and abnormal proliferation of Leydig cells and gonocytes in rats exposed to di(n-butyl) phthalate. *Reprod. Toxicol.* 16:19–28.
- Mylchreest, E., D. G. Wallace, R. C. Cattley, and P. M. D. Foster. 2000. Dose-dependent alterations in androgen-regulated male reproductive development in rats exposed to di(n-butyl) phthalate during late gestation. *Toxicol. Sci.* 55:143–151.
- Nail Manufacturers Council (NMC). 2001. Letter from NMC including ingredient use product type and concentration survey. Unpublished data submitted by NMC of May 7, 2001. 4 pages.⁹
- National Toxicology Program (NTP). 1991. Final report: On the reproductive toxicity of di-n-butyl phthalate (CAS No. 84-72-2) in Sprague-Dawley rats. NTIS Report No. PB92111996.
- NTP. 1995a. NTP technical report on toxicity studies of dibutyl phthalate (CAS No. 84-74-2) administered in feed to F344/N rats and B6C3F1 mice. NTIS Report No. PB95232427.
- NTP. 1995b. Toxicology and carcinogenesis studies of diethylphthalate in F344/n rats and B6C3F1 mice with dermal initiation/promotion study of diethylphthalate and dimethylphthalate in male Swiss (CD-1) mice. Unpublished data submitted by Phthalate Esters Panel, American Chemistry Council. 278 pages.⁹
- Netherlands Organization for Applied Scientific Research and National Institute of Public Health and the Environment. 2001. Risk assessment, dibutylphthalate. Unpublished data submitted by CTFA. 183 pages.⁹
- Ohtani, H., I. Miura, and Y. Ichikawa. 2000. Effects of dibutyl phthalate as an environmental endocrine disruptor on gonadal sex differentiation of genetic males of the frog *Rana rugosa*. *Environ. Health Perspect.* 108:1189–1193.
- Ow, R. K. K., and S. Amrith. 1997. Ocular prosthetics: Use of a tissue conditioner material to modify a stock ocular prosthesis. *J. Prosthet. Dentistry.* 78:218–222.
- Paganet, G., F. Campi, K. Varani, A. Piffanelli, G. Giovannini, and P. A. Borea. 2000. Endocrine-disrupting agents on healthy human tissues. *Pharmacol. Toxicol.* 86:24–29.
- Page, B. D., and G. M. Lacroix. 1992. Studies into the transfer and migration of phthalate esters from aluminum foil-paper laminates to butter and margarine. *Food Add. Contami.* 9:197–212.
- Payan, J. P., J. P. Marty, J. P. Fabry, D. Beydon, I. Boudry, E. Ferrari, F. Canel, M. C. Grandclaude, and C. M. Vincent. 2001. *In vivo* and *in vitro* percutaneous absorption of [¹⁴C]di-n-butylphthalate in rat. *Drug Metab. Distrib.* 29:843–854.
- Peakall, D. B. 1975. Phthalate esters: occurrence and biological effects. *Residue Rev.* 54:1–41.
- Pfordt, J., and E. Bruns-Weller. 1999. Phthalate esters as a group of environmental chemicals with an endocrine disruption potential. Lower Saxony Ministry of Food, Agriculture and Forestry: Hamburg. Unpublished data submitted by the Phthalate Esters Panel, American Chemistry Council. 53 pages.⁹
- Phthalate Esters Panel, American Chemistry Council. 1982a. Toxicity tests of dibutyl phthalate, dibutyl isophthalate, and dibutyl terephthalate. Unpublished data submitted by the Phthalate Esters Panel, American Chemistry Council. 5 pages.⁹
- Phthalate Esters Panel, American Chemistry Council. 1982b. 28-day subacute percutaneous toxicity SL: VE-1005-20, SR: TGSE 1037, P75-006. Unpublished data submitted by the Phthalate Esters Panel, American Chemistry Council. 10 pages.⁹
- Phthalate Esters Panel, American Chemistry Council. 1999. Di-n-butylphthalate (DBP). Unpublished data submitted by the Phthalate Esters Panel, American Chemistry Council. 42 pages.⁹

- Plasterer, M. R., W. S. Bradshaw, G. M. Booth, M. W. Carter, R. L. Schuler, and B. D. Hardin. 1985. Developmental toxicity of nine selected compounds following prenatal exposure in the mouse: Napthalene, p-nitrophenol, sodium selenate, dimethyl phthalate, ethylenethiourea, and four glycol ether derivatives. *J. Toxicol. Environ. Health* 15:25–38.
- Proctor & Gamble Co. 1984. Teratogenicity study of E-2426.01 (diethyl phthalate) by dermal application to rabbits with cover letter dated 05/02/94. NTIS Report No. OTS0572465.
- Pugh, G., J. Isenberg, L. Kamendulis, L. Clare, W. Brown, A. Lington, D. Ackley, J. Smith, and J. Klaunig. 2000. Effects of di-isononyl phthalate, di-2-ethylhexyl phthalate, and clofibrate in cynomolgus monkeys. *Toxicol. Sci.* 56:181–188.
- Research Triangle Institute. 1995. The toxicokinetics and metabolism of di-n-butyl phthalate. Unpublished data submitted by CTFA. 58 pages.⁹
- Research Institute for Fragrance Materials (RIFM). 1984. Acute dermal irritation study. RIFM 1795. Unpublished data submitted by CTFA. 13 pages.⁹
- RIFM. 1984. Acute dermal irritation study. RIFM 3099. Unpublished data submitted by CTFA. 13 pages.⁹
- Saillenfait, A. M., J. P. Payan, J. P. Fabry, et al. 1998. Assessment of the developmental toxicity, metabolism, and placental transfer of di-n-butyl phthalate administered to pregnant rats. *Toxicol. Sci.* 45:212–224.
- Saillenfait, A. M., I. Langonné, and B. Leheup. 2001. Effects of mono-n-butyl phthalate on the development of rat embryos: *in vivo* and *in vitro* observations. *Pharmacol. Toxicol.* 89:104–112.
- Schreck, C. E. 1991. Permethrin and dimethyl phthalate as tent fabric treatments against *Aedes aegypti*. *J. Am. Mosquito Control Assoc.* 7:533–535.
- Sciences International, Inc. 1995. Toxicological profile for diethyl phthalate. Unpublished data submitted by CTFA on August 30, 2002. 3 pages.⁹
- Scientific Committee on Cosmetic Products and Non-Food Products (SCCNFP). 2002. Opinion of the scientific committee on cosmetic products and non-food products intended for consumers concerning diethyl phthalate. Unpublished data submitted by CTFA. 36 pages.⁹
- Scott, R. C., P. H. Dugard, J. D. Ramsey, and C. Rhodes. 1987. *In vitro* absorption of some o-phthalate diesters through human and rat skin. *Environ. Health Perspect.* 74:223–227.
- Selene, C. H., J. Chou, J. Holler, and C. De Rosa. 1998. Minimal risk levels (MRLs) for hazardous substances. *J. Clean Technol. Environ. Toxicol. Occup. Med.* 7:1–24.
- Shultz, V. D., S. Phillips, M. Sar, P. M. D. Foster, and K. W. Gaido. 2001. Altered gene profiles in fetal rats after in utero exposure to di(n-butyl) phthalate. *Toxicol. Sci.* 64:233–242.
- Small, L. D., E. O. Dillingham, and J. Autian. 1973. Tissue distribution and excretion of diethyl phthalate following percutaneous administration to female albino rats. Unpublished data submitted by CTFA. 23 pages.⁹
- Sonde, V., A. D'souza, R. Tarapore, L. Pereira, M. P. Khare, P. Sinkar, S. Krishnan, and C.V. Rao. 2000. Simultaneous administration of diethylphthalate and ethyl alcohol and its toxicity in male Sprague-Dawley rats. *Toxicology*. 147:23–31.
- Srivastava, S., G. B. Singh, S. P. Srivastava, and P. K. Seth. 1990a. Testicular toxicity of di-n-butyl phthalate in adult rats: Effect on marker enzymes of spermatogenesis. *Ind. J. Exp. Biol.* 28:67–70.
- Srivastava, S. P., S. Srivastava, D. K. Saxena, S. V. Chandra, and P. K. Seth. 1990b. Testicular effects of di-n-butyl phthalate (DBP): Biochemical and histopathological alterations. *Arch. Toxicol.* 64:148–152.
- Tarvainen, K. 1995. Analysis of patients with allergic patch test reactions to a plastics and glues series. *Contact Dermatitis* 32:346–351.
- Tavares, I. A., and N. D. Vine. 1985. Phthalic acid esters inhibit arachidonate metabolism by rat peritoneal leucocytes. *J. Pharm. Pharmacol.* 37:67–68.
- Tokunaga, H., Y. Ching, T. Uchino, and M. Ando. 2001. Studies of diethyl phthalate on *in vitro* percutaneous permeation. *J. Soc. Cosmet. Chem. Jpn.* 35:312–316.
- Unilever. 1999. Diethyl phthalate: assessment for oestrogenic potentia I. Unpublished data submitted by CTFA. 126 pages.⁹
- US Environmental Protection Agency. 1994a. Final report, clinical safety evaluation of test article diethyl phthalate control (neat); repeated insult patch test, with cover letter dated 05/09/94 (sanitized). NTIS No. OTS0557311.
- US Environmental Protection Agency. 1994b. Final report, clinical safety evaluation of test article diethyl phthalate in repeated insult patch test, with cover letter dated 05/09/94 (sanitized). NTIS No. OTS0557309.
- Van Beek, L. 1974. Primary skin irritation and eye irritation tests with a perfume concentrate. Unpublished data submitted by CTFA. 9 pages.⁹
- Waliszewski, St. M., and G. A. Szymczyński. 1989. Determination of phthalate esters in human semen. *Andrologia* 229:69–73.
- Walseth, F. and O. G. Nilsen. 1984. Phthalate esters: II. Effects of inhaled dibutyl phthalate on cytochrome P-450 mediated metabolism in rat liver and lung. *Arch. Toxicol.* 55:132–136.
- Walseth, F., and O. G. Nilsen. 1986. Phthalate esters: Effects of orally administered dibutylphthalate on cytochrome P-450 mediated metabolism in rat liver and lung. *Acta Pharmacol. Toxicol.* 59:263–269.
- White, R. D., D. L. Earnest, and D. E. Carter. 1983. The effect of intestinal esterase inhibition on the *in vivo* absorption and toxicity of di-n-butyl phthalate. *Food Chem. Toxicol.* 21:99–101.
- Wilkinson, S. M., and M. H. Bech. 1992. Allergic contact dermatitis from dibutyl phthalate, propylgallate and hydrocortisone in Timodine®. *Contact Dermatitis* 27:197.
- Williams, D. T., and B. J. Blanchfield. 1975. The retention, distribution, excretion, and metabolism of dibutyl phthalate-7-¹⁴C in the rat. *J. Agric. Food Chem.* 23:854–858.
- Wine, R. N., L. H. Li, L. H. Barnes, D. K. Gulati, and R. E. Chapin. 1997. Reproductive toxicity of di-n-butylphthalate in a continuous breeding protocol in Sprague-Dawley rats. *Environ. Health Perspect.* 105:102–107.
- Zacharewski, T. R., M. D. Meek, J. H. Clemons, Z. F. Wu., M. R. Fieldon, and J. B. Mathews. 1998. Examination of the *in vitro* and *in vivo* estrogenic activities of eight commercial phthalate esters. *Toxicol. Sci.* 46:282–293.
- Zhou, Y., M. Fukuoaka, and A. Tanaka. 1990. Mechanism of testicular atrophy induced by di-n-butyl phthalate in rats. Part 3. Changes in the activity of some enzymes in the Sertoli and germ cells, and in the levels of metal ions. *J. Appl. Toxicol.* 10:447–453.

DIMETHICONE COPOLYOL

A safety assessment of Dimethicone Copolyol was published in 1982 with the conclusion that this ingredient “is safe as a cosmetic ingredient in the present practices of use and concentration” (Elder 1982). New studies, along with updated information regarding types and concentrations of use, were considered by the CIR Expert Panel. The Panel determined to not reopen this safety assessment.

Dimethicone Copolyol as an ingredient name has been deleted from the *International Cosmetic Ingredient Dictionary and Handbook* and replaced with several ingredients that adhere to the former definition of Dimethicone Copolyol as “a polymer of dimethylsiloxane with polyoxyethylene and/or polyoxypropylene side chains” (Pepe et al. 2002).

The new ingredients that are described by this definition include: Dimethicone PEG-7 Phosphate, Dimethicone PEG-10 Phosphate, Dimethicone PEG/PPG-7/4 Phosphate, Dimethicone PEG/PPG-12/4 Phosphate, Dimethicone PEG/PPG-20/23 Benzoate, Dimethicone PEG-8 Benzoate, Dimethicone PEG-6 Acetate, Dimethicone PEG-8 Adipate, PEG-3 Dimethicone, PEG-9 Dimethicone, PEG/PPG-20/29 Dimethicone,

Dibutyl, Dimethyl, and Diethyl Phthalate and Butyl Benzyl Phthalate

International Journal of Toxicology
2017, Vol. 36(Supplement 2) 44S-45S
© The Author(s) 2017
Reprints and permission:
sagepub.com/journalsPermissions.nav
DOI: 10.1177/1091581817716148
journals.sagepub.com/home/ijt



F. Alan Andersen¹

Keywords

phthalates, safety, cosmetics

Conclusion

In a 1985 safety assessment of dibutyl, diethyl, and dimethyl phthalate, the Cosmetic Ingredient Review Expert Panel stated that these ingredients are safe for use in cosmetics in the present practice of use and concentration.¹ Subsequently, in 2005, the Panel conducted an extensive rereview of the newly available studies since that assessment, confirmed the decision, and determined to not reopen that report.² In 1992, butyl benzyl phthalate was found safe in the present practice of use and concentration.³ The Panel reviewed studies performed since that assessment as well as updated the use and concentration data in 2007 and confirmed that conclusion.⁴ In 2012, the Panel reviewed 3 new studies on phthalates published in 2012 and confirmed that dibutyl, dimethyl, and diethyl phthalate and butyl benzyl phthalate are safe in cosmetics in the present practices of use and concentration. The Panel did not reopen the safety assessment.

Discussion

The Panel reviewed new studies that focused on the potential for endocrine disruption/reproductive and developmental toxicity on dibutyl, dimethyl, and diethyl phthalate and butyl benzyl phthalate. One study of children aged 5 to 9, who were part of a Manhattan-Bronx cohort, revealed detectable, although varied, levels of phthalates in the urine of all 244 study participants.⁵ Higher levels of both diethyl phthalate and butyl benzyl phthalate were associated with airway inflammation.

Two studies addressed diabetes and phthalates. In 1 study, there were 1,015 men and women 70 years of age from Uppsala, Sweden.⁶ One sample per participant was collected from 2001 to 2004 and analyzed 5 to 8 years later. In this study, blood levels for dimethyl phthalate, diethyl phthalate, diisobutyl phthalate, and diethylhexyl phthalate were measured and correlated with measures of insulin resistance and poor insulin secretion in nondiabetic participants.

In the second study, urinary concentrations of phthalate metabolites measured by the Centers for Disease Control and Prevention and self-reported diabetes in 2,350 women aged 20 to <80 participating in the National Health and Nutrition Examination Survey (NHANES) (2001-2008) were used.⁷ The odds

ratio for diabetes in women with higher levels of n-butyl phthalate, isobutyl phthalate, benzyl phthalate, 3-carboxypropyl phthalate, and the sum of diethylhexyl phthalate metabolites was greater than the odds ratio for women with the lowest concentrations of these phthalates.

The Panel noted that all of these studies identified associations between phthalate metabolites and either diabetes or airway inflammation. Such studies did not suggest a causal link between phthalates and any adverse outcome. The possibility that phthalate metabolites may impact peroxisome proliferation pathways was suggested in the diabetes studies, but that mechanism is not established as a mode of action. The Panel agreed that there is a need for further study of the reported association between phthalates exposures and diabetes and to investigate possible causal links.

Declaration of Conflicting Interests

The author(s) declared no potential conflicts of interest with respect to the research, authorship, and/or publication of this article.

Funding

The author(s) received no financial support for the research, authorship, and/or publication of this article.

References

1. Final Report on the Safety Assessment of Dibutyl Phthalate, Dimethyl Phthalate, and Diethyl Phthalate. *J Am Coll Toxicol.* 1985;4(3):267-303. <http://journals.sagepub.com/doi/abs/10.3109/10915818509078674>.
2. Annual Review of Cosmetic Ingredient Safety Assessments—2002/2003. *Int J Toxicol.* 2005;24 (suppl 1):1-102. PM:15981733.
3. Final Report on the Safety Assessment of Butyl Benzyl Phthalate. *J Am Coll Toxicol.* 1992;11(1):1-23. <http://journals.sagepub.com/doi/abs/10.3109/10915819209141988>.

¹ Former Director, Cosmetic Ingredient Review, Washington, DC, USA

Corresponding Author:

Bart Heldreth, Executive Director, Cosmetic Ingredient Review, 1620 L Street NW, Suite 1200, Washington, DC 20036-4702, USA.

Email: cirinfo@cir-safety.org

4. Andersen FA. Annual Review of Cosmetic Ingredient Safety Assessments: 2007-2010. *Int J Toxicol*. 2011;30(5 suppl): 73S-127S. PM:22068630.
5. Just AC, Whyatt RM, Miller RL, et al. Children's urinary phthalate metabolites and fractional exhaled nitric oxide in an urban cohort. *Am J Respir Crit Care Med*. 2012;186(9):830-837. PM:22923660.
6. Lind PM, Zethelius B, Lind L. Circulating levels of phthalate metabolites are associated with prevalent diabetes in the elderly. *Diabetes Care*. 2012;35(7):1519-1524. PM:22498808.
7. James-Todd T, Stahlhut R, Meeker JD, et al. Urinary phthalate metabolite concentrations and diabetes among women in the National Health and Nutrition Examination Survey (NHANES) 2001-2008. *Environ Health Perspect*. 2012;120(9):1307-1313. PM:22796563.

Concentration of Use by FDA Product Category^{1*}

Dibutyl Phthalate
Diethyl Phthalate
Dimethyl Phthalate

| Ingredient | Product Category | Maximum Concentration of Use |
|-------------------|---|------------------------------|
| Diethyl Phthalate | Skin cleansing (cold creams, cleansing lotions, liquids and pads) | 0.1% |
| Diethyl Phthalate | Face and neck products (not spray) Leave-on | 0.15% |

*The ingredients included in the title of the table but not found in the table were included in the concentration of use survey, but no uses were reported. In response to this survey, it was indicated that Dibutyl Phthalate and Diethyl Phthalate may be present in cosmetics as impurities.

Information collected in 2025
Table prepared: March 27, 2025

¹ The new FDA cosmetic product categories under MoCRA were used for this survey.